

Artificial light at night alters the sexual behaviour and fertilisation success of the common toad

Morgane Touzot, Thierry Lengagne, Jean Secondi, Emmanuel Desouhant,

Marc Thery, Adeline Dumet, Claude H. B. Duchamp, Nathalie Mondy

▶ To cite this version:

Morgane Touzot, Thierry Lengagne, Jean Secondi, Emmanuel Desouhant, Marc Thery, et al.. Artificial light at night alters the sexual behaviour and fertilisation success of the common toad. Environmental Pollution, 2020, 259, pp.113883. 10.1016/j.envpol.2019.113883. hal-02551696

HAL Id: hal-02551696 https://cnrs.hal.science/hal-02551696v1

Submitted on 5 Jan 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés. 1 Artificial light at night alters the sexual behaviour and fertilisation success of the common

2 toad

4	Morgane Touzot ^{1,*} , Thierry Lengagne ¹ , Jean Secondi ^{1,2} , Emmanuel Desouhant ³ , Marc
5	Théry ⁴ , Adeline Dumet ¹ , Claude Duchamp ¹ and Nathalie Mondy ¹
6	¹ Univ Lyon, Université Claude Bernard Lyon 1, CNRS, ENTPE, UMR 5023 LEHNA,
7	Villeurbanne, F-69622, France.
8	² Faculté des Sciences, Université d'Angers, 49045 Angers
9	³ Univ Lyon, Université Claude Bernard Lyon 1, CNRS, UMR5558 LBBE, Villeurbanne, F-
10	69622, France.
11	⁴ Centre National de la Recherche Scientifique, Muséum National d'Histoire Naturelle
12	(MNHN), UMR 7179, Brunoy, F-91800, France.
13	
14	*Corresponding author: Univ Lyon, Université Claude Bernard Lyon 1, CNRS, ENTPE,
15	UMR5023 LEHNA, F-69622, Villeurbanne, France. Fax: (+33) 472 431 141, Tel.: (+33)
16	472 431 520. Email: morgane.touzot@univ-lyon1.fr
17	
18	
19	
20	

21 Abstract

22 Artificial Light At Night (ALAN) is an emerging pollution, that dramatically keeps on increasing worldwide due to urbanisation and transport infrastructure development. In 2016, it 23 nearly affected 23% of the Earth's surface. To date, all terrestrial and aquatic ecosystems have 24 been affected. The disruption of natural light cycles due to ALAN is particularly expected for 25 nocturnal species, which require dark periods to forage, move, and reproduce. Despite the 26 growing number of studies on this subject, our knowledge on the direct influence of nocturnal 27 28 lighting on amphibians, the most endangered vertebrate, is still scarce. Apart from chiropterans, amphibians contain the largest proportion of nocturnal species among 29 vertebrates exhibiting an unfavourable conservation status in most parts of the world and 30 living in ALAN polluted areas. To better understand the consequences of ALAN on the 31 breeding component of amphibian fitness, we experimentally exposed male breeding common 32 toads (Bufo bufo) to ecologically relevant light intensities of 0.01 (control), 0.1 or 5 lux for 12 33 34 days. At mating, exposed males took longer than controls to form an amplexus, *i.e.*, to pair with a female, and broke amplexus before egg laying, while controls never did. These 35 behavioural changes were associated with fitness alteration. The fertilization rate of 5 lux-36 37 exposed males was reduced by 25%. Salivary testosterone, which is usually correlated with reproductive behaviours, was not altered by ALAN. Our study demonstrates that ALAN can 38 affect the breeding behaviour of anuran species and reduce one component of their fitness. 39 40 Given the growing importance of ALAN, more work is needed to understand its long-term consequences on the behaviour and physiology of individuals. It appears essential to identify 41 42 deleterious effects for animal populations and propose appropriate management solutions in an increasingly brighter world. 43

Capsule summarizing the main finding: Artificial light at night decreased the capacity of
male *Bufo bufo* to form a pair with a female and reduced their fertilization success.

Key words: Fitness, reproduction, amphibian, anthropogenic light pollution, behaviour

47 Introduction

In recent decades, the rapid growth in the world's population has led to a sharp increase in 48 human activities necessary to support this growth. One consequence of the expansion of urban 49 areas along with the development of transport infrastructures (Grimm et al. 2008; Gaston et 50 al. 2013) is the major increase of light levels at night. In 2016, nearly 23% of the Earth's 51 surface, 88% of Europe and almost half of the United States experienced brightness levels 52 higher than light levels at night in natural ecosystems (Falchi et al. 2016). Artificial Light At 53 54 Night (ALAN) dramatically expanded. From 2012 to 2016, Earth's artificially lit outdoor area grew by 2.2% per year, with a total radiance increase of 1.8% per year (Kyba et al. 2017). 55 One of the major effects of ALAN is the disruption of the natural photoperiod, which is one 56 of the most important cues for biological timing (Bradshaw & Holzapfel 2010). Among all 57 organisms, nocturnal species, which represent a large proportion of biodiversity, 28% of 58 vertebrates and more than 60% of invertebrates (Hölker et al. 2010), are most likely to 59 experience and to be affected by ALAN (Buchanan 2006; Duffy et al. 2015; Desouhant et al. 60 2019). Nocturnal artificial light is known to affect a wide range of physiological and 61 behavioural phenomena, such as migration (Van Doren et al. 2017), orientation (Tuxbury & 62 Salmon 2005), activity (e.g. Le Tallec et al. 2013; Pulgar et al. 2019; Touzot et al. 2019), 63 foraging (Czarnecka et al. 2019), energy balance (e.g. Welbers et al. 2017; Touzot et al. 2019) 64 and hormonal synthesis (e.g. Brüning et al. 2015; Newman et al. 2015). These effects on 65 individuals have the potential to alter population dynamics (e.g. Gaston et al. 2014a; Grubisic 66 et al. 2017; Sanders et al. 2018). 67

Despite the evidence of profound effects of ALAN on life history traits across
different taxonomic groups with important ecological consequences (*e.g.* Gaston et al. 2014b;
Knop et al. 2017; Bennie et al. 2018), we still have little information on the direct effects of
ALAN on the fitness of individuals although its assessments are key elements in conservation.

