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“Contrasted Microbial community colonization of a bauxite residue deposit marked by a complex geochemical context.”

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Abstract

Bauxite residue is the alkaline byproduct generated during alumina extraction and is commonly landfilled in open-air deposits. The growth in global alumina production have raised environmental concerns about these deposits since no large-scale reuses exist to date. Microbial-driven techniques including bioremediation and critical metal bio-recovery are now considered sustainable and cost-effective methods to revalorize bauxite residues. However, the establishment of microbial communities and their active role in these strategies are still poorly understood. We thus determined the geochemical composition of different bauxite residues produced in southern France and explored the development of bacterial and fungal communities using Illumina high-throughput sequencing. Physicochemical parameters were influenced differently by the deposit age and the bauxite origin. Taxonomical analysis revealed an early-stage microbial community dominated by haloalkaliphilic microorganisms and strongly influenced by chemical gradients. Microbial richness, diversity and network complexity increased significantly with the deposit age, reaching an equilibrium community composition similar to typical soils after decades of natural weathering. Our results suggested that salinity, pH, and toxic metals affected the bacterial community structure, while fungal community composition showed no clear correlations with chemical variations.
1. INTRODUCTION

Bauxite residue is a solid by-product generated during the production of alumina from bauxite. Alumina extraction is most often performed by the Bayer process, where bauxite is digested with large quantities of sodium hydroxide at temperatures between 150 and 250 °C (Evans, 2016). Depending on the process parameters and the origin of the bauxite, 0.7 to 2 tons of bauxite residues are produced per ton of alumina (International Aluminium Institute and European Aluminium, 2015). Over the last 10 years, the average annual production of alumina was 116 ± 15 Mt (World Aluminium, 2020), which corresponds to ~200 Mt of bauxite residue produced per year. In the absence of economically profitable large-scale applications, bauxite residues are commonly landfilled in large open-air deposit areas (BRDA), reaching a total of 4.5 Bt bauxite residue in storage at the present time (Denton et al., 2021).

Due to the physicochemical characteristics of bauxite residue i.e., high alkalinity, high salinity, high metal content and lack of nutrients, BRDAs pose a challenge to most living organisms (Di Carlo et al., 2020; Santini et al., 2015b) and represent a potential source of contamination for the surrounding terrestrial and aquatic ecosystems (Bouchoucha et al., 2019; Ren et al., 2018). Consequently, efforts have been made to remediate these deposits, usually by the addition of amendments (e.g. gypsum and organic matter) that attenuate their harsh conditions prior to the revegetation of the area (Bray et al., 2018; Khaitan et al., 2010). BRDAs have also been considered as a promising secondary source of valuable and critical metals, as these elements are concentrated by a factor of 2 in bauxite residues compared to the initial bauxite ore (Panda et al., 2021; Ujaczki et al., 2017; Vind et al., 2018). Critical metals are chemical elements characterized by their high economic relevance and supply risk (European Commission, 2020) and include metals crucial for information and energy technologies such as Co, Mg, Ba, V, Ge, Nb, Sr, Ga, platinoids (PGMs), and rare earth elements (REEs: Lanthanides (Ln), Y, Sc).

Recently, microbial-driven approaches have gained attention as viable and cost-effective methods for the management and valorization of urban and industrial wastes, including bauxite residue (Lyu et al., 2021; Panda et al., 2021; Santini et al., 2019). During in-situ remediation processes, native microbial communities have been shown to decrease the pH and salinity of the bauxite residue, as well as play an essential role in soil formation and plant growth (Di Carlo et al., 2019; Tian et al., 2020; Wu et al.,
Microbial metal recovery strategies are also promising for the dissolution and recovery of elements of interest (e.g., Fe, Al, critical metals) from waste (Baniasadi et al., 2019; Dev et al., 2020; Maes et al., 2016). Out of all these techniques, bioleaching and microbial electrochemistry are among the most-studied methods (Domínguez-Benetton et al., 2018; Srichandan et al., 2019). In bauxite residues, these techniques are still at an early experimentation stage, although some promising results for REE recovery by bioleaching have been reported (Kiskira et al., 2021; Qu et al., 2019; Zhang et al., 2020).

