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1	Effect of Sb on Precipitation of Biogenic Minerals during
2	the Reduction of Sb-bearing Ferrihydrites
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27 ABSTRACT

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29 Antimony (Sb) is a naturally occurring element; it is enriched in the environment by 30 anthropogenic activities. Like other metalloid species, Sb partitions to mineral phases such as 31 oxyhydroxides. In reducing environments, Fe(III) may serve as a terminal electron acceptor 32 during dissimilatory iron reduction leading to its transformation. Relatively little is known 33 concerning the effect of Sb(V) on the precipitation of biogenic minerals in relation to 34 microbiologically mediated redox reactions. To further our understanding, Sb-bearing 35 ferrihydrites with variable molar ratios [Sb/(Fe + Sb)] were incubated in the presence of 36 Shewanella oneindensis MR-1. A combination of wet chemistry and solid analysis techniques 37 (XRD, Mössbauer and Raman spectroscopies) was used to characterize the reactions.

38 The Sb loading affected the rate and the extent of bio-reduction compared with pure 39 ferrihydrite. Only a minor fraction of the total Sb was released into the solution by the end of 40 the incubation period, suggesting that the metalloid partitioned mainly in a newly formed 41 phase. Furthermore, XPS analyses showed the presence of Sb(V) and Sb(III) species on the 42 biogenic minerals. Magnetite was the main biogenic precipitate in the absence of Sb(V). 43 Increasing of the molar ratios (Sb/(Fe + Sb)) resulted mainly in the precipitation of carbonated 44 green rust and goethite. Abiotic green rust synthesis carried out in the presence of Sb(V) 45 indicated the latter's stabilizing effect on the green rust structure, as for phosphate species. 46 Thus, it is likely that Sb preserves biogenic green rust, hindering its transformation to more 47 thermodynamically stable phases.

48

51 1. INTRODUCTION

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53 Natural deposits (e.g. ores, hot springs) and anthropogenic activities (e.g. mining, road 54 traffic emissions) have led to a significant dispersion of inorganic elements, including antimony (Sb), in the lithosphere (Liu et al., 2010; Anderson, 2012; Okkenhaug et al., 2016; 55 56 Li et al., 2018; He et al., 2019). There is a small but growing body of evidence that 57 demonstrates the deleterious effect of Sb on biota (Gebel, 1997; Oorts et al., 2008; An and 58 Kim, 2009). As a result, both the United States Environmental Protection Agency (U.S. EPA) 59 and the European Union (E.U.) have listed Sb as a "priority pollutant" (Official Journal L 129, 1976; United States Environmental Protection Agency), which implies the need to assess 60 61 the behavior Sb(V) in natural settings. In the environment, antimonite $[Sb(OH)_3]$ and antimonate $[Sb(OH)_6]$ are the dominant chemical species (Filella et al., 2002; Wilson et al., 62 2010). Antimony has been observed interacting with soil components, in particular, iron 63 64 oxides (e.g. goethite, lepidocrocite), and it is positively correlated with Fe content (Craw et al., 2004; Mitsunobu et al., 2010a; Takahashi et al., 2010). 65

66 Iron oxides display a large surface area and high chemical reactivity given their crystallinity and porosity (Cornell and Schwertmann, 2003). They therefore contribute 67 significantly to the speciation and the fate of elements (e.g. As, Sb) that play a central role in 68 69 various biogeochemical cycles in soils and sediments (Dixit and Hering, 2003; Borch and Fendorf, 2007; Mitsunobu et al., 2008a; Frierdich and Catalano, 2012a; Ilgen and Trainor, 70 71 2012; Okkenhaug et al., 2013; Mallet et al., 2013; Perez et al., 2019). Due to their large specific surface area, iron oxides are considered to be an effective sink for Sb in shooting 72 73 range soils (Okkenhaug et al., 2016; Okkenhaug et al., 2018). Thus, the properties of these 74 iron oxides have led to the consideration of Fe-based amendments as a potential strategy for the removal and the immobilization of Sb in contaminated areas. In freshwater, Sb is found to 75

be much more mobile in low Fe systems, suggesting that the adsorption or the co-76 precipitation with iron oxides is a critical factor controlling Sb dynamics (Mitsunobu et al., 77 2010a; Hockmann et al., 2014; Burton et al., 2020). Furthermore, X-ray Absorption Fine 78 Structure (XAFS) spectroscopic analyses have demonstrated the incorporation of Sb(V) into 79 80 the crystallographic structure of iron oxides synthesized in the laboratory, such as goethite 81 and ferrihydrite (Mitsunobu et al., 2013; Karimian et al., 2019b; Burton et al., 2020). This 82 structural incorporation has also been reported to occur in natural soil, suggesting that the 83 substitution of Fe in iron minerals is widespread in the environment (Mitsunobu et al., 84 2010b).

In subsurface environments, iron oxides are also used as electron acceptors for 85 86 dissimilatory iron-reducing bacteria (DIRB) (Lovley, 1993; Coates et al., 1996; Maier and Myers, 2001; Nealson et al., 2002; Sung et al., 2012). The bio-reduction of Sb-bearing iron 87 88 oxide can lead to the potential mobilization into the aqueous phase and (or) the sorption onto biogenic precipitates of the previously retained Sb(V). DIRB such as Geobacter sp. and 89 90 Shewanella sp. couple the oxidation of an organic (e.g. formate, pyruvate, lactate) or 91 inorganic (e.g. H₂) electron source with the reduction of Fe(III) during anaerobic respiration 92 (Lovley et al., 1993; Lovley, 2000). Subsequently, the Fe(II) generated by bio-reduction is 93 released into solutions and/or sorbed on mineral surfaces, which leads to the transformation of 94 Fe-bearing minerals, such as magnetite (Fe₃O₄), siderite (FeCO₃), green rust (GR), goethite (α -FeOOH) and lepidocrocite (γ -FeOOH), depending on the biogeochemical conditions 95 (Kukkadapu et al., 2005; Zegeve et al., 2007; O'Loughlin et al., 2007; Ona-Nguema et al., 96 2009; Eusterhues et al., 2014; Muehe et al., 2016). 97

98 Of these biogenic minerals, GRs are Fe(II-III)-layered double-hydroxysalts constituted 99 of positively charged layers $\{FeII_{(1-x)} FeIII_x (OH)_2\}^{x+}$, which alternate with negatively 100 charged interlayers that include m water molecules per Aⁿ⁻ anion $\{(x/n)A^{n-} \cdot (mx/n)H2O\}^{x-}$

101 (Hansen, 1989; Génin et al., 1996). Two groups of GRs are identified based on X-ray 102 diffraction (XRD) patterns: green rust 1 (GR1) and green rust 2 (GR2) (Bernal, 1959). Their 103 structure and composition depend upon the specific anions they incorporate. GR1 is arranged in rhombohedral geometry and inserts planar or spherical anions (e.g. Cl^{-} , CO_3^{2-}), whereas 104 GR2 is arranged in a hexagonal geometry and inserts three-dimensional anions (e.g. SO_4^{2-} , 105 SeO_4^{2-}) (Hansen et al., 1994; Abdelmoula et al., 1996; Refait et al., 1998; Refait et al., 2000; 106 107 Simon et al., 2003). GRs have demonstrated their potential ability to interact with Sb(V) with 108 an inner sphere complex, thus reducing the metalloid's mobility (Mitsunobu et al., 2009).

109 Relatively little is known concerning the effect of Sb(V) on the precipitation of 110 biogenic minerals such as GRs in relation to microbiologically mediated redox reactions. A 111 recent study has shown that Sb(V) favors feroxyhyte formation over goethite during the bio-112 reduction of Sb(V)-ferrihydrite (Burton et al., 2020). The incorporation of Sb(V) in the iron oxides' structure was also highlighted. To broaden our understanding of Fe and Sb 113 114 biogeochemical cycles, it is pertinent to investigate how various Sb concentrations impact iron oxide bio-reduction processes and the precipitation of biogenic minerals. Furthermore, 115 the interaction of Sb(V) with $GR1(CO_2^{3-})$ resulting from the bio-reduction of an Sb(V)-116 117 bearing iron oxide has not been reported previously.

Our study aims to investigate the impact of Sb(V) loading on the nature of biogenic minerals, such as GRs precipitated during the bio-reduction of Sb(V) bearing ferrihydrites. To achieve this objective, we incubated Sb bearing ferrihydrites with variable molar ratios (Sb/(Fe + Sb) with *Shewanella oneindensis* MR-1. We then assessed the following: (i) the influence of Sb content on the bio-reduction of ferrihydrite (i.e. rate and extent of reduction) and (ii) the nature of biogenic minerals. Furthermore, we synthesized abiotic $GR1(CO_3^{2-})$ in the presence of Sb(V) to examine the Sb(V) effect on this mineral formation and stabilization.

126127 2. METHODS

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129 All reagents and acids were of analytical grade and solutions were prepared using 18.2 M Ω 130 cm⁻¹ deionized water. Anaerobic experiments (bio-reduction, abiotic GR synthesis and aging), 131 sample collection (*i.e.* by centrifugation) and sample drying were conducted under strictly 132 anaerobic conditions in a re-circulating anoxic chamber (Coy Laboratory Products Inc.) in 133 which trace amounts of oxygen were scavenged by palladium catalysts. A 98% N_2 + 2% H_2 134 atmosphere was maintained by the automatized injection of the gas mixture. When needed, 135 the samples were removed from the anaerobic chamber in an air-tight sample holder for solid 136 analyses.

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138 2.1 Synthesis of Sb(V)-bearing ferrihydrite

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140 Ferrihydrite (Fh hereafter) was synthesized by precipitating 0.4 M FeCl₃ with 4 M NaOH 141 according to the method adapted from Cornell and Schwertmann (2003). When the pH 142 approached the desired endpoint of the hydrolysis, a solution of 0.5 M NaOH was used to adjust the pH to = 7.2 ± 0.1 . Antimony co-precipitated Fhs were synthesized in similar 143 144 fashion, except for the addition of K[Sb(OH)₆] stock solution to the ferric salt before 145 hydrolysis at Sb/(Fe + Sb) molar ratios of 0.04, 0.06 and 0.1 (Fh Sb0.04, Fh Sb0.06 and 146 Fh Sb0.1 hereafter), following a modification of the method used by Mitsunobu et al. 147 (2010b). All suspensions were allowed to equilibrate for 4 hours after the synthesis and the 148 pH was readjusted to pH 7.2 with 0.05 M NaOH if necessary. Thereafter, the co-precipitates 149 were washed six times with Milli-Q water by centrifugation and freeze-dried.

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154 2.2 Bio-reduction of Sb(V)-bearing ferrihydrites

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156 Frozen Shewanella oneindensis MR-1 cells from a stock (frozen in 20% glycerol at -80 157 °C) were revived under aerobic conditions on tryptic soy agar. They were sub-cultured once, 158 and the colonies used to prepare a suspension to inoculate 700 mL of trypcase soy broth. Cells 159 were grown, continuously agitated at 300 rpm and 30°C, harvested after 14 hours of culture 160 by centrifugation (10 000 x g at 20°C for 10 minutes), washed twice with sterile 0.9% NaCl, 161 and concentrated in the same medium. The cells were purged for 30 minutes by bubbling with 162 filter-sterilized N₂ filtration through a membrane of pore size 0.2 µm, and used to inoculate 163 batches with ferrihydrite.

164 The bio-reduction were performed in a culture medium composed of 22 mM NH₄Cl, 1.5 M NaCl, 1.2 mM KCl, 1.1 mM MgSO₄•7H₂O, 0.71 mM nitrilotriacetic acid (NTA), 0.67 M 165 166 CaCl₂, 0.27 mM MnSO₄•H₂O, 86 µM ZnCl₂, 38 µM CoSO₄•7H₂O, 32 µM FeSO₄•7 H₂O, 167 9.3µM Na2MoO4 •2 H2O, 9.1 µM NiCl2 •6H2O, 6.8 µM Na2WO4•2 H2O, 3.6 µM CuSO4•5 168 H₂O, 1.9 µM AlK(SO₄)₂•12 H₂O, 1.5 µM H₃BO₃. The medium was heat-sterilized and purged 169 with filter-sterilized N₂, and dispensed into sterile 100 mL serum bottles in the presence of 0.5 170 g of Sb-bearing Fh, and crimp sealed using butyl rubber stoppers. Sodium methanoate, as the 171 sole electron source, and anthraquinone-2,6-disulfonate (AQDS), as an electron shuttle, were 172 sterilized by filtration (filter pore size 0.2 µm), purged with filter-sterilized N₂, and introduced in the culture medium under aseptic conditions in order to achieve 50 mM (i.e. excess of \approx 173 55% with respect to Fe(III)) and 100 µM respectively in each 80 mL assay. S. oneindensis 174 cell suspensions were added to obtain final concentrations of 1.0×10^8 cell mL⁻¹. The initial 175 pH after mixing all the components was 7.2 ± 0.1 . The assays were incubated in the dark for 176

177 25 days, in triplicates. The control experiments were cell-free but otherwise identical to the178 biotic assays.