Fitness, *i.e.* the average contribution to the gene pool of the next generation that is made by 72 73 individuals of a specified genotype or phenotype in a given environment, is often estimated through measurements of mortality and breeding success. Most of the studies conducted to 74 75 date focused on the effects of ALAN on the mortality component of fitness (e.g. Rodríguez et al. 2014; Willmott et al. 2018). The breeding component of fitness has been less examined, 76 even if However, observation-based studies did not control for other environmental factors 77 78 that are often correlated with ALAN in the wild (temperature of urban environment, habitat fragmentation, prey availability...), while most experiment-based studies used higher 79 intensities than those actually encountered by individuals in their habitat (May et al. 2019). a 80 81 growing number of studies (reviewed in Ouyang et al. 2018) reported that the timing and the physiology of seasonal reproductive processes differ between individuals living in lit areas 82 and their conspecifics living in darker areas. Alterations of the period of functional 83 84 development of reproductive organs, reproductive hormonal synthesis (estradiol, ketotestosterone), number of eggs produced by females, egg hatchling success, number of 85 offspring and birth schedule in response to ALAN exposure were found in mammals (e.g. Le 86 Tallec et al. 2015; Robert et al. 2015), birds (e.g. Dominoni et al. 2013; de Jong et al. 2015), 87 88 fishes (e.g. Brüning et al. 2018, Fobert et al. 2019) and insects (e.g. Willmott et al. 2018; 89 McLay et al. 2018). Photoperiod modifications have also been shown to affect sexual maturation and sex steroids synthesis, especially in fish (*e.g.* Norberg et al. 2004). The 90 suppression of the dark period triggered a decrease of spermatocytes in male Asian toads, 91 92 Bufo melanostictus, (Biswas et al. 1978) and a severe reduction of sexual calls in male green frogs, Rana clamitans melanota, (Baker & Richardson 2006). On contrary, other studies have 93 94 found no effect of ALAN on the time for copulation or between mating and laying egg, on the number of eggs or egg sacs laid and on sperm viability (e.g. Durrant et al. 2018; McLay et al. 95 2018; Willmott et al. 2018). A recent study investigating the influence of ALAN on the 96

97 fertilization success of fish, ultimately showed no effect (Fobert et al. 2019). In view of the
98 consequences of ALAN on the breeding behaviour and physiology, it is necessary to
99 investigate the direct influence of ALAN on the breeding component of fitness in amphibians.

In this context, we experimentally studied the effect of three ecologically relevant light 100 intensities at night (0.01, 0.1 or 5 lux), which correspond to light levels measured in areas 101 102 hosting amphibians (Secondi et al. 2017), on both the reproductive behaviour and the fertilisation success of male common toads, Bufo bufo. This nocturnal amphibian is one of the 103 most common and ubiquitous amphibians in France, and as a result is often a useful indicator 104 of ecosystem health and function (Hilty & Merenlender 2000). Moreover, the common toad is 105 an explosive breeding species with a breeding period lasting only a few days, thus limiting the 106 107 number of pairing opportunities (Wells 1977). The operational sex ratio of this species is biased towards males, leading to scrambling competition among males to mate with a female. 108 During the breeding season, common toads are frequently found in urban and peri-urban areas 109 110 with wetlands (Beebee 1979), especially small ones (ponds for instances) which are subjected to ALAN (Secondi et al. 2017). In addition, due to their high nocturnal visual sensitivity, 111 amphibian activities, such as foraging and breeding, are expected to be affected by changes in 112 113 night brightness (Buchanan 2006; Grant et al. 2009; Yovanovich et al. 2017). We have previously shown that ALAN exposure decreased common toad activity during the night and 114 increased allocation of energy to maintenance (Touzot et al. 2019). This suggested that 115 116 breeding activities occurring at dusk or at night may be influenced by ALAN. Considering 117 this, we predicted that (i) ALAN exposure would alter male breeding behaviour, particularly 118 their ability to maintain pairing with a female, which is a costly activity for this species (Lengagne et al. 2007), (ii) these behavioural alterations could be due to changes in male 119 testosterone concentrations, and (iii) ALAN exposure would reduce male fertilisation success. 120

121 Methods

122 Animal collection and housing conditions

123 A total of 60 male common toads were collected during the breeding season (8-9 March 2018) in La Burbanche, France (45°N, 5°E). This site was chosen for its low levels of ALAN 124 regardless of weather conditions and the lunar phase (≤ 0.01 lux). Upon arrival at the animal 125 care facility (EcoAquatron, University of Lyon), males were weighed (LAB 800-3000, 126 127 precision: 0.1 g, B3C pesage, Sérénité) and housed individually in boxes (47 x 36 x 25 cm) containing a 15 cm section of PVC tubing (diameter 10 cm) for shelter and 4 cm of litter that 128 129 was moistened daily. Ambient temperature and relative humidity were kept during the whole experiment at 15.9 ± 0.6 °C and 55.2 ± 2.9 %, respectively. In these conditions, males were 130 active during the night. Toads were fed ad libitum with live domestic crickets, Acheta 131 domesticus, except during the mating trials. Sixty females of common toad collected at the 132 same breeding site were grouped by 5 in large boxes (60 x 42 x 39 cm) with 10 cm of litter 133 that was moistened daily. The ambient temperature was kept at $5.0 \pm 1.0^{\circ}$ C. At this 134 135 temperature, females were inactive which prevented early spawning. Females were not exposed to ALAN. Two days before the end of male ALAN exposure, the temperature of the 136 compartment containing females was increased to reach 15°C, a temperature which turns 137 138 them active during the night.