More generally, microbial-driven strategies for waste management imply the use of microbial communities native to contaminated sites, which tend to thrive better in such harsh environments due to their unique metabolisms developed through natural selection (Ghosh et al., 2018; Ma et al., 2019; Roy et al., 2018; Sajjad et al., 2020). In addition, microbial communities are known to perform complex functions and are more robust to environmental fluctuations compared to pure cultures (Perez-Garcia et al., 2016; Wang et al., 2020; Zhang et al., 2008). Metagenomics studies based on Next Generation Sequencing have significantly expanded the identification and dynamics of microbial communities involved in different bioprocesses. However, the evolution of microbial communities in metal biorecovery experiments is still limited to a few weeks (Ma et al., 2017; Sajjad et al., 2020; Wang et al., 2020). Regarding bauxite residues, the identification of pioneer microbial communities and their dynamics during primary succession have been highlighted as research needs to improve microbial-driven bioremediation (Santini et al., 2015a). Nevertheless, to date, studies on microbial diversity in bauxite residue have focused on the responses of bacterial communities to remediation strategies, overlooking their active role in the process (Banning et al., 2011; Fourrier et al., 2020; Ke, 2021; Krishna et al., 2014; Schmalenberger, 2013; Wu et al., 2020).

This study aimed at exploring the microbial dynamics during primary succession in a BRDA from Southern France using an integrated physicochemical and biological approach. The specific objectives of this study were to (a) evaluate the effect of the deposit age and the ore origin on the geochemical characteristics of bauxite residue; (b) identify the first microbial communities colonizing BRDAs and their role in the establishment of new species; (c) determine the microbial community structure in the equilibrium stage of primary succession in BRDAs and the main geochemical factors
driving it; (d) explore possible implications of this study for microbial-driven bioremediation and critical metal recovery.

2. MATERIALS AND METHODS

2.1 Site description and sampling

The bauxite residues come from a refinery operating the Bayer aluminum extraction process since 1894 in Gardanne, Southern France. The sampling took place at the Mange-Garri bauxite residue disposal area (BRDA) in Bouc-Bel-Air, Southern France. This region is characterized by a hot-summer Mediterranean climate, with average annual precipitation and temperature of 485 mm and 16.3°C (Copernicus Climate Data Store, 2021). We selected four distinct areas of the BRDA based on their deposit age (1, 2, 90, and 100 years) and bauxite origin (lateritic bauxite from Boké, Guinea (Bo) and karstic bauxite from Provence, France (Pr)), named Bo1, Bo2, Pr90 and Pr100 (Fig. 1). At the Bo1 and Bo2 sites, bauxite residues were produced from Guinean bauxite and were landfilled one and two years before sampling respectively. These deposits are occasionally watered to prevent dust dispersion and were not amended at the time of sampling. The Pr90 and Pr100 sites contain bauxite residues that were produced from Provençal bauxite and were deposited around 90 and 100 years ago respectively. These sites were amended with a soil layer in the 1960s and a slight coverage with low-lying vegetation can be seen.

At each of the four sites, samples were collected in triplicate, gathering approximately 500 g of bauxite residue at a depth of 20 to 30 cm in sterile plastic bags. At sites Pr90 and Pr100, samples were collected at the edge of the deposit, where bauxite residues were distant from the added soil layer. In addition, freshly produced bauxite residue (4 replicates) was also selected to analyze the initial bio-geochemical characteristics before landfill. Samples were divided into two groups based on the following analyses. Samples undergoing physicochemical analysis were oven-dried at 70°C, grounded in a mortar, passed through a 130 μm sieve, and stored in metal-free plastic tubes. Samples for microbial community analyses were stored at 4°C.
2.2 Physicochemical analyses

2.2.1 Elemental composition

First, pH and electrical conductivity (EC) were measured at a solid/MilliQ water ratio of 1:5. Then, the elemental composition of the bauxite residues was determined after alkaline fusion of the samples (Rivera et al., 2019). Briefly, 1 g of each sample was digested with 750 mg of lithium tetraborate (Li$_2$B$_4$O$_7$) at 1000°C for 30 min (WiseTherm F/FH 0-1200°C) and then immediately dissolved in 40 mL of HNO$_3$ 1N. Prior to elemental analysis, the samples were dissolved 200 times. Subsequently, inductively coupled plasma mass spectrometry (ICP-MS, Perkin Elmer NexIon 300X) was used to analyze elements with concentrations up to 1000 µg/L, while elements with concentrations above 1000 µg/L were analyzed by inductively coupled plasma optical emission spectroscopy (ICP-OES, Perkin Elmer 4300 DV). All analyses were conducted for major and trace elements commonly found in bauxite residues, namely Fe, Ti, Al, Ca, Zr, Cr, Mn, Zn, Th, Ni, Pb, Cu, Sn, Rb, Cs as well as critical raw elements (as defined by the European Commission (2020): P, Mg, V, Lanthanides (Ln), Sr, Y, Nb, Ba, Sc, Ga, Hf, Ge, Co, Sb, W, Ta). Total carbon (TC) and nitrogen were determined by dry combustion using an Elemental Analyzer (Flash EA, Thermo Scientific). To measure the total organic carbon (TOC), samples were treated with HCl to remove carbonates before analysis.