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180 2.3 **Precipitation of abiotic GR1(CO_3^2)**

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182 Carbonate green rusts were synthesized according to the method adapted from Bocher 183 et al. (2004). Ferrous sulfate heptahydrate FeSO₄ and ferric sulfate Fe₂(SO₄)₃ were dissolved 184 in 30 mL of ultra-pure water with a Fe(II) / Fe(III) ratio of 2:1 and titrated with 30 mL of an appropriate concentration of a basic solution (NaOH, $[OH^-] / [Fe_t] = 2$, and Na₂CO₃, $[CO_3^{2-}] / [Fe_t] = 2$ 185 186 $[Fe_t] = 7/6$). Other sets of GR synthesis were conducted by mixing either (i) phosphate solution (PO₄³⁻ / Fe_t = 0.016) or (ii) Sb^V solution (Sb(V) / Fe_t = 0.09) with the ferric/ferrous 187 188 solution before the co-precipitation. Phosphate was used to hinder the transformation of GR 189 and to stabilize the mineral. A further synthesis was run with the Sb(V) solution being added 190 directly after GR precipitation (~ 10 min) in order to obtain a Sb(V) / Fet ratio of 0.09. The 191 different precipitates were characterized after one week of aging in an anaerobic chamber.

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- 193 2.4 Analytical procedure
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- 195 2.4.1 Chemical analyses
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197 The Sb/(Fe + Sb) molar fractions of the initial Fh minerals were measured by 198 Inductively Coupled Plasma Optical Emission Spectroscopy (ICP-OES) after being dissolved 199 in acid, (6 M HCl) (Amstaetter et al., 2012).

At each time point, samples (0.5 mL) were removed from the bio-reduction assays and extracted with 4 M HCl in an anaerobic chamber in order to quantify total Fe (Fe_{tot}) and total Fe(II) (Fe^{II}_{tot}) using the ferrozine method (Viollier et al., 2000). The apparent initial reduction rates were computed from the first derivative of a nonlinear curve fit (1)

204
$$\%(Fe^{II}_{tot}/Fe_{tot})_t = \%(Fe^{II}_{tot}/Fe_{tot})_{max} [1 - exp(-k_{obs}t)] (1)$$

where $%(Fe^{II}_{tot}/Fe_{tot})_{t}$ is the (Fe^{II}_{tot}/Fe_{tot}) (%) produced at time t, $%(Fe^{II}_{tot}/Fe_{tot})_{max}$ is the maximum $%(Fe^{II}_{tot}/Fe_{tot})$ observed at the end of the reduction period (i.e. 25 days), and k_{obs} is the pseudo first-order constant.

At the end of the incubation period, 1 mL was withdrawn from the assays, filtered 208 209 through 0.22 µm pore filters, and placed in 2 M HCl in an anaerobic chamber to measure the 210 concentration of Sb in solution (Sb_{sol}) by Hydride Generation - Atomic Fluorescence Spectroscopy (HG-AFS). Stibnite was generated in a continuous flow system (PS Analytical's 211 Millennium ExcaliburTM) using 0.7%, m/v NaBH₄ solution (stabilized with 0.1 M NaOH) and 212 25% v/v HCl as carrier solution. The sample solution (flow rate: 9 mL min⁻¹) and carrier 213 solution (NaBH₄ flow rate: 4.5 mL min⁻¹; HCl flow rate: 9 mL min⁻¹) were pumped 214 215 simultaneously into the sample valve for mixing. The standard and the sample solution were 216 prepared in a 30% v/v HCl medium and 1% m/v KI / 0.2% m/v ascorbic acid mix to achieve 217 the pre-reduction of the Sb(V). The mixture was transferred into the gas-liquid separator, and 218 the generated gaseous hydrides were transported to the atomizer flame of the AFS. A boosted 219 discharge hollow cathode lamp (BDHCL) for Sb was used throughout the experiments.

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- 226 2.4.2 Solid analyses

228 2.4.2.1 X-Ray Diffraction (XRD)

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230	The X-Ray Diffraction (XRD) data of the Fh and Fh_Sb series and the biogenic
231	minerals were collected using a Bruker D8 diffractometer equipped with a monochromator
232	and position-sensitive detector. The X-ray source was a Co anode ($\lambda = 0.17902$ nm). The
233	diffractogram was recorded in the 3 – 84° 2 θ range, with a 0.0359° step size and collection
234	time of 3 seconds per point. Crystalline size was calculated using the Scherrer equation (2)
235	$D = K\lambda/\beta \cos\theta (2)$
236	where D is the mean size of the crystalline domains, K is a dimensionless shape factor (0.94),
237	λ is the wavelength of the X-ray radiation, β is the line broadening at the full width at half
238	maximum (FWHM) and θ is the Bragg angle of the particular peak of interest.
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240	2.4.2.2 Transmission Mössbauer Spectroscopy (TMS)
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241 242	Transmission Mössbauer Spectroscopy (TMS) was performed using a ⁵⁷ Co source to
	Transmission Mössbauer Spectroscopy (TMS) was performed using a ⁵⁷ Co source to identify and quantify iron phases and magnetic ordering formed after the bio-reduction of Fh,
242	
242 243	identify and quantify iron phases and magnetic ordering formed after the bio-reduction of Fh,
242 243 244	identify and quantify iron phases and magnetic ordering formed after the bio-reduction of Fh, Fh_Sb(0.04), Fh_Sb(0.06) and Fh_Sb(0.1). ⁵⁷ Fe Mössbauer spectra were collected using a
242243244245	identify and quantify iron phases and magnetic ordering formed after the bio-reduction of Fh, Fh_Sb(0.04), Fh_Sb(0.06) and Fh_Sb(0.1). ⁵⁷ Fe Mössbauer spectra were collected using a conventional spectrometer in transmission geometry coupled with an Advances Research
 242 243 244 245 246 	identify and quantify iron phases and magnetic ordering formed after the bio-reduction of Fh, Fh_Sb(0.04), Fh_Sb(0.06) and Fh_Sb(0.1). ⁵⁷ Fe Mössbauer spectra were collected using a conventional spectrometer in transmission geometry coupled with an Advances Research Systems (USA) cold head cryostat, equipped with in-house made vibration isolation stand.
 242 243 244 245 246 247 	identify and quantify iron phases and magnetic ordering formed after the bio-reduction of Fh, Fh_Sb(0.04), Fh_Sb(0.06) and Fh_Sb(0.1). ⁵⁷ Fe Mössbauer spectra were collected using a conventional spectrometer in transmission geometry coupled with an Advances Research Systems (USA) cold head cryostat, equipped with in-house made vibration isolation stand. The 50 mCi ⁵⁷ Co in Rh matrix radioactive source was mounted in a constant acceleration
 242 243 244 245 246 247 248 	identify and quantify iron phases and magnetic ordering formed after the bio-reduction of Fh, Fh_Sb(0.04), Fh_Sb(0.06) and Fh_Sb(0.1). ⁵⁷ Fe Mössbauer spectra were collected using a conventional spectrometer in transmission geometry coupled with an Advances Research Systems (USA) cold head cryostat, equipped with in-house made vibration isolation stand. The 50 mCi ⁵⁷ Co in Rh matrix radioactive source was mounted in a constant acceleration velocity transducer. Solid samples were prepared under an N ₂ atmosphere in an anoxic glove

252 room temperature and low-temperature studies by comparing the hyperfine split obtained at 253 low temperature with the spectra measured at room temperature. The hyperfine interaction 254 parameters were determined by fitting the experimental spectra by a least-squares method 255 using Recoil software (Lagarec and Rancourt, 1998) with Voigt-based fitting analysis or the 256 Lorentzian model, except in the case of the appearance of a magnetic order Fe(II) phases 257 (siderite), whereas the full static Hamiltonian site model was required at very low temperature 258 (15 K) to solve the mixed hyperfine interactions. Center shifts were reported with respect to 259 that of 25 μ m-thick α -Fe foil at room temperature.

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261 2.4.2.3 Raman Spectroscopy

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Raman spectra were collected on a Renishaw inViaTM Qontor® microspectrometer equipped with a confocal microscope and an Olympus X50 objective (numerical aperture = 0.55). The samples were prepared in an anaerobic chamber and enclosed in an airtight cell with sapphire windows. 785 nm exciting radiation was used with a laser power below 0.1 mW for all samples, in order to prevent their degradation. The spot area was a few μ m². Several locations were probed on each sample. The spectral resolution was about 4 cm⁻¹ and the precision on the wavenumber was below than 1 cm⁻¹.

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271 2.4.2.4 X-ray Photoelectron Spectroscopy (XPS)

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The surface physicochemical properties of Fh_Sb(0.1) were characterized before and after bio-reduction using X-ray Photoelectron Spectroscopy (XPS). Spectra were recorded on a Kratos Axis Ultra X-ray photoelectron spectrometer equipped with a monochromatic Al Kα source (1486.6 eV) operated at 120 W. Powder samples were deposited onto a copper 277 adhesive tape in an anaerobic chamber and kept in sealed aluminum pouches that were reopened in an anaerobic glove box connected to the UHV preparation chamber at 10^{-7} mb. The 278 analysis area was equal to $300 \times 700 \ \mu\text{m}^2$. The spectra were collected at a take-off angle of 279 90° and the pressure in the analysis chamber was approximately 10^{-9} mbar. A pass energy of 280 281 20 eV was used for the high-resolution spectra. Charge compensation was applied by flooding 282 low-energy electrons. Binding energies were calibrated by assigning the adventitious carbon 283 C 1s peak to 284.6 eV. Curve fitting was performed using a Gaussian/Lorentzian (70/30) peak 284 shape after Shirley's background subtraction and using X-vision 2.2.11 software.

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286 2.4.2.5 Magnetism

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288 A Quantum Design Superconducting Quantum Interference Device (SOUID) was used to 289 perform Zero field cooled (ZFC) and field cooled (FC) measurements. The ZFC procedure 290 consisted of demagnetizing the samples at 300 K using an oscillatory decaying field from 1 T 291 to zero field (resp. magnetized under 7.5 mT), and then cooling them to 2 K at a rate of 1 292 K/min. ZFC magnetic moment was then measured while the temperature was ramped back up 293 from 2 K to 300 K at a rate of 1 K/min under 7.5 mT constant field. The FC procedure 294 consisted of cooling down the sample from 300K under 7.5 mT and then measuring moment 295 while heating it from 2 K back up to 300 K.

296

3. RESULTS

300

301 3.1 Sb-bearing ferrihydrite characterization

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303 The Fe and Sb contents of the co-precipitates were measured by ICP-OES; the 304 concentrations obtained were approximately equal to the initial molar ratio of Sb/(Sb + Fe) (i.e. 0.04, 0.06 and 0.1) used to precipitate the solids. X-ray diffraction was used to 305 306 characterize the co-precipitated phases (Fig. 1), and the signal displayed two broad humps (at $2\theta \sim 41^{\circ}$ and 75°) with d-spacing of ~ 0.258 nm and ~ 0.149 nm indicative of Two-line 307 ferrihydrite. No other Fe(III) oxides were detected. We observed a broadening of the XRD 308 309 peaks as the Sb content of the Fh increased; this was due to either a decrease in crystallinity as a result the incorporation of Sb in the structure of the Fe oxyhydroxide as revealed by 310 311 Mitsunobu et al. (2010b) or a decrease in crystallite size as hypothesized in the case of arsenic 312 substitution (Waychunas et al., 1996).