139 Light treatments

During 7-8 days of acclimation, toads were exposed during the daytime to 254 ± 4 lux (mean \pm SEM) provided by light tubes (Philips Master TL-D 58W/865 and Exo Terra Repti Glo 2.0, 40W T8) and remained in the dark (< 0.01 lux) during the scotophase. After the acclimation period, under the natural photoperiod at the date of the experiment, male toads were assigned to one of the three treatments (n=20 for each treatment) and exposed to their respective light treatment at night. To reproduce artificial light by night, we used white light-emitting diode

(LED) ribbons (white cold Light Plus, 6000-6500 K, 14W, 60 LED/m). White LEDs were 146 147 chosen because they are increasingly used for street lighting worldwide (Falchi et al. 2016). A LED ribbon of 95 cm (57 LED) was fixed 20 cm above the bottom of each box for each light 148 149 treatment (see Appendix Supplementary Material S1 for details). All the ribbons in a given light treatment were connected to a dimmer (manual dimmer, 12V max, 8A) and a laboratory 150 151 power supply (15V/DC max, 3A), which allowed to finely adjust the light intensity of each 152 light treatment. The boxes assigned to one light treatment were isolated from each other, so that the light intensities did not interfere with each other. Under the natural photoperiod at the 153 date of the experiment, during the night, the control group was exposed to 0.01 lux, 154 155 corresponding to the illuminance of a sky under clear conditions with a quarter moon (Gaston et al. 2013). The first experimental group was exposed to 0.10 lux (hereafter called the "0.1 156 lux-group"), corresponding to the illuminance of urban skyglow (Gaston et al. 2013). The 157 158 second experimental group was exposed to 5 lux (hereafter called the "5 lux-group"), which corresponds to the light level of a residential street (Gaston et al. 2013). The daylight and 5 159 160 lux-group intensities were measured with a luxmeter (Illuminance meter T-10A, Konica Minolta). Intensities for the 0.10 lux and control groups were measured with a highly 161 sensitive light meter (Sky Quality Meter-SQM-L, Unihedron). Consequently, to compare 162 163 intensities, SQM measured values were converted into lux (see Appendix Supplementary Material S2 for details). Light intensities were measured at the bottom of the boxes and 164 checked every week (0.01 \pm 0.001 lux for the control group, 0.05 \pm 0.01 lux for the 0.10 lux-165 166 group and 5.01 \pm 0.06 lux for the 5 lux-group, mean \pm SEM). The exposure period lasted twelve days (D1 to D12), which corresponds to the average duration of a breeding period for 167 this species in the wild (Wells 1977; Reading & Clarke 1983). 168

169 *Reproductive behaviours*

On D13, exposure to nocturnal light was stopped, and the litter contained in the box was 170 171 replaced with 211 of water (one-third of dechlorinated water and two-third of pond water) and tree branches for oviposition. Females that had not been exposed to ALAN were assigned to a 172 partner (we have checked that both female body mass and the ratio between male and female 173 size did not statistically differ between light treatments). During the following 27 hours, we 174 175 recorded hourly males that amplexed a female. Amplexus can be either successful (*i.e.*, the 176 male maintains the female until egg laying) or not (*i.e.*, the male stops maintaining the female in amplexus). As parameter we recorded the latency to obtain successful amplexus, *i.e.*, the 177 duration required for males to form an amplexus leading to egg laying, and the separation 178 179 rate, *i.e.*, the percentage of pairs that broke amplexus before egg laying. The following days 180 (between 1 and 7 days), we determined the duration of amplexus before the clutch was laid. After laying, animals were removed, weighed and measured (Snout-to-Vent length, SVL) 181 182 because previous studied showed that male pairing success was linked to their size (Davies & Halliday 1978; Lengagne et al. 2007) and fertilisation rate was linked to the ratio between 183 male and female size (Davies & Halliday 1977). Four days after laying, we quantified the 184 number of unfertilised eggs at the blastulation stage (Lengagne et al. 2007) and the total 185 186 number of laid eggs. These data allowed us to calculate the fertilisation rate, *i.e.*, the 187 proportion of fertilised eggs to the total number of laid eggs by a female.

188 Hormonal assay

On D13, males were weighed and measured, and a saliva sample was taken to measure their testosterone level. Sampling was carried out between 01:00 a.m. and 03:00 a.m., when testosterone concentration was the highest (see Appendix Supplementary Material S3 for details). Within the 2 min after capture, a dry cotton ball of known weight was introduced into the toad's mouth for 30 s to collect saliva. Each cotton ball was immediately weighed, placed into a microtube containing a filter to retain the cotton ball fibre during centrifugation and

stored at -80°C. We investigated testosterone levels in saliva, which is a non-invasive method 195 196 commonly used for the detection of steroids in wild species, including amphibians (Kutsukake et al. 2009; Janin et al. 2012; Desprat et al. 2015). Saliva was extracted from the cotton ball 197 by the addition of 500 µl of phosphate buffer (1 M phosphate solution containing 1% bovine 198 serum albumin, 4 M sodium chloride, 10 mM ethylenediaminetetraacetic acid, and 0.1% 199 200 sodium azide) and centrifugation (10 000 g at 4°C for 5 min). We used a testosterone Enzyme 201 Linked ImmunoSorbent Assay (ELISA) kit (number 582701, Cayman Chemical Company, Ann Arbor, MI, U.S.A.), validated for amphibians (Desprat et al. 2015) and previously 202 optimized for common toads, to analyzed testosterone from saliva samples. Concentrations 203 204 were calculated using a standard curve run on each plate and were estimated in 1 mg of saliva and expressed in pg mg⁻¹ of saliva. 205