2.2.2 Mineralogical composition

The mineralogical analyses of bauxite residue were carried out by X-ray diffraction (XRD) using a PANalytical X’Pert Pro (Malvern Panalytical, UK) diffractometer equipped with a cobalt tube (λ = 1.79 Å) running at 40kV and 40mA. Samples were
deposited on low background silicon plates and analyzed from 8° to 80° (2θ) with a step size of 0.033° and a total counting time of 7 hours. Samples were also spun at 15 rpm to improve statistics. Phase identification was performed using the X’pert Highscore plus software (PANalytical) together with the PDF-2 ICDD database (International Center for Diffraction data, Powder Diffraction Files 2). Profex software (Doebelin and Kleeberg, 2015) was used for Rietveld refinement to semi-quantify the proportions of the minerals in the bauxite residues. The following parameters were refined: zero point shift, sample displacement, cell parameters, preferred orientation and peak broadening resulting from the size of the crystallite and the micro strain.

2.3 DNA extraction, PCR amplification and sequencing

The total DNA was extracted from 5 g of each bauxite residue samples using the FastDNA® Spin Kit for soil (MP Biomedicals, USA) following the manufacturer’s protocol. Extracted DNA was used as template in separate PCR reactions amplifying the bacterial 16S and the fungal ITS rRNA gene sequences. For bacterial diversity analysis, we used the primers 341F (5’-CCTAYGGGGRBGCASCAG-3’) and 806R (5’-GGACTACNNNGGTATCTAAT-3’), targeting the V3 and V4 variable regions of the bacterial 16S rRNA (Caporaso et al., 2011; Muyzer et al., 1993). For fungal diversity analysis, we used the primers fITS7 (5’-GTGARTCATCGAATCTTTG-3’) and ITS4 (5’-TCCTCCGCTTATTGATATGC-3’), targeting the ITS2 region (Ihrmark et al., 2012). The amplification conditions were as follows: initial denaturation at 95 °C for 2 min; followed by 34 cycles of denaturation at 95 °C for 30 s, annealing at 55 °C for 30 s, and extension at 72 °C for 1 min; and a final extension at 72 °C for 5 min (Muller et al., 2021). PCR products were then purified using ProNex® Size-Selective Purification System (Promega, USA) and sequenced on Illumina MiSeq platform (Biofidal, Vaulx-en-Velin, France). The raw sequence reads generated from this study have been deposited in the National Center for Biotechnology Information (NCBI) Sequence Read Archive under the accession number PRJNA748554.

2.4 Sequencing data processing

Microbiome bioinformatics were performed by the open-source software QIIME2, version 2019.10 (https://qiime2.org) (Bolyen et al., 2019). Raw reads were quality-filtered, denoised and chimera-checked using DADA2 (Callahan et al., 2016). DADA2 uses a parametric model to infer true biological sequences from reads. The model
relies on input read abundances (true reads are likely to be more abundant) and the pairwise similarity between sequences. The taxonomic annotation of the resulting sequence variants (ASVs) was assigned using the feature-classifier command with default parameters in QIIME2 and sequences were matched against the Greengenes 13_8 database (McDonald et al., 2012). Finally, scaling with ranked subsampling (SRS) curves (Beule and Karlovsky, 2020) were drawn to determine whether the sequencing depth was sufficient to represent the true diversity of the samples.

2.5 Biodiversity parameters and microbial biomarker discovery

Alpha diversity was explored through observed ASVs, Chao1, and Shannon. The observed ASVs and the Chao1 estimator were selected to identify community richness, and Shannon index was used to assess community diversity (Callahan et al., 2017; Hill et al., 2003). Beta diversity analysis was used to evaluate distribution patterns in samples based on bacterial and fungal ASV composition (Anderson et al., 2011; Callahan et al., 2017). For this purpose, a principal coordinate analysis (PCoA) based on weighted UniFrac distances (Lozupone et al., 2007) was conducted. All alpha and beta diversity metrics were calculated using QIIME 2 after normalization to 59388 and 4778 sequences per sample for bacteria and fungi, respectively. Linear discriminant analysis effect sizes algorithm (LEfSe) (Segata et al., 2011) was performed on the Galaxy platform (https://huttenhower.sph.harvard.edu/galaxy/) to identify bacterial biomarkers characterizing the samples. LEfSe couples Kruskal–Wallis tests for measuring statistical significance with quantitative tests for biological consistency (Wilcoxon rank sum test).