313

314 3.2 Microbial bio-reduction of Sb-bearing ferrihydrite

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We quantitatively monitored the bio-reduction of Sb-bearing ferrihydrite series over time. Bio-reduction started without a lag phase and reached a plateau after 25 days (Fig. 2). No measurable Fe(III) reduction took place in the non-inoculated treatments (data not shown). The apparent rate of reduction was $16.40 \pm 0.07\%$ (Fe^{II}_{tot}/Fe_{tot}) day⁻¹ for Fh, while the rate was less than 50% slower for Fh_Sb (Table 1). Additionally, the extent of reduction reached ~ 35% for pure Fh, and ~ 30% for Fh_Sb(0.04) and Fh_Sb(0.06), whereas it plateaued at ~ 50% for Fh_Sb(0.1) (Table 1). After 25 days of incubation, only a minor fraction - less than 0.5% - of the Sb ions were present in the aqueous phase irrespective of the molar ratio Sb/(Sb + Fe) of the initial Fhs.

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327 3.3 **Biogenic minerals**

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329 The mineralogical transformation of Fh and Fh_Sb series was monitored after the incubation330 period using a suite of spectroscopic techniques.

331

332 3.3.1 Biogenic minerals formed by Sb free Fh bio-reduction

The biogenic minerals displayed X-ray diffraction peaks with d-spacing characteristic of magnetite and siderite (Fig. 3). Based on the FWHM of the strongest peaks at d = 0.251 (hkl: 311) and 0.294 nm (hkl: 220), the estimated crystalline size of the magnetite was 9.4 nm which is in the same range as the sizes observed in bio-reduction experiments described in the literature (Kukkadapu et al., 2005; Byrne et al., 2011).

338 We performed Mössbauer temperature profiles at room temperature (RT), 143 K, 78 K 339 and 15 K (Fig. 4) to identify and quantify the biogenic minerals precipitated during Fh bio-340 reduction. At RT and 143 K, Mössbauer spectra are complex, with the superposition of a 341 paramagnetic doublet as well as the significant broadening of a magnetic component. At RT 342 the quadrupole doublet is characterized by large center shifts (CS) and a quadrupole splitting (Δ) values of 1.14 mm s⁻¹ and 1.92 mm s⁻¹ respectively, which are similar to those published 343 for siderite at room temperature, i.e. $CS = 1.14 \text{ mm s}^{-1}$ and $\Delta = 1.98 \text{ mm s}^{-1}$ (Génin et al., 344 345 2002). The magnetic component was fitted using two distributions of the hyperfine magnetic 346 field with widened lines, possibly due to magnetite observed using XRD. We fitted this 347 component with two broad overlapping sextets with two weak mean magnetic fields at 40.8 T and 13.2 T respectively. However, the two sextets are usually clearly defined and fitted with a
Lorentzian model when dealing with large-particle magnetite. This discrepancy in the line
shape of the Mössbauer spectrum at RT is due to the small particle size (less than 20 nm),
which may induce a greater distribution in the magnetic behavior (Silva et al., 2013; Alves et
al., 2016).

353 During cooling, the spectral features of the magnetic component change considerably. At 77 354 K, the ferrous doublet of the siderite phase observed at RT remains and the CS and the Δ 355 increase as expected for the hyperfine parameters of the Fe(II), and adopt a high spin 356 configuration at the octahedral site. The magnetic contribution of the spectrum exhibits five 357 sextets that were deconvoluted using a Lorentzian model; they correspond to the five iron 358 sites observed for magnetite below the Verwey temperature (Srivastava et al., 1981; Vandenberghe et al., 2000). At 15 K, the spectrum becomes more complex in its central part. 359 360 This indicates that all the carbonate-containing Fe(II) sites are now magnetically ordered and 361 represented by an octet versus paramagnetic doublet at 77 K (Maitte et al., 2015). Modelling 362 of this spectrum component, using the software package Recoil (Lagarec and Rancourt, 363 1998), is intricate, owing to the mixing of states, particularly given the fact that, for Fe(II) 364 compounds below magnetic transition, the electric quadrupole and magnetic hyperfine 365 interactions are generally of the same order of magnitude. Thus, it is not likely that this 366 component can be analyzed in terms of six-line patterns (sextets) as is the case for the magnetism of ferric compounds (Duboscq et al., 2020). The octet comprises eight lines that 367 368 have to be computed fully by solving the eigenvalues of the full Hamiltonian. The octet is characterized by hyperfine parameters (Table 2), a CS of 1.76 mm s⁻¹, electric quadrupole 369 interaction ($e^2qQ/2$) of 2.72 mm s⁻¹, and a hyperfine field H of 18.7 T typical of Fe(II) in 370 371 siderite (Maitte et al., 2015). The relative abundance of siderite was ~ 9 %.

372 The second interesting feature in the spectrum at 15 K is the appearance of resonance 373 lines on the high energies which correspond, after spectral deconvolution, to the five sextets 374 S1, S2, S3, S4 and S5. Their hyperfine parameters, listed in Table 2, are consistent with those generally reported for magnetite of five-sextets as observed in the literature (Srivastava et al., 375 376 1981) at low temperature. Moreover, the particle size constraints generate the phenomenon of 377 superparamagnetism and influence the magnetic properties, which could explain the gradual 378 evolution with regards to temperature from a broad sextet component to sharp sextets as the 379 temperature decreases. This behavior is typical of superparamagnetic particles (Goya et al., 380 2003), and may be related to the widening of the diffraction lines assigned to magnetite by 381 XRD (Fig. 3). The relative abundance of magnetite was \sim 91 %.

The Raman spectra realized at different locations on the sample (analyzed spot size of a few μ m) exhibited a strong band at 680 cm⁻¹ and a weaker one in the region of 310 cm⁻¹, indicating the precipitation of magnetite (Fig. A1) (Shebanova and Lazor, 2003; Jubb and Allen, 2010). Additional bands at 280, 725 and 1084 cm⁻¹ were attributed to siderite (Fig. A1) (Hanesch, 2009; Maitte et al., 2015). The fact that different locations display a different mineral indicates that the system is heterogeneous, and that various mineral precipitation pathways are involved.

389

390 3.3.2 Biogenic minerals formed by Fh_Sb bio-reduction

Neither magnetite nor siderite was detected as a result of the transformation of Sbbearing minerals at the level discernible by XRD (Fig. 3). Instead, the d-spacing displayed the sharp diffraction peaks of green rust 1 (GR1) and goethite as biogenic minerals irrespective of the Sb/(Fe + Sb) molar ratio. Peaks of an unidentified mineral (indicated by *) were also present on the diffractograms independently of the Sb/(Sb + Fe) molar ratio of the initial Fhs. No evidence for the formation of other Sb-hosting minerals such as hydroxyferroromeite or Fe- and Sb-bearing minerals such as tripulyite was detected. Based on the experimental setup, the reduction medium composition and the electron source used in our study (i.e. methanoate), the precipitated GR1 is likely a carbonated GR ($GR1(CO_3^{2-})$) as described elsewhere (Ona-Nguema et al., 2004; Zegeye et al., 2007; O'Loughlin et al., 2007).

We used the FWHM of the strongest peaks at d = 0.755 nm (hkl : 003) and d = 0.378 nm (hkl 202 : 006) to estimate the average GR1 crystal size. Our goal was not to determine the exact 203 crystal size but to monitor any variation due to the loading of Sb(V). The estimated crystal 204 sizes of the GR1(CO₃²⁻) bio-precipitated from the Fh_Sb(0.04), Fh_Sb(0.06) and Fh_Sb(0.1) 205 bio-reductions were 77 nm, 40 nm and 66 nm for respectively.

At room temperature, the Mössbauer spectra Fh_Sb(0.04), Fh_Sb(0.06), and Fh_Sb(0.1) were computer fitted with 5, 3 and 5 quadrupole doublets respectively (Fig. A2). Their hyperfine parameters are typical of ferrous and ferric cations (Table A1), but their mineralogical identification remains difficult. Nevertheless, the presence of $GR1(CO_3^{2-})$ and superparamagnetic goethite phases cannot be ruled out on the basis of the XRD characterization.

412 Iron oxides are often sensitive to effects such as small particle size in the nanometer range, 413 which complicates their identification and characterization. These constraints require 414 Mössbauer spectra to be recorded at a lower temperature. At 15 K, all the Fh Sb(0.04), 415 Fh Sb(0.06) and Fh Sb(0.1) samples displayed magnetically ordered components on which 416 paramagnetic doublets were superimposed (Fig. 5). The magnetic components exhibit broad 417 and asymmetric lines with a distribution of the hyperfine magnetic fields centered on three components at 45 T, 47.5 T and 49.4 T. The quadrupole shift obtained (- 0.22 mm s⁻¹) is 418 419 characteristic of goethite, also identified by XRD but with widened lines. The broadening of 420 magnetic lines is the result of relaxation effects of the smallest particles and/or a poor

421 crystallinity. This could explain the spectral behavior observed at RT, with a significant
422 doublet suspected as being superparamagnetic goethite (Kone et al., 2009).

The paramagnetic components were fitted with two quadrupole doublets (D1 and D2). 423 Doublet D1 with values of CS = 1.21 mm s⁻¹ and Δ = 2.9 mm s⁻¹ differs from doublet D2 424 with smaller values for CS (0. 38 mm s⁻¹) and Δ (0.74 mm s⁻¹), which are ascribed to 425 (GR1(CO₃²⁻) (Génin et al., 2006). It is also possible to detect another magnetic component 426 appearing as an octet characterized by hyperfine parameters (Table 2) with a CS of 1.52 mm 427 s^{-1} , electric quadrupole interaction ($e^2qQ/2$) of 2.47 mm s^{-1} and a hyperfine field H of 11 T, 428 which is typical of Fe(II) in probable association with Sb. This component unambiguously 429 distinguishes it from both siderite, which has a much higher magnetic field, and $GR1(CO_3^{2-})$, 430 431 which is a mixed-valence (ferrous-ferric) compound leading to two magnetic ordering temperatures, 5.2 and 7 K (Rusch et al., 2008). The relative abundance of $GR1(CO_3^{2-})$ and of 432 433 unspecified Fe(II) species increases with the increase of Sb(V) loading (Table 2).

434 We performed field cooled (FC) and zero-field cooled (ZFC) moment measurement 435 (Fig. A3). FC and ZFC curves overlap from 300 K down to approximately 15 K. At 436 temperatures lower than 15 K, there are two maxima in the ZFC curve. In a system including 437 particles, such maxima are usually associated with the blocking process of particles, thus determining the onset of superparamagnetic behavior (Néel, 1949; Brown, 1963). The 438 439 blocking temperatures estimated for the maximum magnetization of ZFC curves were 10.5 K 440 and 4.9 K. We expect that these two blocking temperatures are relative to the two main 441 minerals detected by Mössbauer spectroscopy and XRD, namely goethite and green rust.

The characterization by Raman spectroscopy of different locations on the biogenic minerals precipitated from the Fh_Sb series samples revealed different biogenic minerals (Fig. 6A). Thus, the results exhibited strong bands at 435 and 510 cm⁻¹ and a weak band at 220 cm⁻¹, characteristics of $GR1(CO_3^{2-})$ (Legrand et al., 2001b; Refait et al., 2011). We also 446 observed bands related to the precipitation of magnetite and siderite that were not detected by447 XRD, implying that these phases were present as minor species.

448 XPS analysis was carried out to investigate the chemical composition of the surface of biogenic solid, with special emphasis on determining the chemical speciation of Fe, Sb and C. 449 450 The Fe 2p, Sb 3d and C 1s spectra of unreacted and bio-reacted Fh Sb(0.1) samples are 451 compared in Fig. 7. The Fe 2p high-resolution spectrum of unreacted ferrihydrite Fh Sb(0.1) displays a maximum at \sim 711.0 eV and a broad satellite peak at \sim 719 eV, as is commonly 452 453 observed for Fe(III) species in iron (oxyhydr)oxides (Mallet et al., 2013; Fan et al., 2016). 454 Four multiplet peaks at 710.2, 711.3, 712.4 and 713.5 (FHWM = 1.3 eV) and peak area ratios 455 relative to the first Fe(III)-O component of 1.1, 0.69 and 0.38 respectively were used to fit the 456 2p_{3/2} spectrum in agreement with Gupta and Sen, 1974; Gupta and Sen, 1975 predictions (Fig. 457 7A, top). An additional surface peak at 714.7 eV is required to fit the data (Grosvenor et al., 458 2004).