206 Statistical Analysis

We performed a linear model (LM) to test the effects of the light treatment (0.01, 0.1, 5 lux) 207 on male body mass gain during ALAN exposure and on male body mass on D13. To confirm 208 209 that there were no confounding differences in body size between light treatments, we performed a LM to test the effect of light treatment (0.01, 0.1, 5 lux) on male SVL 210 211 measurement and on the ratio between male and female size on D13. Successful amplexus latency data were analysed as survival data following a Weibull distribution. Light treatment 212 213 (0.01, 0.1, 5 lux), male body mass, female body mass, male salivary testosterone 214 concentrations, and the interactions between light treatment and male body mass and between 215 light treatment and male salivary testosterone concentrations were included as explanatory 216 variables. The effect of light treatment (0.01, 0.1, 5 lux), male body mass, female body mass, 217 male salivary testosterone concentrations and the interactions between light treatment and 218 male body mass and between light treatment and male salivary testosterone concentrations on the probability of separation before egg laying (separation = 1; no separation = 0) was 219

analysed through a generalized linear model (GLM) with a binomial distribution (logit link). 220 221 As the results of the model indicated a marginal effect for the chosen significance threshold (5%), we calculated the effect size (Cohen's h for proportions). We performed a GLM with a 222 223 Gamma distribution to test the effect of the light treatment (0.01, 0.1, 5 lux), male body mass, female body mass and the interaction between light treatment and male body mass on 224 225 amplexus duration before egg laying. For this analysis, one value is missing in the 5 lux-226 group, as one female paired to a 5 lux-exposed male did not lay eggs. We performed a LM to test the effects of the light treatment (0.01, 0.1, 5 lux) and male body mass and their 227 interaction on male salivary testosterone concentrations. For male salivary testosterone 228 229 concentrations, one data point was excluded because the value was below the colour range of the ELISA kit. Male fertilisation rate was analysed through a GLM with a binomial 230 distribution (logit link) using male body mass, female body mass, light treatment and the 231 232 interaction between light treatment and male body mass as predictors. Because clutches are synchronous and, given the number of eggs to be counted, we counted fertilised eggs on a 233 234 sub-sample of pairs (n = 10 for control, n = 9 for 0.1 lux and n = 8 for 5 lux), which corresponds to 94 102 eggs counted. For the analyses, the fitted models were simplified by a 235 backward procedure to keep only the significant variables. For all models, multiple 236 237 comparisons were performed using a Tukey adjusted post hoc test. All data were analysed using the software R 3.5.1 (R Core Team 2018) and the packages *lsmeans* (for successful 238 amplexus latency and male fertilisation rate analysis), pwr (for the probability of separation 239 240 before egg laying) and *survival* (for successful amplexus latency analysis).

241 *Ethical note*

242 The capture of common toads was authorized by the Préfecture de l'Ain (DDPP01-16-145)

and by the French government (*APAFIS*#3655-2016011914372094) in accordance with the

ethical committee of Lyon 1 University. The animal care structure "EcoAquatron" (University

of Lyon) received an agreement of veterinary services (approval DSV 692661201). At the end
of the experiments, all the toads and the egg clutches were released to their original site. **Results ALAN affected reproductive behaviours of males**(Fig. 1)

251

Male body mass gain during ALAN exposure was on average of 10.66 ± 1.27 g in the control 252 group (mean \pm SEM) and did not significantly differ between the light treatments (see 253 254 Appendix Supplementary Material S4 for details). At mating, on D13, male body mass, male SVL measurement and the ratio between male and female size did not significantly differed 255 between the light treatments (see Appendix Supplementary Material S4 for details). For all 256 257 light treatments, each male performed an amplexus with the female. However, ALAN 258 strongly affected the reproductive behaviours of male common toads. First, successful amplexus latency significantly differed between pairs with an exposed male and pairs with a 259 260 control male (Fig. 1; $X_2^2 = 14.01$, P < 0.001). Males exposed to 0.1 or 5 lux took much longer to achieve successful amplexus as shown by the 3-fold increase (Tukey post hoc test P <261 0.001) and the 1.5-fold increase in latency (Tukey post hoc test P = 0.004), respectively, 262 compared with control males. No significant difference was found between males exposed to 263 0.1 or 5 lux (Tukey post hoc test P = 0.911). Male body mass negatively affected successful 264 amplexus latency ($X_1^2 = 4.58$, P = 0.032). However, this variable was not affected by female 265 body mass, by male salivary testosterone concentrations, by the interaction between light 266 treatment and male body mass, nor by the interaction between light treatment and male 267

salivary testosterone concentrations. Second, pairs with a control male never split after the 268 269 amplexus was initiated, whereas the separation rate was by 20% and 10% for males exposed to 0.1 or 5 lux, respectively ($X^2_2 = 5.99$, P = 0.050). Even if the model indicated a marginal 270 effect for the chosen significance threshold, the effect size was strong (Cohen's test $|\mathbf{h}| =$ 271 0.795), reflecting a clear biological difference. The marginal effect may likely due to low 272 statistical power. A stronger effect may have been observed with more replicates. We did not 273 274 observe an effect of male body mass, of female body mass, of male salivary testosterone concentrations, of the interaction between light treatment and male body mass or of the 275 interaction between light treatment and male salivary testosterone concentrations on the 276 277 separation rate. The time spent in amplexus before egg laying was not altered by light treatment, as pairs spent on average 55 ± 4 hours (mean \pm SEM) in amplexus before egg 278 laying. We neither observed an effect of male body mass, female body mass, nor of the 279 280 interaction between male body mass and light treatment on the time spent in amplexus before egg laying (see Appendix Supplementary Material S4 for details). 281

282 ALAN did not modify male salivary testosterone concentrations

After 12 days of exposure, the mean salivary testosterone concentration of the control group was 3.7 ± 0.69 pg mg⁻¹ of saliva (mean \pm SEM). Exposure to ALAN did not modify male salivary testosterone concentrations. We neither observed an effect of male body mass nor of the interaction between male body mass and light treatment on testosterone levels (see Appendix Supplementary Material S4 for details).

288 Exposure to ALAN reduced male fertilisation success

289

290 (**Fig. 2**)

292 When all light treatments were combined, the average clutch size laid by a female was $3347 \pm$ 272 eggs (mean \pm SEM). Male fertilisation success was affected by the interaction between 293 the male body mass and light treatment (Fig. 2; $X^2_2 = 248.5$, P < 0.001). By visually 294 interpreting the difference of slopes, we noticed that in the control and 5 lux groups, the 295 296 relationship between the fertilisation rate and male body mass was weakly negative, whereas in the 0.1 lux group, this relationship was weakly positive. Moreover, control males had an 297 298 average fertilisation rate of $75.14 \pm 6.23\%$ (mean \pm SEM), while a 25%-decrease was highlighted in the 5 lux-group compared with control males (Tukey post hoc test P < 0.001). 299