2.6 Statistical analysis

All statistical analyses were performed with the open-source software R (R Core Team, 2020) using the packages “dplyr” (Wickham et al., 2021), “vegan” (Oksanen et al., 2020), “car” (Fox and Weisberg, 2019), “ggpubr” (Kassambara, 2020), and “rstatix” (Kassambara, 2021). Figures were produced with the package “ggplot2” (Wickham, 2009). To study the ore-dependent differences in chemical and microbiological compositions of bauxite residue, unpaired two-sided T-tests and Wilcoxon rank sum tests were conducted respectively. Furthermore, Games-Howell post-hoc tests were used to assess the age-dependent variations in chemical and microbial compositions. Pearson correlations between each variable and the deposit age were used to further
explore these relationships. Also, significant differences in alpha diversity indices were tested by Wilcoxon rank sum tests.

Principal component analysis (PCA) was used to identify the variables that explain most of the variation in chemical composition. In addition, unsupervised hierarchical clustering was applied to the PCA to group the samples according to their chemical similarity. Both PCA and hierarchical clustering were performed using the packages “FactoMineR” (Lê et al., 2008) and “factoextra” (Kassambara and Mundt, 2020). PERMANOVA via the adonis function was conducted in both PCA and PCoA to test for the chemical and microbiological dissimilarities based on the deposit age and the bauxite origin. To study the factor age alone, nested PERMANOVA were calculated using the parameter strata to exclude the effect of bauxite origin. Furthermore, multiple co-inertia analysis (MCIA) was performed using the package “omicade4” (Meng et al., 2014) to determine the relationships between the four datasets used in this study (chemical characteristics, mineralogy, 16S rRNA sequences and ITS sequences).

2.7 Co-occurrence network construction

Co-occurrence analyses were implemented for a better understanding of bacterial and fungal interactions in the four bauxite residues. Co-occurrence networks were constructed based on pairwise Pearson correlations calculated between bacterial and fungal ASVs by using the base R function cor (Berry and Widder, 2014; Williams et al., 2014). To avoid including false positives in the network due to spurious or random interactions, the ASV table we permuted 100 times and a p-value for each possible pairwise interaction was calculated to test its validity. The p-values were then adjusted using the Benjamini-Hochberg procedure (Benjamini and Hochberg, 1995), and only edges with a p-value below 0.01 that corresponded to an absolute correlation higher than 0.4 were retained. To describe the topology properties of the networks, a set of network indexes including graph density, average degree of nodes, transitivity, modularity, average geodesic distance, betweenness centrality, and density were calculated with the R package “igraph” (Csardi and Nepusz, 2006). The Wilcoxon test was employed to assess significant differences in topological parameters between networks. The connectivity of the network nodes was determined by their within-module connectivity (Zi) and among-module connectivity (Pi). Nodes were then classified into four categories, according to Poudel et al. (2016): peripherals (Zi < 2.5 and Pi < 0.62, nodes with few links to other species), connectors (Pi > 0.62, nodes
that connect modules), module hubs (Zi > 2.5, highly connected nodes within
modules), and network hubs (Zi > 2.5 and Pi > 0.62, highly connected nodes among
and within modules).

3. RESULTS

3.1. pH, EC and geochemical composition

The pH of the bauxite residue ranged from 8.70 ± 0.34 to 11.90 ± 0.01, with the highest
value in fresh samples and the lowest in Pr100 (Fig. 2a, Table S1). The EC of the
bauxite residue ranged from 0.29 ± 0.15 mS/cm to 2.53 ± 0.04 mS/cm, and again,
fresh samples showed the highest value and Pr100 the lowest (Fig. 2a, Table S1).

The elements with the highest concentrations in all bauxite residue samples were Fe
(268,894 ± 45,706 mg/kg), Ti (51313 ± 17317 mg/kg), Al (33093 ± 5662 mg/kg), Ca
(22425 ± 4952 mg/kg), and TC (17162 ± 9439 mg/kg) (Fig. 2b, Table S2). Fe
represents at least 22 ± 3 % of the dried mass in all samples. The main critical
elements present in all samples were Ti (51313 ± 17317 mg/kg), P (2743 ± 1274
mg/kg), Mg (1877 ± 987 mg/kg), V (849 ± 423 mg/kg), Ln (775 ± 339 mg/kg), Sr (240
± 142 mg/kg), Y (159 ± 59 mg/kg), Nb (99 ± 16 mg/kg), Sc (73 ± 12 mg/kg), Ba (73 ±
15 mg/kg), and Ga (56 ± 24 mg/kg) (Fig. 2c-d, Table S2). TN concentrations were
only above the machine’s limit of detection (100 mg/kg) in the Provence samples, with
values of 287 ± 183 mg/kg for Pr90 and 470 ± 158 mg/kg for Pr100 (Table S2).