A slight shift of the Fe $2p_{3/2}$ maximum component towards lower binding energy is observed, 459 460 and the satellite peak characteristic of Fe(III) species is strongly attenuated (Fig. 7A, bottom). 461 In addition to the four Fe(III) multiplet peaks, three multiplet peaks located at 708.6, 709.6 462 and 710.7 eV that characterize the presence of Fe(II) are required to fit the spectra of the Fh Sb(0.1) bio-reduced sample (Fig. 7A, bottom). The peak area ratio of the multiplet 463 464 components relative to the first Fe(II)-O component is 1.1 and 0.5, and the FHWM is 1.6 eV. 465 The presence of Fe(II) in the bio-reduced sample agrees well with the formation of GR1 (and 466 minor amounts of magnetite and siderite) as observed in the XRD and TMS characterizations. 467 The binding energy of Sb(III) and Sb(V) compounds bonded to oxygen species are too close 468 to allow individual quantification (Fig. 7B). Previous studies have reported that the binding 469 energy of Sb 3d_{3/2} core for Sb(III) and Sb(V) within antimonic oxides were usually 538.8-470 539.4 and 539.8-541.7 eV respectively (Huang and Ruiz, 2006; Mittal et al., 2011). The binding energy of Sb $3d_{5/2}$ peak position of unreacted Fh_Sb(0.1) at 540.0 eV agrees well with that reported for Sb(V)-O species (Fig 7B, top). Interestingly, the Sb $3d_{3/2}$ peak of the bio-reduced sample shifted to a slightly lower binding energy of 539.6 eV which may cautiously suggest that some reduction of Sb(V) into Sb(III) occurred at the surface of the sample (Fig 7B, bottom). In particular, Xu et al. (2011) reported Sb $3d_{3/2}$ binding energy of 539.7 eV for Sb(III) sorption onto goethite.

477 Lastly, the C 1s peak of unreacted Fh_Sb(0.1) is usually fitted with the components at 284.6 478 eV, 286.1 eV and 288.7 attributed to C-C, H, C-O and C=O, -O environments (Fig. 7C, top). 479 The presence of carbon reflects some surface contamination often observed in XPS 480 experiments. Bio-reduced samples show a significant increase in the carbonate components 481 located at 288.0 eV (Fig. 7C, bottom), in agreement with the formation of $GR1(CO_3^{2-})$ as 482 observed by XRD and TMS.

- 483
- 484 3.3.3 Stability of abiotic $GR1(CO_3^{2-})$
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 $GR1(CO_3^{2-})$ was synthesized as described by Bocher et al. (2004) and aged for one week 486 before being analyzed by Raman spectroscopy (Fig. 6B). The GR1(CO3²⁻) synthesized 487 without any additive (i.e. PO_4^{3-} or Sb(V) had transformed by the end of the aging time (Fig. 488 6B a). The Raman spectrum displayed a strong band at 670 cm⁻¹, which is assigned to 489 490 magnetite. Similarly, the study carried out by Bocher et al. (2004) demonstrated the 491 transformation of $GR1(CO_3^{2-})$ toward magnetite (Fe₃O₄) and ferrous hydroxide (Fe(OH)₂). Moreover, Guilbaud et al. (2013) used the same procedure to synthesize $GR1(CO_3^{2-})$ and 492 observed its transformation towards magnetite after 69 hours of aging. The Raman spectrum 493 of the $GR1(CO_3^{2-})$ synthesized in the presence of phosphate with an PO_4^{3-} / Fet ratio of 0.016, 494 exhibited two strong bands at 435 and 510 cm⁻¹ and a weak band at 220 cm⁻¹, indicating the 495

occurrence of $GR1(CO_3^{2-})$ (Fig. 6B b). No specific band of either magnetite or of siderite was 496 497 detected. This result agrees with the data produced by Benali et al. (2001) and Bocher et al. (2004), where the presence of phosphate stabilized the mixed-valence mineral and hindered 498 mineralogic transformation. When the $GR1(CO_3^{2-})$ precipitation occurred in the presence of 499 Sb(V), the Raman spectrum exhibited the two strong bands at 435, and 510 cm⁻¹ 500 characteristics of $GR1(CO_3^{2-})$ plus an additional unassigned phase, probably an iron-bearing 501 502 solid doped with Sb (Fig. 6B d, e). However, when the Sb(V) was added after the precipitation (~ 10 min) the characteristic Raman bands of $GR1(CO_3^{2-})$ and magnetite were 503 observed (Fig. 6B c). 504

507 **4. DISCUSSION**

508

509 4.1 Rate and extent of bio-reduction and Sb partitioning

510

511 The bio-reduction of Fh was not inhibited by Sb(V), which is in line with recent work 512 by Burton et al. (2019). It was expected that Sb(V) loading, like other elements substitutions, 513 would alter the physicochemical properties of the Fh (Tsapin et al., 1996; Fredrickson et al., 514 2001; Jentzsch and Penn, 2006; Ekstrom et al., 2010; Frierdich and Catalano, 2012b; Massey 515 et al., 2014), and consequently affects the rate and the extent of bio-reduction. In this study, 516 the rate of reduction decreased by ~ 50% in the presence of Sb(V) independently of the 517 Sb/(Sb+Fe) molar ratio (Table 1, Fig. 2). The observed decrease in the rate of reduction may 518 be related to the decline of a significant number of reactive sites on the Fh surface due to 519 Sb(V) loading (Fredrickson et al., 2001; Ekstrom et al., 2010; Voegelin et al., 2019).

520 Fh is a metastable mineral that transforms promptly to more thermodynamically stable 521 phases such as magnetite and goethite when reacted with Fe(II) (Hansel et al., 2005; Pedersen et al., 2005) under anoxic and non-sulfidic conditions. Electron transfer between surface-522 523 associated Fe(II) and structural Fe(III) has been shown to initiate the dissolution of Fh and the 524 precipitation of a thermodynamically stable secondary minerals (Williams and Scherer, 2004; 525 Yang et al., 2010; Boland et al., 2014). The bio-generation and sorption on the Fh surface of 526 the minimum amount of Fe(II) necessary to trigger Fh transformation would hamper the bio-527 reduction process. In the present study, the assays run with an Sb/(Sb+Fe) molar ratio inferior 528 to 0.06 displayed the same extent of reduction as that of pure Fh (Table 1, Fig. 2). The transformation of the host mineral was induced by ~ 35% of bio-generated Fe(II), which 529 530 subsequently halted the bio-reduction process. This result indicates that Sb(V) loading with an

531 Sb/(Sb+Fe) molar ratio of up to 0.06 has a minimal effect on Fh_Sb lability or reducibility
532 compared with pure Fh.

533 The extent of reduction of Fh Sb(0.1) reached ~ 50%, indicating that a higher Fe(II) loading was required to induce Fh Sb(0.1) transformation (Table 1, Fig. 2). It may be considered that 534 535 this Sb loading probably reduced the lability or reducibility of the Fh Sb(0.1) by likely 536 decreasing the transfer of electrons between the bio-generated Fe(II) and the Fh Sb(0.1). A 537 higher Fe(II) concentration was thus needed to trigger Fh Sb(0.1) transformation (Williams 538 and Scherer, 2004; Masue-Slowey et al., 2011; Frierdich et al., 2012). The inhibitory effect of 539 a high Sb(V) concentration (0.25 mmol of Sb(V) per gram of goethite) was also observed 540 during the Fe(II)-catalyzed recrystallization of goethite (Burton et al., 2020). The amount of 541 Sb(V) in the Fh Sb(0.1) probably increased the extent of reduction by likely retarding the 542 mineralogical transformation of the Fh Sb(0.1) and preserving the Fh Sb(0.1) surface area 543 for further reduction (Masue-Slowey et al., 2011).

544 At the end of the bio-reduction period, a minor fraction of Sb(V) was measured in the 545 solution. Our results indicate that a substantial amount of Sb(V) was probably associated with 546 secondary minerals, in line with the data observed elsewhere under biotic and abiotic 547 conditions (Mitsunobu et al., 2010b; Burton et al., 2019; Karimian et al., 2019a). We 548 hypothesize that a significant proportion of Sb(V) was likely adsorbed on the newly formed 549 Fe(II) minerals or otherwise occluded within aggregates of biogenic minerals. Based on 550 earlier reported data for the incorporation of Sb(V) into iron-bearing minerals (Mitsunobu et 551 al., 2013b; Karimian et al., 2017; Burton et al., 2019; Burton et al., 2020), we assume that 552 Sb(V) may structurally substitute for Fe(III) in the secondary minerals due to the similarity in ion size (Fe: 0.65 Å; Sb: 0.60 Å). The fact that Fe-O and Sb-O bound distances are in the 553 same range (Fe-O: 1.97-2 Å; Sb-O: 1.91 Å) in the Fe^{III}O₆ and Sb^VO₆ octahedra (Mitsunobu et 554 555 al., 2010b; Mitsunobu et al., 2013b) further supports this assumption. Mössbauer analyses

indicated the precipitation of an unidentified Fe(II)-bearing solid. Given our experimental conditions, the composition of the reduction medium, and hyperfine parameters, this nonidentified solid could plausibly be constituted of Sb, which could be an additional reason for the low amount of Sb present in the aqueous phase.

560 A slight reduction of the solid associated Sb(V) to Sb(III) was observed by the XPS analyses (Fig. 7B); this may have been triggered by electron transfer from Fe(II). Karimian et 561 562 al. (2019b) noted that humic acids (HA) and an Fe(II)-rich environment induced a substantial 563 reduction of Sb(V) by the potential formation of a ternary complex composed of Fe(II), HA 564 and Sb(V), thereby facilitating the electron transfer. Fe(II)-(III) minerals such as magnetite are also able to reduce Sb(V) to Sb(III), as reported by Kirsch et al. (2008). The spiking of 565 Sb(V) during the precipitation of $GR2(SO_4^{2-})$ triggered its adsorption on the newly formed 566 $GR2(SO_4^{2-})$ and its reduction to Sb(III). However, no reduction of Sb(V) was observed when 567 the latter was supplemented in the medium after the precipitation of $GR2(SO_4^{2-})$ (Mitsunobu 568 569 et al., 2008b; Mitsunobu et al., 2009). Although there is no data in the literature describing the interaction of $GR1(CO_3^{2-})$ and Sb(V), the slight reduction observed in our study was probably 570 induced by the Fe(II) / Fe(III) complex during the precipitation of $GR1(CO_3^{2-})$. 571

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573 4.2 Precipitation of biogenic minerals

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575 No detectable biogenic mineral was observed in the control assays (without bacteria), 576 whatever the Sb loading. This observation is in line with previous studies showing that the 577 transformation of Fh to a thermodynamically more stable mineral necessitates a long period 578 (months or years) in our experimental conditions (i.e. near-neutral pH and room temperature) 579 (Schwertmann et al., 2004).

582 4.2.1 Biogenic minerals formed by bio-reduction of Sb-free Fh

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584 The ratio of electron donor to electron acceptor is a significant parameter, as it impacts 585 the Fe(II) to Fe(III) fraction that controls the nature of the biogenic minerals in the reduction 586 media. Indeed, a lower Fe(II) / Fe(III) ratio (< 1.0 mmol Fe(II)/g Fh) promotes the 587 precipitation of ferric phases, such as lepidocrocite and goethite. At the same time, a higher 588 ratio leads to the precipitation of ferrous and ferrous/ferric phases such as siderite, magnetite 589 and GR (Hansel et al., 2003; Yang et al., 2010). In this study, we conducted the Fh bio-590 reduction experiments with an excess of electron source, and the resulting biogenic minerals 591 were magnetite (91%) and siderite (9%) (Fig. 8). Iron oxides particles have a high affinity to 592 bacteria cells walls and associated extracellular polymeric substances (Fein et al., 1997; 593 Lower et al., 2001; Luef et al., 2013), leading to the formation of aggregated structures. 594 Furthermore, the Fh drying procedure used in this study entails the aggregation of Fh particles 595 (Gilbert et al., 2009). Such bio-inorganic structures are likely to lead to the formation of 596 micro-environments establishing distinct chemical and redox zones relative to the bulk 597 solution (Zegeve et al., 2010; Luef et al., 2013). Similar conditions may induce the 598 precipitation of various biogenic minerals due to the occurrence of a differential Fe(II) 599 gradient. In the literature, the precipitation of siderite is commonly observed in the presence 600 of an HCO_3^- buffered medium (Fredrickson et al., 1998) or in an atmosphere containing CO_2 601 (Roh et al., 2003), which was not the case in our experimental setup. As siderite precipitation 602 requires high local bicarbonate concentration, we hypothesized that its formation occurs in the 603 micro-environments of aggregates where the mass transfer is probably limited or slowed 604 down, thereby allowing the simultaneous accumulation of both Fe(II) and bicarbonate (Pallud et al., 2010). 605

On the other hand, magnetite precipitation is thought to involve a topotactic transformation process (Tronc et al., 1992; Pedersen et al., 2005; Usman et al., 2018). The sorption of Fe(II) generated by bacterial activity induces interfacial electron transfer to the structural Fe(III) and catalyzes the subsequent precipitation of magnetite as the thermodynamically more stable phase.