300 A difference between the 0.1 lux- and 5 lux-exposed males was also observed (Tukey post 301 hoc test P < 0.001).

302 **Discussion**

303 Here, we demonstrated that exposure to low light intensities at night during the breeding period alters both the mating behaviours and the fertilisation success of common toad males. 304 Indeed, ALAN affected mating behaviour as male common toads previously exposed to 305 ALAN needed several attempts to maintain a female in amplexus until clutch laying, although 306 all males finally paired in our experimental setup. In line with this first result, the latency to 307 pair successfully with a female was strongly increased by previous exposure to ALAN as well 308 309 as male body mass. Similarly, Drosophila melanogaster reared under 10 lux took longer to 310 initiate mating than control flies (McLay et al. 2018). In the wild, male-male competition is 311 strong in common toad with values of operational sex ratio as high as 4 or 5 males to one female (Davies & Halliday 1979). Therefore, ALAN might increase male-male competition as 312 313 females will be caught and released several times before one male manages to stay in amplexus until egg laying is complete. The strong operational sex ratio should probably 314

prevent males that release females from having another possibility to gain a female during the 315 316 breeding season. In accordance with the important role of photoperiod as a cue to initiate the reproductive season in many species, most studies have shown that ALAN exposure advances 317 318 the date of reproduction (Ikeno et al. 2014; Dominoni et al. 2016; Le Tallec et al. 2016). Other direct effects of artificial light on sexual behaviour have also been highlighted in various 319 320 species. For example, Botha et al. (2017) reported that Australian black field crickets, 321 Teleogryllus commodus, reared under a very high light intensity (100 lux) were more likely to mate than individuals reared under lower light (0, 1 and 10 lux). Further, male Blue tits, 322 *Cyanistes caeruleus*, occupying territories exposed to streetlights (> 0.06 lux) were twice as 323 324 successful in obtaining extra-pair mates as males occupying territories further away inside the forest, but no information is given about their breeding success with their social partners 325 326 (Kempenaers et al. 2010). In addition, ALAN reduced the proportion of mated females Winter 327 moth, Operophtera brumata, compared with control females, and the reduction was strongest under green ALAN than red ALAN (Van Greffen et al. 2015). These observations suggest 328 329 that ALAN may differentially affect mating behaviour depending on animal species and on spectral composition. In our study, we used white cold LED, that emit a prominent peak of 330 331 blue wavelength. It is well known that many species, including amphibians (Hailman & 332 Jaeger 1974), are sensitive to blue wavelength light, which provokes behavioural and physiological modifications (Musters et al. 2009). It would therefore, be highly interesting to 333 study the effects of low colour temperature and other spectra on the fitness of amphibians. 334

ALAN negatively affected male fertilisation success, a direct measurement of the breeding component of male fitness that does not provide information on offspring quality but instead informs on their potential number. Present data showed that large males may be more sensitive to high light intensity since under 5 lux, the fertilisation rate was 25% lower than in controls and decreased as body mass rose. The weak negative relationship between

fertilisation rate and male body mass in the controls contrasts with earlier studies showing 340 341 either a positive or no relationship between fertilisation rate and the ratio between male and female size in wild common toads (Davies & Halliday 1977; Lengagne et al. 2007). The 342 343 mechanism underlying the deleterious effect of ALAN in large males is unclear. Interestingly, it was suggested that the energetic cost of amplexus increases with size in an anuran species 344 (McLister 2003). Furthermore, we also noticed an effect of body mass on the ALAN-induced 345 346 rise in basal metabolic rate of common male toads. In males exposed to 5 lux during the night, the largest males showed the highest increase in metabolic rate (Touzot, unpublished data). 347 Further investigations are needed to clarify the link between amplexus cost and fertilisation 348 349 rate. In amphibian species, large males often have more opportunities to access reproduction than small males (Davies & Halliday 1977). By altering behaviour and fertilisation success, 350 especially in larger males, ALAN may lead to an increase in the genetic contribution of poor-351 352 quality males in populations naturally exposed to ALAN. Another scenario could be that larger males would win the competition for female access, but as being the most affected by 353 354 ALAN, fail at fertilisation, and thus contribute to the reduction in the overall reproductive success of the population. 355

The physiological mechanisms responsible for the 25% reduction in fertilisation rate 356 357 in male exposed to 5 lux at night as compared with control are unknown but may be linked to 358 spermatogenesis. Effectively, several studies have reported that exposure to continuous light causes spermatocyte degeneration (Rastogi 1976; Biswas et al. 1978). It should however be 359 noted that in anurans, spermatogenesis lasts 5 or 6 weeks on average (Jørgensen 1984). With 360 361 regard to the rather short duration of ALAN exposure used in our protocol (males have been 362 exposed to ALAN for 12 days only), a marked deleterious effect on spermatogenesis would 363 be limited. Another possibility is that the fertilisation rate decrease can likely be explained by an alteration of reproductive hormone synthesis. In vertebrates, the effects of ALAN occur at 364

all levels of the hypothalamic-pituitary-gonadal axis, leading to variations in testosterone 365 366 levels (reviewed in Ouyang et al. 2018; Brüning et al. 2018). Testosterone would not be involved in the reduced fertilisation rate of common toads as ALAN did not modify salivary 367 testosterone levels. Present observation in toads therefore contrasts with results in birds in 368 369 which ALAN modified the testosterone secretion pattern during a breeding period (Dominoni 370 et al. 2013) and increased corticosterone levels (Ouyang et al. 2015). The quality of females 371 paired with males cannot be an explanation for our results since females were not subjected to ALAN and we have controlled that female body mass and the ratio between male and female 372 size did not statistically differ among light treatments, thus excluding a female effect. Finally, 373 374 an alteration of the synchronization of gametes released and/or a reduction in the amount of sperm released by the male can likely explain the fertilisation rate reduction. Indeed, as in all 375 376 external fertilisation situations, the male must synchronize its sperm release so that it 377 coincides with egg release (Davies & Halliday 1977). Since males exposed to ALAN show increased basal metabolic rate but reduced activity-related energy expenditure (Touzot et al. 378 379 2019), sperm emission that requires smooth muscle contractions might also be impaired.