Statistical analysis showed that Ti, Cr, V, and Ga concentrations were significantly
higher in samples coming from Boké bauxite (p < 0.05, Table S2), while samples from
Provence bauxite showed greater abundances of TC, TOC, P, Mg, Ln, Mn, Sr, Y, Sc,
Pb, Ni, and Co (p < 0.05, Table S2). Within these two groups, some elements
concentrations changed over time. In the Boké samples, a significant decrease in Fe,
Ca, Cr, V, Ga, Cu, Ge, and W was observed as sample age increased (p < 0.05, Table
S2) while Mg, and Sr concentrations increased with age (p < 0.05, Table S2). In the
Provence samples, Mg decreased significantly with age, while P and TOC
concentrations increased (p < 0.05, Table S2). These relationships between the age
of the sample and their chemical properties were further confirmed by Pearson
correlations (R ≥ 0.9, p < 0.05, Fig. S1).
PCA was used to detect patterns in the chemical dataset and describe linear relations between the analyzed physicochemical parameters (Fig. 3a). The first three principal component axis (PC1, PC2, and PC3) explained 84.3% of the data variability between samples. PC1 explained 53.0% of the variance and included two groups: the first one
composed of pH, EC, Ga, Cr, V, Cu, Fe, Ge, Ca, Ti, and W whereas the second one comprised TOC, Co, Ni, Sr, Ln, Mg, Pb, Mn, Y, Sb, Cs, and Sn. PC2 explained 18.4% of the variance among samples and was mainly composed of Zr, Sc, Hf, Ta, Al, Zn, Nb, and Ti. Finally, PC3 (not shown) explained 12.9% of variance and was primarily composed of P, Th, Rb, Zn, Cs, and Sn. PERMANOVA analysis corroborated that the factors “age” (p < 0.01) and “bauxite origin” (p < 0.05) significantly explained the differences in chemical properties across samples. The factor “age” (tested by nested PERMANOVA to exclude the effect of the bauxite origin) explained 65.8% of the variance in chemical parameters while the factor “bauxite origin” explained 28.18%.

The hierarchical cluster analysis (Fig. S2) provided statistical confirmation of the tendencies observed in the PCA and bauxite residue samples were clustered based on their chemical similarity. The first cluster contained the fresh samples, the second cluster Bo1 samples, the third cluster Bo2 samples, and the fourth cluster contained both Pr90 and Pr100 samples.

To further understand the effect of the factor “age” on residues generated from the same bauxite ore, a second PCA was conducted considering only the Boké samples (fresh, Bo1 and Bo2) (Fig. 3b). The first two dimensions explained 82.5% of the variability and the factor “age” separated the samples significantly (PERMANOVA, p < 0.01). Co, Sr, Pb, Mg, and Sb were mostly associated with Bo2, while pH, EC, W, Ca, Ge, Fe, Cu, V, Cr, and Ga had their highest values in fresh samples and decreased progressively with age. In contrast, TOC, Y, Mn, Ln, Ni, Sn, Al, and Ta did not show any age-dependent trends during the first years of storage (Table S2). Globally, this statistical analysis is in line with the expected behavior of the elements, such as the colocalization of Cr and V and their correlation with Fe (Markus Gräfe et al., 2011) as well as the REEs affinity for Mn phases (Vind et al., 2018).
Figure 3. Principal component analysis (PCA) of chemical parameters in A) all and B) Boké bauxite residue samples. Individuals are represented by dots and colored by sample. Individuals are grouped according to hierarchical clustering results. Variables are represented by arrows and colored by chemical nature. Only variables with a \( \cos^2 > 0.5 \) are shown.
3.1.3 Mineralogical composition

XRD analysis (Fig. 2d, Table S3) showed that bauxite residues were principally composed of iron and aluminum oxides, hydroxides, and silicates. Goethite (FeO(OH)), hematite (Fe₂O₃), gibbsite (Al(OH)₃), and vishnevite (Na₈(AlSiO₄)₆O₂₄(SO₄)) are the dominant phases and accounted for at least 70% of the crystalline fraction in all the samples. Rutile (TiO₂) and calcite (CaCO₃) were also detected in all samples, in the range of 3-5% and 2-14% respectively. The hematite, anatase (TiO₂) and boehmite (AlO(OH)) were significantly more present in Provence samples, and interestingly, neither anatase nor boehmite were found in fresh samples (Table S3). In Boké samples, a significant decrease in goethite and Katoite (Ca₃Al₂(SiO₄)₁.5(OH)₆) was observed as the age of the samples increased.