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4.2.2 Biogenic minerals formed by Fh_Sb bio-reduction

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The bio-reduction of Sb-bearing Fh led mainly to the precipitation of $GR1(CO_3^{2-})$. 614 615 goethite and an unidentified Fe(II)-bearing solid irrespective of the Sb/Fe ratio. However, the 616 relative abundance of this unidentified Fe(II) bearing solid rises with the increase of Sb loading (Fig. 8). Its relative abundance was $\sim 7\%$ and reached $\sim 17\%$ for the assays run with 617 618 Fh Sb(0.04) and Fh Sb(0.1) respectively (Fig. 8). While which secondary minerals form 619 depends on the partitioning of the Fe(II) between aqueous and solid associated forms, the 620 presence of Sb also seems to orient the nature of the biogenic minerals. The precipitation of 621 magnetite as a very minor phase (Fig. 6A) may be a consequence of the presence of Sb(V). 622 Fe(II) was generated almost to its maximum concentration just after four days of incubation in 623 the absence of Sb(V). The generation of a high Fe(II) concentration in a very short time 624 shifted the precipitation toward magnetite, as witnessed elsewhere in the literature (Ona-Nguema et al., 2002). The presence of Sb(V) reduced the rate of reduction by 50%, resulting 625 626 in slower production of Fe(II) over time and consequently hindering the formation of magnetite. 627

In a recent study focusing on the same topic (*i.e.* the bio-reduction of Sb bearing ferrihydrite), feroxyhyte (δ '-FeOOH) and goethite were the main biogenic minerals (Burton et al., 2019). The main difference between the two studies is that in Burton et al. (2019) the electron source was in deficit relative to ferrihydrite concentration. As a result, the Fe(II) concentration produced in their research was lower, thus providing the chemical condition for the precipitation of Fe(III)-bearing biogenic minerals. In the current study, however, the excess of electron source permitted a more substantial input of Fe(II), allowing the precipitation of both $GR1(CO_3^{2^-})$ and goethite. These observations are in agreement with Hansel et al. (2003), who showed a threshold for segregating the formation of magnetite and goethite based on the Fe(II) / Fe(III) ratio in a biotic and abiotic systems.

638 It is commonly accepted that the precipitation of ferrous hydroxides, Fe(III) oxides and 639 oxyhydroxides is the result of the condensation phenomena of iron cations initially present in aqueous solution. Jolivet et al. (2004) described the different steps involved in these 640 641 precipitation processes and revealed how the structure of polymeric precursors determines the 642 nucleation and growth of the solid phases in a Fe-H₂O system. These authors posited that 643 growth of goethite by the dissolution/crystallization process might be interpreted as taking 644 place on the basis of the planar tetramer $[M_4(OH)_{12} (OH_2)_4]_0$. In our study, the 645 superparamagnetic goethite, identified by Mössbauer analysis, appears to be in line with these 646 mechanisms based on the precipitation of nanosized solid phases. We assumed that the ZFC 647 maxima at 10.5 K (Fig A3) originates from goethite (Guyodo et al., 2003; Barrero et al., 2006), and considering an anisotropy constant of 103 J m^{-3} , we calculated the average 648 649 goethite volume, corresponding to a sphere of about 20 nm diameter. This calculated particle 650 size agrees with the intriguing line shapes of the Mössbauer spectrum obtained at RT. Indeed, 651 as was expected for a poorly crystallized goethite, we observed a total absence of any magnetic signal (sextet) which would suggest the presence of nano-crystalline goethite in a 652 653 superparamagnetic state (Janot et al., 1968; Olowe et al., 1990).

As $GR1(CO_3^{2-})$ double-layered structure bears Fe(II), Fe(III) and CO_3^{2-} , we hypothesized that its formation was governed by a dissolution/precipitation process. The intriguing feature is

that $GR1(CO_3^{2-})$ is a metastable mineral that transforms relatively quickly (> 24 h) into a 656 thermodynamically more stable mineral such as magnetite or goethite, depending on the 657 658 geochemical conditions (Benali et al., 2001; Guilbaud et al., 2013). However, the increased Sb(V) loading in our system tended to favor the precipitation of $GR1(CO_3^{2-})$ rather than 659 660 goethite (Fig. 8): thus, the relative abundance of goethite declined by 25% when the Sb/(Fe + Sb) molar ratios were raised from 0.06 to 0.1, even though goethite has a much lower 661 standard-state Gibbs energy (- 489.8 \pm 1.2 kJ/mol) (Hiemstra, 2015) than of GR1(CO₃²⁻) (-662 4044.24 kJ/mol) (Drissi et al., 1995). Given the higher thermodynamic stability of goethite 663 compared with $GR1(CO_3^{2-})$, it was reasonable to assume that the system would move to 664 goethite. In addition, $GR1(CO_3^{2-})$ needs to be in equilibrium with aqueous Fe(II) in order to 665 circumvent any mineralogical transformation. But this aqueous Fe(II) would be depleted by 666 sorption on the goethite surface, thereby inducing the dissolution of $GR1(CO_3^{2-})$. For 667 668 example, Ona-Nguema et al. (2002) demonstrated the depletion of aqueous Fe(II) which was in equilibrium with $GR1(CO_3^{2-})$ by sorption on the remaining lepidocrocite surface, thereby 669 inducing the dissolution of $GR1(CO_3^{2-})$ and the precipitation of magnetite in a biotic system. 670 Similarly, under anoxic conditions, $GR1(CO_3^{2-})$ transforms into magnetite, $Fe(OH)_2$ or 671 siderite (Benali et al., 2001), but in the presence of phosphate it is likely that the adsorption of 672 the latter species upon the lateral faces of abiotic GR1(CO₃²⁻) crystals likely hindered this 673 transformation, by acting as a physical barrier (Bocher et al., 2004). Mitsunobu et al. (2009) 674 also observed the stabilization of abiotic $GR2(SO_4^{2-})$ when precipitated in the presence of 675 Sb(V). The adsorption of Sb(V) on the edge sites of $GR2(SO_4^{2-})$ with an inner sphere complex 676 was the main parameter hampering its transformation toward magnetite. To the authors' 677 knowledge, no comparable information regarding Sb(V) and GR1(CO32-) is available. We 678 therefore hypothesized that the adsorption of Sb(V) species on the $GR1(CO_3^{2-})$ crystals 679 680 hampered its dissolution by acting as a physical barrier. The observation, after the ageing period, of magnetite and $GR1(CO_3^{2^-})$ when Sb(V) was added after $GR1(CO_3^{2^-})$ precipitation implies that the stabilizing effect of Sb(V) is probably more marked when Sb(V) and iron are co-precipitated together. Stabilization efficiency seems to be related to the interaction of Sb(V) and iron during the $GR1(CO_3^{2^-})$ crystallite formation rather than to Sb(V) adsorption after the crystal growth (Fig 6B).

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687 *4.3 Geochemical significance*

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689 While this study addresses the effect of Sb (V) on the bio-reduction of Fh and the 690 resulting secondary minerals, our results may also be of interest in improving our 691 understanding of the mobility of Sb in the presence of iron-reducing bacteria. In natural soils, 692 antimony has been shown to be closely associated with iron oxides. Adsorption and 693 precipitation reactions with Fe(III) oxides play a crucial role in controlling the mobility and 694 the fate of Sb(V) in natural settings.

695 In subsurface environments, seasonal water table oscillation caused by rainfall and irrigation 696 establishes oxic-anoxic cycles. Under reducing conditions, the iron oxides may undergo 697 dissolution and mineralogy transformation upon reaction with iron-reducing bacteria that can potentially affect the fate and the mobility of sorbed Sb(V). During Sb(V)-bearing bio-698 699 reduction of Fh. goethite and feroxyhyte formation is observed in carbon-deficient setups (Burton et al., 2019). In contrast, goethite and $GR1(CO_3^{2-})$ is precipitated in carbon-rich 700 701 systems (this study). In both cases, Sb(V) partitioned mainly to secondary minerals, 702 suggesting that labile organic carbon's availability does not affect Sb(V) 's sorption into the 703 structure of secondary minerals. Furthermore, even a partial formation of Fe(II) can trigger 704 the mineralogical transformation of iron oxide and lead to Sb(V) sorption in a crystallized 705 secondary mineral. At redox transitions zones, iron oxides undergo numerous redox cycling that will likely lead to the precipitation of various secondary minerals displaying different affinity to Sb(V) depending on their properties. Ultimately, Sb(V) repartitioning and speciation will vary among environments but should be controlled by the flux of Fe(II) which dictates the secondary minerals' nature.

710 Our study also demonstrates that Sb(V)-loading affects the mineralogy of secondary minerals 711 indicating that Sb(V) concentration controls the Fe(II) flux. The increase of Sb(V) led to an substantial Fe(II) formation (Fe(II)/Fe_{tot} ~ 0.5) leading to the simultaneous rise of metastable 712 $GR1(CO_3^{2-})$ and a decrease of goethite. The Sb(V) sorption on $GR1(CO_3^{2-})$ seems to drive the 713 mineral's stability. In contaminated subsurface environments, $GR1(CO_3^{2-})$ displays the ability 714 715 to remove arsenic (As) species and to be a host phase for As (Perez et al., 2021). It is, therefore, necessary to study the fate of Sb(V) sorbed on $GR1(CO_3^{2-})$ particles to decipher the 716 717 role played by these metastable Fe(II)/Fe(III) minerals in natural settings in regards to the 718 (im)mobilization of metalloids. Our study's significance is that the parameters governing the 719 biogeochemical flux of Fe(II) should be understood before considering iron oxides and 720 secondary minerals as a reliable Sb(V) sink for long-term storage in Sb-contaminated 721 environments.

722

725 **5. CONCLUSIONS**

726

This study highlights how Sb affects the rate and the extent of Fh bio-reduction as well as 727 the nature of biogenic minerals. By using complementary characterization methods, we were 728 729 able to capture the precipitation of a relative abundance of various biogenic minerals with a 730 relative abundance that were governed by multiple biomineralization pathways. The presence 731 of Sb reduced by almost 50% the rate of bio-reduction and enhanced the extent of reduction 732 of Fh with a Sb/(Sb + Fe) molar ratio of 0.1. Note should also be taken of the total Fh 733 transformation irrespective of the Sb(V) concentration. Thus, magnetite precipitated in the 734 absence of Sb(V), while GR1(CO₃²⁻), goethite and an unidentified Fe(II)-Sb-bearing solid 735 formed in its presence. Importantly, this study shows that the efficiency of Sb in stabilizing the metastable $GR1(CO_3^{2-})$ by mitigating its mineralogical transformation toward a 736 737 thermodynamically more stable mineral such as goethite. The question raised by this study is, how does Sb(V) interact with $GR1(CO_3^{2-})$ crystals? A natural next step would be to carry out 738 research to reveal the interaction of Sb(V) with $GR1(CO_3^{2-})$ at the crystallographic level. 739 740

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753 **6. REFERENCES**

- Abdelmoula M., Trolard F., Bourrié G. and Génin J. M. R. (1998) Evidence for the Fe(II)Fe(III) green rust "Fougerite" mineral occurrence in a hydromorphic soil and its
 transformation with depth. *Hyperfine Interact.* 112, 235–238.
- Ai J., Yin W. and Hansen H. C. (2019) Fast Dechlorination of Chlorinated Ethylenes by
 Green Rust in the Presence of Bone Char. *Environ. Sci. Technol. Lett.* 6, 191–196.
- Alves A. F., Mendo S. G., Ferreira L. P., Mendonça M. H., Ferreira P., Godinho M., Cruz M.
- M. and Carvalho M. D. (2016) Gelatine-assisted synthesis of magnetite nanoparticles for
 magnetic hyperthermia. *J. Nanoparticle Res.* 18, 27.
- 762 Amstaetter K., Borch T. and Kappler A. (2012) Influence of humic acid imposed changes of
- ferrihydrite aggregation on microbial Fe(III) reduction. *Geochim. Cosmochim. Acta* 85, 326–
 341.
- An Y. J. and Kim M. (2009) Effect of antimony on the microbial growth and the activities of
 soil enzymes. *Chemosphere* 74, 654–659.
- Anderson C. G. (2012) The metallurgy of antimony. *Chemie der Erde* **72**, 3–8.
- Barrero C. A., Betancur J. D., Greneche J. M., Goya G. F. and Berquó T. S. (2006)
 Magnetism in non-stoichiometric goethite of varying total water content and surface
 area. *Geophys. J. Int.* 164, 331–339.
- 771 Benali O., Abdelmoula M., Refait P. and Génin J. M. R. (2001) Effect of orthophosphate on
- the oxidation products of Fe(II)-Fe(III) hydroxycarbonate: The transformation of green
 rust to ferrihydrite. *Geochim. Cosmochim. Acta* 65, 1715–1726.
- Bocher F., Géhin A., Ruby C., Ghanbaja J., Abdelmoula M. and Génin J. M. R. (2004)
 Coprecipitation of Fe(II-III) hydroxycarbonate green rust stabilised by phosphate
 adsorption. *Solid State Sci.* 6, 117–124.
- 777 Boland D. D., Collins R. N., Miller C. J., Glover C. J. and Waite T. D. (2014) Effect of