380 Conclusion

381 Although some studies have investigated the effects of ALAN on the mortality component of fitness (Rodríguez et al. 2014; Van Doren et al. 2017; Willmott et al. 2018), reports of effects 382 on the breeding component are still scarce, especially in amphibian species. This study 383 384 showed that relevant nocturnal artificial light intensities can have a major effect on the 385 breeding component of fitness of animals in the wild. Even if a reduction in fertilisation rate was observed at the rather high intensity of 5 lux, the effect although weaker probably occurs 386 387 at lower light intensities. The separation rate during amplexus, for instance, was affected at an intensity as low as 0.1 lux, which is equivalent to a full moon. If extrapolated to habitats 388 exposed to human activity-related ALAN, present experimental findings suggest that many 389

- toad populations in Europe are expected to incur reproductive losses in their breeding
- 391 habitats. A better mechanistic understanding of the effects of ALAN, particularly on
- 392 reproduction, is necessary to predict long term consequences on wildlife and propose
- appropriate management solutions in an increasingly lighted world.

394 Acknowledgements

We thank A. Clair, J. Ulmann and L. Averty for their technical assistance in the EcoAquatron;

L. Guillard for his help with the setting up of LED ribbons; and Elsevier for English

397 corrections.

398 Funding

This work was supported by the French Government [PhD grants 2017–2020], by the

400 'LABEX IMU Laboratoire d'excellence Intelligence des Mondes Urbains', by IDEX

401 Initiative d'excellence Université Lyon. The authors declare no competing or financial402 interests.

403 **References**

404 Baker BJ, Richardson JML. 2006. The effect of artificial light on male breeding-season

behaviour in green frogs, *Rana clamitans melanota*. Canadian Journal of Zoology 84: 1528–
153.

- 407 Beebee TJC. 1979. Habitats of the british amphibians (2): suburban parks and gardens.
- 408 Biological Conservation 15:241–257.
- 409 Bennie J, Davies TW, Cruse D, Inger R, Gaston KJ. 2018. Artificial light at night causes top-
- 410 down and bottom-up trophic effects on invertebrate populations. Journal of Applied Ecology

411 55:2698–2706.

- Biswas NM, Chakraborty J, Chanda S, Sanyal S. 1978. Effect of continuous Light and
 Darkness on the testocular histology of toad (*Bufo melanostictus*). Endocrinologia Japonica
 25:117–180.
- 415 Botha LM, Jones TM, Hopkins GR. 2017. Effects of lifetime exposure to artificial light at
- 416 night on cricket *Teleogryllus commodus* courtship and mating behavior. Animal Behavior
- 417 129:181–188. doi:10.1016/j.anbehav.2017.05.020.
- 418 Bradshaw WE, Holzapfel CM. 2010. Light, Time, and the Physiology of Biotic Responses to
- 419 Rapid Climate Change in Animals. Annual Review of Physiology 72:147–166.
- 420 doi:10.1146/annurev-physiol-021909-135837.
- 421 Brüning A, Hölker F, Franke S, Preuer T, Kloas W. 2015. Spotlight on fish: Light pollution
- 422 affects circadian rhythms of European perch but does not cause stress. Science of the Total

423 Environment 511:516–522. doi.org/10.1016/j.scitotenv.2014.12.094.

- 424 Brüning A, Kloas W, Preuer T, Hölker F. 2018. Influence of artificially induced light
- 425 pollution on the hormone system of two common fish species, perch and roach, in a rural
- 426 habitat. Conservation Physiology 6. doi: 10.1093/conphys/coy016.
- 427 Buchanan BW. 2006. Observed and potential effects of artificial night lighting on anuran
- 428 amphibians. Pages 192–220 in Rich C, Longcore T, editors. Ecological consequences of
- 429 artificial night lighting. Washington D.C.: Island Press.
- 430 Czarnecka M, Kakareko T, Jermacz L, Pawlak R, Kobak J. 2019. Combined effects of
- 431 nocturnal exposure to artificial light and habitat complexity on fish foraging. Science of the
- 432 Total Environment 684:14–22. doi.org/10.1016/j.scitotenv.2019.05.280.
- 433 Davies NB, Halliday TR. 1977. Optimal mate selection in the toad *Bufo bufo*. Nature 269:56–
 434 58.

- 435 Davies NB, Halliday TR. 1978. Deep croaks and fighting assessments in toads *Bufo bufo*.
 436 Nature 274:683–685.
- 437 Davies NB, Halliday TR. 1979. Competitive mate searching in male common toads, *Bufo*438 *bufo*. Animal Behavior 27:1253–1267.
- de Jong M, Ouyang JQ, Da Silva A, van Grunsven RHA, Kempenaers B. Visser ME,
- 440 Spoelstra K. 2015. Effects of nocturnal illumination on life-history decisions and fitness in
- two wild songbird species. Philosophical transactions of the royal society of London B,
- **442** 370:20140128.
- 443 Desouhant E, Gomes E, Mondy N, Amat I. 2019. Mechanistic, ecological and evolutionary
- 444 consequences of artificial light at night for insects: review and prospective. Entomologia
- 445 Experimentalis et Applicata 1–22. doi:10.1111/eea.12754.
- 446 Desprat J, Lengagne T, Dumet A, Desouhant E, Mondy N. 2015. Immunocompetence
- 447 handicap hypothesis in tree frog: trade-off between sexual signals and immunity? Behavioral
- 448 Ecology 264:1138-1146. doi:10.1093/beheco/arv057.
- 449 Dominoni DM, Borniger JC, Nelson RJ. 2016. Light at night, clocks and health: from humans
- 450 to wild organisms. Biology Letters 12:20160015. doi:10.1098/rsbl.2016.0015.
- 451 Dominoni D, Quetting M, Partecke J. 2013. Artificial light at night advances avian
- 452 reproductive physiology. Proceedings of the Royal Society of London Part B 280:20123017.
- 453 doi:10.1098/rspb.2012.3017.
- 454 Duffy JP, Bennie J, Durán AP, Gaston KJ. 2015. Mammalian ranges are experiencing erosion
- 455 of natural darkness. Scientific reports 5:12041. doi:10.1038/srep12042.