3.2 Microbial community diversity, structure and dynamics

3.2.1 Alpha diversity

The influence of bauxite residue’s age and geochemical composition on microbial communities was assessed using 16S and ITS metabarcoding. As expected, no DNA could be extracted from the fresh residue samples. For Bo1, Bo2, Pr90, and Pr100, the sequencing achieved a coverage of more than 55000 16S rRNA sequences and 4700 ITS sequences per sample. The SRS curves (Fig. S3) indicated that the sequencing depth was sufficient to identify the majority of ASVs within bacterial and fungal communities in all the samples (except fresh samples). Alpha-diversity derived from the number of ASVs, Chao1 richness, and Shannon’s index showed differences between samples regarding bacterial and fungal communities (Table S4). A total number of 7994 bacterial ASVs were identified, with 350 to 1318 ASVs per sample. The number of bacterial ASVs in the Provence samples (Pr90 and Pr100) was 6928, significantly greater than the 2867 ASVs found in the Boké samples (Bo1 and Bo2). The Provence samples also showed the highest richness and Shannon values for bacteria, compared with Boké samples (Fig. 4a). For fungi, 892 ASVs ranging from 45 to 156 ASVs per sample were obtained, with again more ASVs and Chao1 richness in Provence samples. However, there were no significant changes in Shannon’s index between both group of samples, indicating that the specific diversity of the fungal communities was similar in Provence and Boké samples.
Figure 4. Bacterial alpha and beta diversity of bauxite residues. A) Chao1 richness and Shannon diversity index. B) Principal coordinate analysis (PCoA) based on weighted UniFrac distances between bauxite residue samples.
3.2.2 Microbial community structure

PCoA ordinations based on weighted UniFrac distances (Fig. 4b) revealed that the bacterial community structure differed between the bauxite residue samples. Samples were significantly separated based on bauxite origin (PERMANOVA, \( p < 0.05 \)), and nested PERMANOVA confirmed the difference in bacterial community structure across age (\( p < 0.001 \)). The first two axes of PCoA explained 74.95% of the community dissimilarity. Axis 1 was most correlated with EC (\( R = 0.97 \), \( p = 1.1 \times 10^{-7} \)), followed by Co (\( R = 0.96 \), \( p = 9.8 \times 10^{-7} \)), Ni (\( R = 0.95 \), \( p = 1.4 \times 10^{-6} \)), Cr (\( R = 0.94 \), \( p = 1.1 \times 10^{-7} \)), V (\( R = 0.92 \), \( p = 2.5 \times 10^{-5} \)), Ti (\( R = 0.9 \), \( p = 5.4 \times 10^{-5} \)), and pH (\( R = 0.9 \), \( p = 8.1 \times 10^{-5} \)); while Axis 2 correlated with P (\( R = 0.75 \), \( p = 0.005 \)). In contrast, the fungal community structure did not seem to follow a clear pattern among samples according to PCoA built on Bray-Curtis distances (Fig. S4).

The bacterial phyla that dominated all bauxite residue samples were Proteobacteria, Actinobacteria, Chloroflexi, Bacteroidetes, Gemmatimonadetes, Planctomycetes, Verrucomicrobia, Acidobacteria and Firmicutes, accounting for more than 90% of the total bacterial communities (Fig. 5). When comparing the samples from Provence and Boké, significant taxonomic differences were observed. Actinobacteria and Gammaproteobacteria were the dominant taxa in the Boké samples with an average abundance of 28.0% and 21.3% respectively, followed by Chloroflexi (9.8%) and Bacteroidetes (9.8%). In the Provence samples, Actinobacteria (18.4%) were still dominant while Gammaproteobacteria (4.8%) and Bacteroidetes (3.6%) lost prominence in favor of Alphaproteobacteria (12.0%), Planctomycetes (9.6%), Betaproteobacteria (6.3%), and Deltaproteobacteria (3.2%). Boké samples were significantly enriched in Actinobacteria, Gammaproteobacteria, Bacteroidetes and Firmicutes, whereas samples from Provence showed greater abundances of Alphaproteobacteria, Planctomycetes, Betaproteobacteria and Deltaproteobacteria (\( p < 0.05 \)). Acidobacteria (10.3%), barely found in Boké samples, showed a strong increase in Provence samples (\( p < 0.05 \)). Chloroflexi, Gemmatimonadetes and Verrucomicrobia were nearly constant across samples.
At the level of bacterial classes, differences between Boké and Provence samples were also found (Fig. 6). Significantly higher proportions of Gammaproteobacteria, Nitriliruptoria, Thermomicrobia, Gemm3, Bacilli, Rhodothermi, TK17, Deinococci and Clostridia were identified in the Boké samples ($p < 0.05$). Within this group, variations between samples of different ages were also observed. Bo1 was enriched in Thermomicrobia, Thermoleophilia, Acidimicrobiia and Bacilli, whereas Opitutae, Verrucomicrobiae and Flavobacteriaia showed greater abundances in Bo2 ($p < 0.05$).