- solution and solid-phase conditions on the Fe(II)-accelerated transformation of
 ferrihydrite to lepidocrocite and goethite. *Environ. Sci. Technol.* 48, 5477–5485.
- 780 Borch T. and Fendorf S. (2007) Phosphate Interactions with Iron (Hydr)oxides:
- 781 Mineralization Pathways and Phosphorus Retention upon Bioreduction. In *Adsorption of*
- 782 Metals by Geomedia II: Variables, Mechanisms, and Model Applications (eds. M. O.
- 783 Barnett and D. B. Kent). pp. 321–348.
- Bourrié G., Trolard F., Jaffrezic J. M. R. G., Maître V. and Abdelmoula M. (1999) Iron
 control by equilibria between hydroxy-Green Rusts and solutions in hydromorphic soils. *Geochim. Cosmochim. Acta* 63, 3417–3427.
- Brown W. F. (1963) Thermal Fluctuations of a Single-Domain Particle. *Phys. Rev.* 130,
 1677–1686.
- Burton E. D., Hockmann K. and Karimian N. (2020) Antimony Sorption to Goethite: Effects
 of Fe(II)-Catalyzed Recrystallization. *ACS Earth Sp. Chem.* 4, 476–487.
- Burton E. D., Hockmann K., Karimian N. and Johnston S. G. (2019) Antimony mobility in
 reducing environments: The effect of microbial iron(III)-reduction and associated
 secondary mineralization. *Geochim. Cosmochim. Acta* 245, 278–289.
- 794 Byrne J. M., Telling N. D., Coker V. S., Pattrick R. A. D., Van Der Laan G., Arenholz E.,
- Tuna F. and Lloyd J. R. (2011) Control of nanoparticle size, reactivity and magnetic
 properties during the bioproduction of magnetite by Geobacter sulfurreducens. *Nanotechnology* 22, 1–22.
- Christiansen B. C., Balic-Zunic T., Dideriksen K. and Stipp S. L. S. (2009) Identification of
 Green Rust in Groundwater. *Environ. Sci. Technol.* 43, 3436–3441.
- Coates J. D., Phillips E. J. P., Lonergan D. J., Jenter H. and Lovley D. R. (1996) Isolation of
 Geobacter species from diverse sedimentary environments. *Appl. Environ. Microbiol.* 62,
- 802 1531–1536.

- 803 Cornell R. M. and Schwertmann U. (2003) *The iron oxides: structure, properties, reactions,*804 *occurrences and uses: second ed.*, Wiley-VCH:Wheinhein, Germany, 2003.
- Craw D., Wilson N. and Ashley P. M. (2004) Geochemical controls on the environmental
 mobility of Sb and As at mesothermal antimony and gold deposits. *Appl. Earth Sci.* 113,
 3–10.
- Dixit S. and Hering J. G. (2003) Comparison of Arsenic(V) and Arsenic(III) Sorption onto
 Iron Oxide Minerals: Implications for Arsenic Mobility. *Environ. Sci. Technol.* 37,
 4182–4189.
- B11 Drissi S. H., Refait P., Abdelmoula M. and Génin J. M. R. (1995) The preparation and
 thermodynamic properties of Fe(II)Fe(III) hydroxide-carbonate (green rust 1); Pourbaix
 diagram of iron in carbonate-containing aqueous media. *Corros. Sci.* 37, 2025–2041.
- B14 Duboscq J., Abdelmoula M., Rémazeilles C., Jeannin M., Sabot R. and Refait P. (2020) On
 B15 the formation and transformation of Fe(III)-containing chukanovite, FeII2B16 xFeIIIx(OH)2-xOxCO3. J. Phys. Chem. Solids 138, 109310.
- Ekstrom E. B., Learman D. R., Madden A. S. and Hansel C. M. (2010) Contrasting e ff ects of
 Al substitution on microbial reduction of Fe (III) (hydr) oxides. *Geochim. Cosmochim. Acta* 74, 7086–7099.
- Erbs M., Bruun Hansen H. C. and Olsen C. E. (1999) Reductive Dechlorination of Carbon
 Tetrachloride Using Iron(II) Iron(III) Hydroxide Sulfate (Green Rust). *Environ. Sci. Technol.* 33, 307–311.
- 823 Eusterhues K., Hädrich A., Neidhardt J., Küsel K., Keller T. F., Jandt K. D. and Totsche K. U.
- 824 (2014) Reduction of ferrihydrite with adsorbed and coprecipitated organic matter:
 825 Microbial reduction by Geobacter bremensis vs. abiotic reduction by Na-dithionite.
 826 *Biogeosciences* 11, 4953–4966.
- 827 Fan J.-X., Wang Y.-J., Fan T.-T., Dang F. and Zhou D.-M. (2016) Effect of aqueous Fe(II) on

- 828 Sb(V) sorption on soil and goethite. *Chemosphere* **147**, 44–51.
- Feder F., Trolard F., Klingelhöfer G. and Bourrié G. (2005) In situ Mössbauer spectroscopy:
 Evidence for green rust (fougerite) in a gleysol and its mineralogical transformations
 with time and depth. *Geochim. Cosmochim. Acta* 69, 4463–4483.
- Fein J. B., Daughney C. J., Yee N. and Davis T. A. (1997) A chemical equilibrium model for
 metal adsorption onto bacterial surfaces. *Geochim. Cosmochim. Acta* 61, 3319–3328.
- Filella M., Belzile N. and Chen Y. W. (2002) Antimony in the environment: A review
 focused on natural waters I. Occurrence. *Earth-Science Rev.* 57, 125–176.
- 836 Fredrickson J. K., Zachara J. M., Kennedy D. W., Dong H., Onstott T. C., Hinman N. W. and
- Li S. M. (1998) Biogenic iron mineralization accompanying the dissimilatory reduction
- 838 of hydrous ferric oxide by a groundwater bacterium. *Geochim. Cosmochim. Acta* 62,
 839 3239–3257.
- 840 Fredrickson J. K., Zachara J. M., Kukkadapu R. K., Gorby Y. A., Smith S. C. and Brown C.
- F. (2001) Biotransformation of Ni-substituted hydrous ferric oxide by an Fe(III)reducing bacterium. *Environ. Sci. Technol.* 35, 703–712.
- Frierdich A. J., and Catalano J. G. (2012a) Controls on Fe(II)-activated trace element release
 from goethite and hematite. *Environ. Sci. Technol.* 46, 1519–1526.
- Frierdich A. J., and Catalano J. G. (2012b) Fe(II)-mediated reduction and repartitioning of
 structurally incorporated Cu, Co, and Mn in iron oxides. *Environ. Sci. Technol.* 46,
 11070–11077.
- Frierdich A. J., Scherer M. M., Bachman J. E., Engelhard M. H., Rapponotti B. W. and
 Catalano J. G. (2012) Inhibition of trace element release during Fe(II)-activated
 recrystallization of Al-, Cr-, and Sn-substituted goethite and hematite. *Environ. Sci.*
- 851 *Technol.* **46**, 10031–10039.
- 852 Gebel T. (1997) Arsenic and antimony: comparative approach on mechanistic toxicology.

- 853 *Chem. Biol. Interact.* **107**, 131–144.
- Génin J.-M. R., Abdelmoula M., Ruby C. and Upadhyay C. (2006) Speciation of iron;
 characterisation and structure of green rusts and FeII–III oxyhydroxycarbonate fougerite. *Comptes Rendus Geosci.* 338, 402–419.
- Génin J.-M. R., Dhouibi L., Refait P., Abdelmoula M. and Triki E. (2002) Influence of
 Phosphate on Corrosion Products of Iron in Chloride-Polluted-Concrete-Simulating
 Solutions: Ferrihydrite vs Green Rust. *Corrosion* 58, 467–478.
- Génin J.-M. R., Olowe A. A., Refait P. and Simon L. (1996) On the stoichiometry and
 pourbaix diagram of Fe(II)-Fe(III) hydroxy-sulphate or sulphate-containing green rust 2:
- An electrochemical and Mössbauer spectroscopy study. *Corros. Sci.* **38**, 1751–1762.
- Gilbert B., Ono R. K., Ching K. A. and Kim C. S. (2009) The effects of nanoparticle
 aggregation processes on aggregate structure and metal uptake. *J. Colloid Interface Sci.*339, 285–295.
- Goya G. F., Berquó T. S., Fonseca F. C. and Morales M. P. (2003) Static and dynamic
 magnetic properties of spherical magnetite nanoparticles. *J. Appl. Phys.* 94, 3520–3528.
- 868 Grosvenor A. P., Kobe B. A., Biesinger M. C. and McIntyre N. S. (2004) Investigation of
- 869 multiplet splitting of Fe 2p XPS spectra and bonding in iron compounds. *Surf. Interface*870 *Anal.* 36, 1564–1574.
- Guilbaud R., White M. L. and Poulton S. W. (2013) Surface charge and growth of sulphate
 and carbonate green rust in aqueous media. *Geochim. Cosmochim. Acta* 108, 141–153.
- Gupta R. P. and Sen S. K. (1975) Calculation of multiplet structure of core p-vacancy levels.
 II. *Phys. Rev. B* 12, 15–19.
- Gupta R. P. and Sen S. K. (1974) Calculation of multiplet structure of core p-vacancy levels. *Phys. Rev. B* 10, 71–77.
- 877 Guyodo Y., Mostrom A., Lee Penn R. and Banerjee S. K. (2003) From Nanodots to

- Nanorods: Oriented aggregation and magnetic evolution of nanocrystalline goethite. *Geophys. Res. Lett.* **30**, 1512.
- Halevy I., Alesker M., Schuster E. M., Popovitz-Biro R. and Feldman Y. (2017) A key role
 for green rust in the Precambrian oceans and the genesis of iron formations. *Nat. Geosci.*10, 135–139.
- Hanesch M. (2009) Raman spectroscopy of iron oxides and (oxy)hydroxides at low laser
 power and possible applications in environmental magnetic studies. *Geophys. J. Int.* 177,
 941–948.
- Hansel C. M., Benner S. G. and Fendorf S. (2005) Competing Fe(II)-induced mineralization
 pathways of ferrihydrite. *Environ. Sci. Technol.* 39, 7147–7153.
- Hansel C. M., Benner S. G., Neiss J., Dohnalkova A., Kukkadapu R. K. and Fendorf S.
 (2003) Secondary mineralization pathways induced by dissimilatory iron reduction of
 ferrihydrite under advective flow. *Geochim. Cosmochim. Acta* 67, 2977–2992.
- Hansen H. C. B. (1989) Composition, stabilization, and light absorption of Fe(II)Fe(III)
 hydroxy-carbonate ('green rust'). *Clay Miner.* 24, 663–669.
- He M., Wang N., Long X., Zhang C., Ma C., Zhong Q., Wang A., Wang Y., Pervaiz A. and
 Shan J. (2019) Antimony speciation in the environment: Recent advances in
 understanding the biogeochemical processes and ecological effects. *J. Environ. Sci. (China)* 75, 14–39.
- Hiemstra T. (2015) Formation, stability, and solubility of metal oxide nanoparticles: Surface
 entropy, enthalpy, and free energy of ferrihydrite. *Geochim. Cosmochim. Acta* 158, 179–
 198.
- Hockmann K., Lenz M., Tandy S., Nachtegaal M., Janousch M. and Schulin R. (2014)
 Release of antimony from contaminated soil induced by redox changes. *J. Hazard. Mater.* 275, 215–221.