- 456 Duffy JP, Bennie J, Durán AP, Gaston KJ. 2015. Mammalian ranges are experiencing erosion
- 457 of natural darkness. Scientific reports 5:12041. doi:10.1038/srep12042.
- 458 Falchi F, Cinzano P, Duriscoe D, Kyba CCM, Elvidge CD, Baugh K, Portnov BA, Rybnikova
- 459 NA, Furgoni R. 2016. The new world atlas of artificial night sky brightness. Science
- 460 Advances 2 e1600377. doi:10.1126/sciadv.1600377.
- Fobert EK, Burke da Silva K, Swearer SE. 2019. Artificial light at night causes reproductive
 failure in clownfish. Biology letters 15:20190272.
- 463 Gaston KJ, Bennie J, Davies TW, Hopkins J. 2013. The ecological impacts of nighttime light
- 464 pollution: a mechanistic appraisal. Biological Reviews 88:912–927.
- 465 Gaston KJ, Bennie J. 2014b. Demographic effects of artificial nighttime lighting on animal
- 466 populations. Environmental Reviews 22:323–330.
- 467 Gaston KJ, Duffy JP, Gaston S, Bennie J, Davies TW. 2014a. Human alteration of natural
- light cycles: causes and ecological consequences. Oecologia 176:917–931.
- 469 Grant RA, Chadwick EA, Halliday T. 2009. The lunar cycle: a cue for amphibian
- 470 reproductive phenology? Animal Behaviour 78: 349–357.
- 471 Grimm NB, Faeth SH, Golubiewski NE, Redman CL, Wu J, Bai X, Briggs JM. 2008. Global
- 472 Change and the Ecology of Cities. Science 319:756–760.
- 473 Grubisic M, Singer G, Bruno MC, van Grunsven RHA, Manfrin A, Monaghan MT, Hölker F.
- 474 2017. Artificial light at night decreases biomass and alters community composition of benthic
- 475 primary producers in a sub-alpine stream. Limnology and Oceanography.
- 476 doi:10.1002/lno.10607.

- Hailman JP, Jaeger RG. 1974. Phototactic responses to spectrally dominant stimuli and use of
 colour vision by adult anuran amphibians: a comparative survey. Animal Behavior 22:757–795.
- 479 Hilty J, Merenlender A. 2000. Faunal indicator taxa selection for monitoring ecosystem health.
 480 Biological Conservation. 92:185-197.
- 481 Hölker F, Wolter C, Perkin EK, Tockner K. 2010. Light pollution as a biodiversity threat.
 482 Trends in Ecology & Evolution 25:681–682.
- 483 Ikeno T, Weil ZM, Nelson RJ. 2014. Dim light at night disrupt the short-day response in

484 Siberian hamsters. General and Comparative Endocrinology 197:56–54.

- 485 Janin A, Léna JP, Deblois S, Joly P. 2012. Use of stress-Hormone levels and habitat selection
- 486 to assess functional connectivity of a landscape for an amphibian. Conservation Biology
 487 265:923–931.
- 488 Jørgensen CB. 1984. Growth and Reproduction. In: Feder ME, Burggren WW, eds.
- Environmental Physiology of the Amphibians. The University of Chicago Press, pp 439–467.
- 490 Kempenaers B, Borgström P, Loës P, Schlicht E, Valcu M. 2010. Artificial night lighting
- 491 affects dawn song, extra-pair siring success and lay date in songbirds. Current Biology
- 492 20:1735–1739. doi:10.1016/j.cub.2010.08.028.
- Knop E, Zoller L, Ryser R, Gerpe C, Hörler M, Fontaine C. 2017. Artificial light at night as a
 new threat to pollination. Nature 548:206–209.
- 495 Kutsukake N, Ikeda K, Honma S, Teramoto M, Mori Y, Hayasaka I, Yamamoto R, Ishisa T,
- 496 Yoshikawa Y, Hasegawa T. 2009. Validation of salivary cortisol and testosterone assays in
- 497 chimpanzees by liquid chromatography-tandem mass spectrometry. American Journal of
- 498 Primatology 71:696–706.

- 499 Kyba CCM, Kuester T, de Miguel AS, Baugh K, Jechow A, Hölker F, Bennie J, Elvidge CD,
- 500 Gaston KJ, Guanter L. 2017. Artificially lit surface of Earth at night increasing in radiance
- and extent. Sciences Advances 3:e1701528.
- Lengagne T, Arthaud F, Cormier M, Joly P. 2007. Cost of sexually embracing a large female
- 503 offset by the number of eggs fertilized for small male *Bufo bufo L*. Biological Journal of the
- 504 Linnean Society 92, 755–762.
- 505 Le Tallec T, Perret M, Théry M. 2013. Light pollution modifies the expression of daily
- rhythms and behavior patterns in a nocturnal primate. PLoS One 8: e79250.
- 507 doi:10.1371/journal.pone.0079250.
- Le Tallec T, Théry M, Perret M. 2016. Melatonin concentration and timing of seasonal

reproduction in male mouse lemurs *Microcebus murinus* exposed to light pollution. Journal of
Mammalogy 973:753–760.

- 511 Le Tallec T, Théry M, Perret M. 2015. Effects of light pollution on seasonal estrus and daily
- rhythms in a nocturnal primate. Journal of Mammalogy 96:438–445.
- 513 May D, Shidemantle G, Melnick-Kelley Q, Crane K, Hua J. 2019. The effect of intensified
- 514 illuminance and artificial light at night on fitness and susceptibility to abiotic and biotic
- stressors. Environmental Pollution 251:600–608.
- 516 McLay LK, Nagarajan-Radha V, Green MP, Jones TM. 2018. Dim artificial light at night
- 517 affects mating, reproductive output and reactive oxygen species in *Drosophila melanogaster*.
- 518 Journal of Experimental Zoology 1–10. doi:10.1002/jez.2164.
- 519 McLister JD. 2003. The metabolic cost of amplexus in the grey tree frog (*Hyla versicolor*):
- assessing the energetics of male mating success. Canadian Journal of Zoology 81:388–394.