In contrast, Provence samples were significantly more enriched in Alphaproteobacteria, Planctomycetia, Acidobacteria6, Betaproteobacteria, Thermoleophilia, Deltaproteobacteria, Gemmatimonadetes, Pedosphaerae, Anaerolineae, Phycisphaerae, Spartobacteria, Gemm1, Chloracidobacteria, Ellin6529, S085, Nitrospira, Saprospirae, TK10, Solibacteres and TM71 ($p < 0.05$). In the Provence samples, the only differences seen over ages were Pedosphaerae and Actinobacteria, which were more abundant in Pr90 than in Pr100 ($p < 0.05$).
3.2.3 Taxonomic characteristics of bacterial and fungal communities

To further explore these temporal scale differences, LEfSe analyses were conducted to detect asymmetrically distributed bacterial taxa. Among the Boké samples, 9 indicator bacterial taxa were found in Bo1 against 37 in Bo2 (Fig. 7a). LEfSe confirmed the highest abundance of three families of Bacilli in Bo1, as well as other taxa belonging to the phyla Actinobacteria and Proteobacteria, and the archaeal class Halobacteria. In line with our previous results, Opitutae, Verrucomicrobiae and Flavobacteriia were significantly more abundant in Bo2, together with other taxa belonging to the phyla Actinobacteria, Proteobacteria, Firmicutes, Bacteroidetes, Planctomycetes and Nitrospirae. For the Provence samples, only two indicator

Figure 6. Z-score hierarchical clustering heat map representing the relative abundance of the top 40 bacterial classes in different bauxite residue samples.
bacterial taxa (classified as Bacteroidetes) were identified in Pr90, against 28 in Pr100 (classified as Actinobacteria, Proteobacteria, Planctomycetes, Bacteroidetes, Verrucomicrobia, and Chloroflexi) (Fig. 7b).
Figure 7. LEfSe analysis of in the bauxite residue samples according to sample age. Histograms of LDA scores of 16S gene sequences in Boké (A) and Provence (C) samples. Only taxa with a LDA score (log10) above 2.0 and a p-value lower than 0.05 for Kruskal–Wallis tests are shown. B and D Cladograms are derived from LEfSe analysis. The central point denotes the root of the tree and expanded to each ring representing the next lower taxonomic level from phylum to genus.
3.2.4 Microbial co-occurrence networks

Co-occurrence networks were constructed using the ASV data of 16S rRNA and ITS sequences. The Boké (Fig. 8) and Provence (Fig. 9) samples did not differ significantly in network topological parameters such as average geodesic distance and modularity (Table S5). However, the network complexity (number of nodes and edges, degree of nodes, and transitivity) was different between the Boké and Provence networks (Table S5). The two Boké networks (Bo1 and Bo2) contained a similar number of nodes (366 and 369), although Bo1 showed more edges (516 versus 324), higher average degree (2.46 versus 1.76), and transitivity (0.15 versus 0.11). The two Provence networks were more complex, as both Pr90 and Pr100 showed more nodes (603 and 571), more edges (1254 and 1322), higher average degree (4.16 and 4.63), and higher transitivity (0.18 and 0.12) compared to Boké.