38

- Huang Y. and Ruiz P. (2006) The nature of antimony-enriched surface layer of Fe–Sb mixed
 oxides. *Appl. Surf. Sci.* 252, 7849–7855.
- 905 Ilgen A. G. and Trainor T. P. (2012) Sb(III) and Sb(V) sorption onto Al-rich phases: Hydrous
- Al oxide and the clay minerals kaolinite KGa-1b and oxidized and reduced nontronite
 NAu-1. *Environ. Sci. Technol.* 46, 843–851.
- Janot C., Chabanel M. and Hertzog E. (1968) Etude d'une limonite par effet Mössabuer. *Bull. Soc. Fr. Miner. Cristal.* 91, 166–171.
- Jentzsch T. L. and Penn R. L. (2006) Influence of aluminum doping on ferrihydrite
 nanoparticle reactivity. J. Phys. Chem. B 110, 11746–11750.
- 912 Johnson C. A., Murayama M., Küsel K. and Hochella M. F. (2015) Polycrystallinity of green
- 913 rust minerals and their synthetic analogs: Implications for particle formation and
 914 reactivity in complex systems. *Am. Mineral.* 100, 2091–2105.
- Jolivet J. P., Chanéac C. and Tronc E. (2004) Iron oxide chemistry. From molecular clusters
 to extended solid networks. *Chem. Commun.* 4, 477–483.
- Jubb A. M. and Allen H. C. (2010) Vibrational spectroscopic characterization of hematite,
 maghemite, and magnetite thin films produced by vapor deposition. *ACS Appl. Mater.*
- 919 *Interfaces* **2**, 2804–2812.
- 920 Karimian N., Burton E. D. and Johnston S. G. (2019a) Antimony speciation and mobility
- 921 during Fe(II)-induced transformation of humic acid-antimony(V)-iron(III) coprecipitates.
 922 *Environ. Pollut.* 254, 113112.
- Karimian N., Burton E. D., Johnston S. G., Hockmann K. and Choppala G. (2019b) Humic
 acid impacts antimony partitioning and speciation during iron(II)-induced ferrihydrite
 transformation. *Sci. Total Environ.* 683, 399–410.
- 926 Karimian N., Johnston S. G. and Burton E. D. (2017) Antimony and Arsenic Behavior during
- 927 Fe(II)-Induced Transformation of Jarosite. *Environ. Sci. Technol.* **51**, 4259–4268.

Kirsch R., Scheinost A. C., Rossberg A., Banerjee D. and Charlet L. (2008) Reduction of
antimony by nano-particulate magnetite and mackinawite. *Mineral. Mag.* 72, 185–189.

930 Koeksoy E., Sundman A., Byrne J. M., Lohmayer R., Planer-Friedrich B., Halevy I.,

- Konhauser K. O. and Kappler A. (2019) Formation of green rust and elemental sulfur in
 an analogue for oxygenated ferro-euxinic transition zones of Precambrian oceans. *Geology* 47, 211–214.
- Kone T., Hanna K., Abdelmoula M., Ruby C. and Carteret C. (2009) Reductive
 transformation and mineralization of an azo dye by hydroxysulphate green rust
 preceding oxidation using H2O2 at neutral pH. *Chemosphere* **75**, 212–219.
- Wickadapu R. K., Zachara J. M., Fredrickson J. K., Kennedy D. W., Dohnalkova A. C. and
 McCready D. E. (2005) Ferrous hydroxy carbonate is a stable transformation product of
 biogenic magnetite. *Am. Mineral.* 90, 510–515.
- 940 Lagarec K. and Rancourt D. G. (1998) Recoil Mössbauer Spectral Analysis Software.
- Legrand L., Sagon G., Lecomte S., Chausse A. and Messina R. (2001b) A Raman and
 infrared study of a new carbonate green rust obtained by electrochemical way. *Corros. Sci.* 43, 1739–1749.
- Li J., Zheng B. H., He Y., Zhou Y., Chen X., Ruan S., Yang Y., Dai C. and Tang L. (2018)
 Antimony contamination, consequences and removal techniques: A review. *Ecotoxicol.*
- 946 *Environ. Saf.* **156**, 125–134.
- 947 Liu F., Le X. C., McKnight-Whitford A., Xia Y., Wu F., Elswick E., Johnson C. C. and Zhu
- 948 C. (2010) Antimony speciation and contamination of waters in the Xikuangshan 949 antimony mining and smelting area, China. *Environ. Geochem. Health* **32**, 401–413.
- 950 Lovley D. R. (1993) Dissimilatory Metal Reduction. Annu. Rev. Micobiology 47, 262–290.
- 951 Lovley D. R. (2000) Fe(III) and Mn(IV) Reduction. In Environmental Microbe-Metal
- 952 *Interactions* (ed. D. R. Lovley). ASM Press, Washington, DC. pp. 3–30.

- 953 Lovley D. R., Giovannoni S. J., White D. C., Champine J. E., Phillips E. J. P., Gorby Y. A.
- and Goodwin S. (1993) Geobacter metallireducens gen. nov., a microorganism
- capable of coupling the complete oxidation of organic compounds to the reduction of
 iron and other metals. *Arch. Microbiol.* 159, 336–344.
- Lower S. K., Hochella M. F. and Beveridge T. J. (2001) Bacterial Recognition of Mineral
 Surfaces: Nanoscale Interactions Between Shewanella and α-FeOOH. *Science* 292,
 1360–1363.
- Luef B., Fakra S. C., Csencsits R., Wrighton K. C., Williams K. H., Wilkins M. J., Downing
 K. H., Long P. E., Comolli L. R. and Banfield J. F. (2013) Iron-reducing bacteria
 accumulate ferric oxyhydroxide nanoparticle aggregates that may support planktonic
 growth. *ISME J.* 7, 338–350.
- Maier T. M. and Myers C. R. (2001) Isolation and Characterization of a Shewanella
 putrefaciens MR-1 Electron Transport Regulator etrA Mutant : Reassessment of the Role
 of EtrA. J. Bact. 183, 4918–4926.
- 967 Maitte B., Jorand F. P. A., Grgic D., Abdelmoula M. and Carteret C. (2015) Remineralization
- 968 of ferrous carbonate from bioreduction of natural goethite in the Lorraine iron ore
 969 (Minette) by Shewanella putrefaciens. *Chem. Geol.* 412, 48–58.
- Mallet M., Barthélémy K., Ruby C., Renard A. and Naille S. (2013) Investigation of
 phosphate adsorption onto ferrihydrite by X-ray Photoelectron Spectroscopy. *J. Colloid Interface Sci.* 407, 95–101.
- Massey M. S., Lezama-Pacheco J. S., Michel F. M. and Fendorf S. (2014) Uranium
 incorporation into aluminum-substituted ferrihydrite during iron(ii)-induced
 transformation. *Environ. Sci. Process. Impacts* 16, 2137–2144.
- Masue-Slowey Y., Loeppert R. H. and Fendorf S. (2011) Alteration of ferrihydrite reductive
 dissolution and transformation by adsorbed As and structural Al: Implications for As

- 978 retention. *Geochim. Cosmochim. Acta* **75**, 870–886.
- Matta R., Hanna K. and Chiron S. (2008) Oxidation of phenol by green rust and hydrogen
 peroxide at neutral pH. *Sep. Purif. Technol.* 61, 442–446.
- Mitsunobu S., Harada T. and Takahashi Y. (2008a) Comparison of antimony behavior with
 arsenic under various soil redox conditions. *Chinese J. Geochemistry* 25, 98–99.
- 983 Mitsunobu S., Muramatsu C., Watanabe K. and Sakata M. (2013) Behavior of antimony(V)
- 984 during the transformation of ferrihydrite and its environmental implications. *Environ.*985 *Sci. Technol.* 47, 9660–9667.
- Mitsunobu S., Takahashi Y. and Sakai Y. (2008b) Abiotic reduction of antimony(V) by green
 rust (Fe4(II)Fe2(III)(OH)12SO4 · 3H2O). *Chemosphere* 70, 942–947.
- Mitsunobu S., Takahashi Y., Sakai Y. and Inumaru K. (2009) Interaction of synthetic sulfate
 green rust with antimony(V). *Environ. Sci. Technol.* 43, 318–323.
- Mitsunobu S., Takahashi Y. and Terada Y. (2010a) μ-XANES evidence for the reduction of
 Sb(V) to Sb(III) in soil from Sb mine tailing. *Environ. Sci. Technol.* 44, 1281–1287.
- Mitsunobu S., Takahashi Y., Terada Y. and Sakata M. (2010b) Antimony(V) incorporation
 into synthetic ferrihydrite, goethite, and natural iron oxyhydroxides. *Environ. Sci. Technol.* 44, 3712–3718.
- Mittal V. K., Bera S., Narasimhan S. V. and Velmurugan S. (2011) Sorption of Sb(III) on
 carbon steel surface in presence of molybdate and selenite in citric acid medium. *Appl. Surf. Sci.* 258, 1525–1530.
- Muehe E. M., Morin G., Scheer L., Le Pape P., Esteve I., Daus B. and Kappler A. (2016)
 Arsenic(V) Incorporation in Vivianite during Microbial Reduction of Arsenic(V)Bearing Biogenic Fe(III) (Oxyhydr)oxides. *Environ. Sci. Technol.* 50, 2281–2291.
- Murad E. and Fisher W. R. (1987) The geobiochemical cycle of iron. In *Iron in Soils and Clay Minerals* (eds. J. W. Stucki, B. A. Goodman, and U. Schwertmann). Springer

- 1003 Netherlands, Dordrecht. pp. 1–18.
- Nealson K. H., Belz A. and Mckee B. (2002) Breathing metals as a way of life : geobiology in
 action. *Anton. Leeu. Int. J. G.* 81, 215-222.
- 1006 Néel L. (1949) Théorie du traînage magnétique des ferromagnétiques en grains fins avec
 1007 applications aux terres cuites. *Ann. Geophys.* 5, 99–136.
- 1008 O'Loughlin E., Gorski C. and Scherer M. (2015) Effects of Phosphate on Secondary Mineral
- Formation During the Bioreduction of Akaganeite (B-FeOOH): Green Rust Versus
 Framboidal Magnetite. *Curr. Inorg. Chem.* 5, 214–224.
- 1011 O'Loughlin E. J., Larese-Casanova P., Scherer M. and Cook R. (2007) Green Rust Formation
- 1012 from the Bioreduction of γ –FeOOH (Lepidocrocite): Comparison of Several Shewanella
 1013 Species. *Geomicrobiol. J.* 24, 211–230.
- 1014 Official Journal L 129 18/05/1976. Council of the European Communities. (1976) CEC.
- 1015 Council directive 76/464/EEC of 4 May 1976 on pollution caused by certain dangerous
- 1016 substances dis- charged into the aquatic environment of the Community.,
- 1017 Okkenhaug G., Amstätter K., Lassen Bue H., Cornelissen G., Breedveld G. D., Henriksen T.
- and Mulder J. (2013) Antimony (Sb) contaminated shooting range soil: Sb mobility and
 immobilization by soil amendments. *Environ. Sci. Technol.* 47, 6431–6439.
- 1020 Okkenhaug G., Grasshorn Gebhardt K.-A., Amstaetter K., Lassen Bue H., Herzel H.,
- 1021 Mariussen E., Rossebø Almås Å., Cornelissen G., Breedveld G. D., Rasmussen G. and
- 1022 Mulder J. (2016) Antimony (Sb) and lead (Pb) in contaminated shooting range soils: Sb
- and Pb mobility and immobilization by iron based sorbents, a field study. *J. Hazard. Mater.* 307, 336–343.
- 1025 Okkenhaug G., Smebye A. B., Pabst T., Amundsen C. E., Sævarsson H. and Breedveld G. D.
- 1026 (2018) Shooting range contamination: mobility and transport of lead (Pb), copper (Cu)
- and antimony (Sb) in contaminated peatland. J. Soils Sediments 18, 3310–3323.

Olowe A. A., Refait P. and Génin J. M. R. (1990) Superparamagnetic behaviour of goethite
 prepared in sulphated medium. *Hyperfine Interact.* 57, 2037–2043.