Musters CJM, Snelder DJ, Vos P. 2009. The effects of coloured light on nature: A literature
study of the effects of part of the spectrum of artificial light on species and communities.
Institute of Environmental Sciences, Leiden University. 43 pp.

524 Newman RC, Ellis T, Davison PI, Ives MJ, Thomas RJ, Griffiths SW, Riley WD. 2015. Using

novel methodologies to examine the impacts of artificial light at night on the cortisol stress

response in dispersing Atlantic salmon *Salmo salas L. fry*. Conservation Physiology 3: doi:

527 10.1093/conphys/cov051.

528 Norberg B, Brown CL, Halldorsson O, Stensland K, Björnsson BT. 2004. Photoperiod regulates

529 the timing of sexual maturation, spawing, sex steroid and thyroid hormones profiles in the

530 Atlantic cod (*Gadus morhua*). Aquaculture 229:451–467.

- 531 Ouyang JQ, Davies S, Dominoni D. 2018. Hormonally mediated effects of artificial light at
- night on behavior and fitness: linking endocrine mechanism with function. Journal of

533 Experimental Biology 221. doi: 10.1242/jeb.156893.

- 534 Ouyang JQ, de Jong M, Hau M, Visser ME, van Grunsven RHA, Spoelstra K. 2015. Stressfull
- colours: corticosterone concentrations in a free-living songbird vary with the spectral

composition of experimental illumination. Biology Letters 11: 20150517.

- 537 Pulgar J, Zeballos D, Vargas J, Aldana M, Manriquez P, Manriquez K, Quijón PA,
- 538 Widdicombe S, Anguita C, Quintanilla D, et al. 2019. Endogenous cycles, activity patterns
- and energy expenditure of an intertidal fish is modified by artificial light pollution at night
- 540 ALAN. Environmental Pollution. doi:10.1016/j.envpol.2018.10.063.
- 541 R Core Team. 2018. R: A language and environment for statistical computing. R foundation
- 542 for statistical computing, Vienna, Austria. URL <u>https://www.R-project.org/</u>.

- 543 Rastogi RK. 1976. Seasonal cycle in anuran amphibia testis: the endocrine and environmental
 544 controls. Italian Journal of Zoology 43:151–172.
- Reading CJ, Clarke RT. 1983. Male breeding behaviour and mate acquisition in the Common
 toad, *Bufo bufo*. Journal of Zoology 201:237–246.
- 547 Robert KA, Lesku JA, Partecke J, Chambers B. 2015. Artificial light at night desynchronizes
- 548 strictly seasonal reproduction in a wild mammal. Proceedings of the Royal Society of London
- 549 Part B 282:20151745. doi:10.1098/rspb.2015.1745.
- 550 Rodríguez A, Burgan G, Dann P, Jessop R, Negro JJ, Chiaradia A. 2014. Fatal attraction of
- 551 Short-Tailed Shearwaters to Artificial lights. PLoS One 9: e110114.
- 552 doi:10.1371/journal.pone.0110114.
- 553 Sanders D, Kehoe R, Cruse D, van Veen KJK, Gaston KJ. 2018. Low levels of artificial light
- at night strengthen Top-Down control in insect food web. Current Biology 28:1–5.
- 555 doi:10.1016/j.cub.2018.05.078.
- 556 Secondi J, Dupont V, Davranche A, Mondy N, Lengagne T, Théry M. 2017. Variability of
- surface and underwater nocturnal spectral irradiance with the presence of clouds in urban and
- peri-urban wetlands. PLoS One 12: e0186808. doi:10.1371/journal.pone.0186808.
- 559 Touzot M, Teulier L, Lengagne T, Secondi J, Théry M, Libourel PA, Guillard L, Mondy N.
- 560 2019. Artificial light at night disturbs the activity and energy allocation of the common toad
- during the breeding period. Conservation Physiology 71: coz002.
- 562 doi:10.1093/conphys/coz002.
- 563 Tuxbury SM, Salmon M. 2005. Competitive interactions between artificial lighting and
- natural cues during seafinding by hatchling marine turtles. Biological Conservation 121:311–
- 565 316.

- Van Doren BM, Horton KG, Dokter AM, Klinck H, Elbin SB, Farnsworth A. 2017. High-566
- 567 intensity urban light installation dramatically alters nocturnal bird migration. PNAS. doi:10.1073/pnas.1708574114. 568
- Welbers AAMH, van Dis NE, Kolvoort AM, Ouyang J, Visser ME, Spoelstra K, Dominoni
- DM. 2017. Artificial light at night reduces daily energy expenditure in breeding great tits 570
- 571 (Parus major). Frontiers in Ecology and Evolution 5: 55. doi:10.3389/fevo.2017.00055.
- 572 Wells KD. 1977. The social behaviour of anuran amphibians. Animal Behavior 25:666-693.
- 573 Willmott N, Henneken J, Selleck CJ, Jones TM. 2018. Artificial light at night alters life
- history in a nocturnal orb-web spider. PeerJ 6:e5599. doi:10.7717/peerj.5599. 574
- Yovanovich CA, Koskela SM, Nevala N, Kondrashev SL, Kelber A, Donner K. 2017. The 575
- dual rod system of amphibians supports colour discrimination at the absolute visual threshold. 576
- Philosophical Transactions of the Royal Society of London B 372:20160066. 577
- doi:10.1098/rstb.2016.0066. 578
- 579

Fig. 1. Successful amplexus latency (hour) of males previously exposed for 12 days to 0.01
(black), 0.1 (grey) or 5 (white) lux (horizontal line: median value; box ends: upper and lower
quartiles; whiskers: upper and lower 1.5 x interquartile range (IQ); dot: upper dot beyond
upper 1.5 x IQ). * < 0.005 ** < 0.001.

Fig. 2. Fertilisation rate (%) of males previously exposed for 12 days to 0.01 (black dots and solid line), 0.1 (grey dots and dashed line) or 5 (white dots and dotted line) lux depending on their body mass. Points represent observed data and the lines represent the predictions of the statistical model (see Statistical Analysis). Shaded areas represent SE calculated from the statistical model.

590





613 Fig. 2.