1030 Ona-Nguema G., Abdelmoula M., Jorand F., Benali O., Géhin A., Block J. C. and Génin J. M.

- 1031 R. (2002) Iron(II,III) hydroxycarbonate green rust formation and stabilization from
 1032 lepidocrocite bioreduction. *Environ. Sci. Technol.* 36, 16–20.
- 1033 Ona-Nguema G., Carteret C., Benali O., Abdelmoula M., Génin J. M. and Jorand F. (2004)
- 1034 Competitive formation of hydroxycarbonate green rust 1 versus hydroxysulphate green 1035 rust 2 in Shewanella putrefaciens cultures. *Geomicrobiol. J.* **21**, 79–90.
- 1036 Ona-Nguema G., Morin G., Wang Y., Menguy N., Juillot F., Olivi L., Aquilanti G.,
- 1037 Abdelmoula M., Ruby C., Bargar J. R., Guyot F., Calas G. and Brown G. E. (2009)
- 1038 Arsenite sequestration at the surface of nano-Fe(OH)2, ferrous-carbonate hydroxide, and
- green-rust after bioreduction of arsenic-sorbed lepidocrocite by Shewanella putrefaciens. *Geochim. Cosmochim. Acta* 73, 1359–1381.
- 1041 Oorts K., Smolders E., Degryse F., Buekers J., Gascó G., Cornelis G. and Mertens J. (2008)
 1042 Solubility and Toxicity of Antimony Trioxide (Sb2O3) in Soil. *Environ. Sci. Technol.*1043 42, 4378–4383.
- Pallud C., Kausch M., Fendorf S. and Meile C. (2010) Spatial patterns and modeling of
 reductive ferrihydrite transformation observed in artificial soil aggregates. *Environ. Sci. Technol.* 44, 74–79.
- Pedersen H. D., Postma D., Jakobsen R. and Larsen O. (2005) Fast transformation of iron
 oxyhydroxides by the catalytic action of aqueous Fe(II). *Geochim. Cosmochim. Acta* 69,
 3967–3977.
- 1050 Perez J. P. H., Tobler D. J., Thomas A. N., Freeman H. M., Dideriksen K., Radnik J. and
- 1051 Benning L. G. (2019) Adsorption and Reduction of Arsenate during the Fe2+-Induced
- 1052 Transformation of Ferrihydrite. *ACS Earth Sp. Chem.* **3**, 884–894.

- Perez J. P. H., Schifler A. A., Rubio S. N., Reischer M., Overheu N. D., Benning L. G. and
 Tobler D. J. (2021) *J. Hazard. Mat.* 401, 123327.
- Roh Y., Zhang C. L., Vali H., Lauf R. J., Zhou J. and Phelps T. J. (2003) Biogeochemical and
 environmental factors in Fe biomineralization: Magnetite and siderite formation. *Clays Clay Miner.* 51, 83–95.
- Rusch B., Génin J.-M. R., Ruby C., Abdelmoula M. and Bonville P. (2008) Ferrimagnetic
 properties in FeII–III (oxy)hydroxycarbonate green rusts. *Solid State Sci.* 10, 40–49.
- 1060 Schwertmann U., Stanjek H. and Becher H.-H. (2004) Long-term in vitro transformation of 2-
- 1061 line ferrihydrite to goethite/hematite at 4, 10, 15 and 25°C. *Clay Miner.* **39**, 433–438.
- Shebanova O. N. and Lazor P. (2003) Raman spectroscopic study of magnetite (FeFe2O4): A
 new assignment for the vibrational spectrum. *J. Solid State Chem.* 174, 424–430.
- 1064 Silva V. A. J., Andrade P. L., Silva M. P. C., Bustamante D. A., De Los Santos Valladares L.
- and Albino Aguiar J. (2013) Synthesis and characterization of Fe3O4 nanoparticles
 coated with fucan polysaccharides. *J. Magn. Magn. Mater.* 343, 138–143.
- 1067 Srivastava C. M., Shringi S. N. and Babu M. V. (1981) Mössbauer study of the low-1068 temperature phase of magnetite. *Phys. Status Solidi* **65**, 731–735.
- Sung H.-R., Yoon J.-H. and Ghim S.-Y. (2012) Shewanella dokdonensis sp. nov., isolated
 from seawater. *Int. J. Syst. Evol. Microbiol.* 62, 1636–1643.
- Takahashi T., Shozugawa K. and Matsuo M. (2010) Contribution of amorphous iron
 compounds to adsorptions of pentavalent antimony by soils. *Water. Air. Soil Pollut.* 208,
 1073 165–172.
- 1074 Thomas N. A., Eiche E., Göttlicher J., Steininger R., G. Benning L., M. Freeman H.,
 1075 Dideriksen K. and Neumann T. (2018) Products of Hexavalent Chromium Reduction by
- 1076 Green Rust Sodium Sulfate and Associated Reaction Mechanisms. *Soil Syst.* **2**, 58.
- 1077 Trolard F., Génin J. M. R., Abdelmoula M., Bourrié G., Humbert B. and Herbillon A (1997)

- 1078 Identification of a green rust mineral in a reductomorphic soil by Mössbauer and Raman
 1079 spectroscopies. *Geochim. Cosmochim. Acta* 61, 1107–1111.
- 1080 Tronc E., Belleville P., Jolivet J. P. and Livage J. (1992) Transformation of ferric hydroxide
 1081 into spinel by iron(II) adsorption. *Langmuir* 8, 313–319.
- 1082 Tsapin A. I., Nealson K. H., Meyers T., Cusanovich M. A., Van Beuumen J., Crosby L. D.,
- Feinberg B. A. and Zhang C. (1996) Purification and properties of a low-redox-potential
 tetraheme cytochrome c3 from Shewanella putrefaciens. *J. Bacteriol.* 178, 6386–6388.
- 1085 United States Environmental Protection Agency (1976) Water related fate of the 129 priority
 1086 pollutants. USEPA, EP-440/4-79- 029A., Washington, DC.
- Usman M., Byrne J. M., Chaudhary A., Orsetti S., Hanna K., Ruby C., Kappler A. and
 Haderlein S. B. (2018) Magnetite and Green Rust: Synthesis, Properties, and
 Environmental Applications of Mixed-Valent Iron Minerals. *Chem. Rev.* 118, 3251–
 3304.
- Usman M., Hanna K., Abdelmoula M., Zegeye A., Faure P. and Ruby C. (2012) Formation of
 green rust via mineralogical transformation of ferric oxides (ferrihydrite, goethite and
 hematite). *Appl. Clay Sci.* 64, 38-43.
- 1094 Vandenberghe R. E., Barrero C. A., da Costa C. M., Van San E. and De Grave E. (2000)
- 1095 Mössbauer characterization of iron oxides and (oxy)hydroxides: the present state of the 1096 art. *Hyperfine Interact.* **126**, 247–259.
- 1097 Viollier E., Inglett P. W., Hunter K., Roychoudhury A. N. and Cappellen P. Van (2000) The
 1098 Ferrozine method revisted. *Appl. Geochemistry* 15, 785–790.
- 1099 Voegelin A., Senn A. C., Kaegi R. and Hug S. J. (2019) Reductive dissolution of As(V)-
- bearing Fe(III)-precipitates formed by Fe(II) oxidation in aqueous solutions. *Geochem*. *Trans.* 20, 1–13.
- 1102 Waychunas G. A., Fuller C. C., Rea B. A. and Davis J. A. (1996) Wide angle X-ray scattering

- (WAXS) study of "two-line" ferrihydrite structure: Effect of arsenate sorption and
 counterion variation and comparison with EXAFS results. *Geochim. Cosmochim. Acta* **60**, 1765–1781.
- Williams A. G. B. and Scherer M. M. (2001) Kinetics of Cr(VI) Reduction by Carbonate
 Green Rust. *Environ. Sci. Technol.* 35, 3488–3494.
- Williams A. G. B. and Scherer M. M. (2004) Spectroscopic evidence for Fe(II)-Fe(III)
 electron transfer at the iron oxide-water interface. *Environ. Sci. Technol.* 38, 4782–4790.
- Wilson S. C., Lockwood P. V., Ashley P. M. and Tighe M. (2010) The chemistry andbehaviour of antimony in the soil environment with comparisons to arsenic: A critical
- 1112 review. *Environ. Pollut.* **158**, 1169–1181.
- Xu W., Wang H., Liu R., Zhao X. and Qu J. (2011) The mechanism of antimony(III) removal
 and its reactions on the surfaces of Fe-Mn Binary Oxide. *J. Colloid Interface Sci.* 363,
 320–326.
- Yang L., Steefel C. I., Marcus M. A. and Bargar J. R. (2010) Kinetics of Fe(II)-catalyzed
 transformation of 6-line ferrihydrite under anaerobic flow conditions. *Environ. Sci. Technol.* 44, 5469–5475.
- 1119 Zegeye A., Bonneville S., Benning L. G., Sturm A., Fowle D. A., Jones C., Canfield D. E.,
- 1120 Ruby C., MacLean L. C., Nomosatryo S., Crowe S. A. and Poulton S. W. (2012) Green
- rust formation controls nutrient availability in a ferruginous water column. *Geology* **40**.
- 1122 Zegeye A., Mustin C. and Jorand F. (2010) Bacterial and iron oxide aggregates mediate
 1123 secondary iron mineral formation: Green rust versus magnetite. *Geobiology* 8, 209–222.
- 1124 Zegeye A., Ruby C. and Jorand F. (2007) Kinetic and thermodynamic analysis during
 1125 dissimilatory *y*-FeOOH reduction: Formation of green rust 1 and magnetite.
- 1126 *Geomicrobiol. J.* **24**, 51–64.
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1130 FIGURES CAPTIONS

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Figure 1: X-ray diffraction pattern of Fh (black line) and Sb substituted Fh (Fh_Sb(0.04) blue
line, Fh_Sb(0.06) red line and Fh_Sb(0.1) green line). The two maxima (0.258 nm and 0.149
nm) are consistent with 2-line Fh.

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Figure 2: Bio-reduction of Fh (black symbols) and Sb substituted Fh, (Fh_Sb(0.04) Fh_Sb(0.06) and Fh_Sb(0.1) blue, red and green symbols respectively) measured as Fe(II)_{tot}/Fe_{tot}. The initial rate of reduction was computed from the first derivation of a nonlinear curve fit represented by the dotted lines. Errors bars represent standard deviation of three triplicates batches incubations.

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Figure 3: X-ray diffractograms of microbially reduced Fh (black line) and Sb substituted Fh.
(Fh_Sb(0.04), Fh_Sb(0.06) Fh_Sb(0.1) blue, red and green lines respectively. "GR1": green
rust 1, "M": magnetite, "S": siderite and "*" unidentified phase.

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Figure 4: TMS spectra of solid phases obtained after the bio-reduction of Fh at different
temperature. Black dots: experimental data; red dashed line: global fitted curve; solid line:
components.

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Figure 5: Raman spectra of biogenic and abiotic minerals A: Raman spectra for two locations
(red and black spectra) on bio-reduced Fh_Sb(0.04) sample and two locations (green and blue
spectra) on bio-reduced Fh_Sb(0.1) sample. B: Raman spectra of (a) abiotic GR1 precipitate,
(b) abiotic GR1 precipitate with phosphate, (c) abiotic GR1 precipitate and addition of Sb(V)
~ 10 minutes after the precipitation, (d), (e) two locations for abiotic GR1 precipitated with

Sb(V). The analyses were done after a week of aging. M: Magnetite, S: siderite, GR: GreenRust, ?: unassigned phase.

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Figure 6: TMS spectra of solid phases obtained after the bio-reduction of Sb substituted Fh at
16 K. Black dots: experimental data; red dashed line: global fitted curve; solid line:
components.

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Figure 7: XPS high resolution spectra of (A) $Fe2p_{3/2}$ core level before (upper panel) and after bio-reduction (lower level), (B) Sb $3d_{5/2}$ core level before (upper panel) and after bioreduction (lower level), (C) C1s core level before (upper panel) and after bio-reduction (lower level).

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Figure 8: Relative abundance distribution of biogenic minerals as a function of Sb/(Sb + Fe)
based on Mössbauer data.

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1171 **TABLES CAPTIONS:**

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Table 1: Rate and extent of Fh and substituted Fh bio-reduction. Aqueous Sb were measured
at the end of the bio-reduction period (*i.e.* 25 days). Numbers in parenthesis are the standard
deviation of triplicates analyses.

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1177 **Table 2:** Mössbauer hyperfine parameters at 15 K and relative abundance of components in 1178 bio-reduced Fh_Sb series. $CS = center shift relative to \alpha-iron at room$

- 1179 temperature, $\Delta(\varepsilon)$ = Quadrupole splitting (shift), H = magnetic hyperfine field, RA = Relative
- 1180 abundance.
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