Most of the nodes from the Boké and Provence networks were classified as peripherals (specialists), and few nodes fell into module hubs (generalists) (Fig. 8-9). Generalists are considered keystone taxa, as they are responsible for structuring the different nodes and modules into a complete community, thus determining the efficiency of energy metabolism and nutrient cycling in habitats (Wang et al., 2019). In this study, more module hubs (generalists) were identified in Pr90 and Pr100 than in Bo1 and Bo2. In Bo1 and Bo2 networks, the module hubs were mainly members of Bacteroidetes, Actinobacteria, Proteobacteria, and Firmicutes for bacteria, and Ascomycota for fungi. In contrast, the module hubs identified in Pr90 and Pr100 networks belonged to Proteobacteria, Actinobacteria, Chloroflexi, Verrucomicrobia, Gemmatimonadetes, Planctomycetes and Firmicutes for bacteria, and Ascomycota and Basidiomycota for fungi.
Figure 8. Co-occurrence network and Zi-Pi plot of Bo1 (A) and Bo2 (B) bauxite residue samples. Each node represents a bacterial class (circles) or a fungal phyla (triangles). Nodes are colored by taxonomical affiliation and their size in the network is proportional to the node betweenness. The color of each link reflects positive (red) or negative (blue) interactions. The topological role of each node is defined by within-module connectivity (Zi) and among-module connectivity (Pi). According to values of Zi (2.5) and Pi (0.62), the roles of nodes are classified into four categories: Peripherals, Module hubs, Connectors and Network hubs.
Figure 9. Co-occurrence network and Zi-Pi plot of Pr90 (A) an Pr100 (B) bauxite residue samples. Each node represents a bacterial class (circles) or a fungal phyla (triangles). Nodes are colored by taxonomical affiliation and their size in the network is proportional to the node betweenness. The color of each link reflects positive (red) or negative (blue) interactions. The topological role of each node is defined by within-module connectivity (Zi) and among-module connectivity (Pi). According to values of Zi (2.5) and Pi (0.62), the roles of nodes are classified into four categories: Peripherals, Module hubs, Connectors and Network hubs.
3.2.5 Multi co-inertia analysis

Multiple co-inertia analysis (MCIA) was used to determine the relationships between the chemistry, mineralogy, bacterial and fungal community composition in our bauxite residue samples. Figure 10 displays the projection of the four datasets onto the first two principal components (PCs) of MCIA for Bo1, Bo2, Pr90 and Pr100 samples. PC1 and PC2 explained 54% and 19% of the total variation respectively (Fig. S6b). The sample space (Fig. S6a) shows a good correlation between the different datasets in all samples, which could be clearly differentiated. Early-stage bauxite residues (Bo1 and Bo2) were separated from equilibrium-stage samples (Pr90 and Pr100) along the PC1. Pseudo-eigenvalues (Fig. S6c) indicated that the chemistry and bacterial dataset contributed the most to the variation observed in PC1, while mineralogy was weighted in PC2. The pair-wise RV coefficient, which is multivariate generalization of the squared Pearson correlation coefficient, indicated higher global similarity between chemistry and bacterial classes (RV score = 0.86) when compared to the similarities between the chemistry and fungal classes (RV score = 0.36), and between the mineralogy and microbial classes (RV score for bacterial classes = 0.41, RV score for fungal classes = 0.22). Fungal classes contributed little to the variation of PC1 or PC2 and were mostly correlated with bacterial classes (RV score = 0.38).
4. DISCUSSION

4.1 BRDAs behave as primary successional environments

The initial geochemical conditions of the fresh bauxite residue, characterized by high alkalinity (pH = 11.90 ± 0.01) and salinity (EC = 2.53 ± 0.04 mS cm⁻¹) in line with previous studies (Cusack et al., 2019; Xue et al., 2016), make BRDAs hostile environments for most domains of life. The high pH and salinity derive from the excess of alkaline anions (hydroxides, carbonates, aluminates, aluminum hydroxides, silicates) and Na⁺ produced during the Bayer process (Cusack et al., 2019; X. Kong et al., 2017) and thus the bauxite origin does not seem to significantly influence the value of these parameters. Fresh and recently deposited bauxite residues (Bo1 and
Bo2) also contained significant amounts of toxic metals for microorganisms, namely Cr (1594 ± 23 mg/kg), V (1452 ± 10 mg/kg) and Al (38420 ± 920 mg/kg). The high alkalinity (pH > 10) of fresh, Bo1 and Bo2 bauxite residues may favor the Cr (III) oxidation to Cr(IV) as well as the predominance of V(V), both highly mobile and toxic for bacteria, fungi and plants (Chen et al., 2021; Economou-Eliopoulos et al., 2016; Liang et al., 2021; Milačič et al., 2012; Xiao et al., 2017). Regarding aluminum, only 1% of the total Al content in bauxite residue has been reported to be soluble at pH values above 9, mainly as mobile [Al(OH)₄]⁻ (Milačič et al., 2012). Nevertheless, its impact on microbial and plant communities should not be neglected, as even small proportions of the total Al correspond to potentially toxic concentrations. Moreover, fresh and recently deposited bauxite residues were inherently poor in most macronutrients required for microbial and plant growth, such as TOC, TN, and P (Table S2). TOC accounted for less than 0.6% of the mass (0.28 % in fresh residue, 0.33 % in Bo1, and 0.23 % in Bo2) which is typically attributed to desert soils (Li et al., 2013; Zech et al., 2014). TN concentrations were below the detection limit (100 mg/kg) in fresh, Bo1 and Bo2 samples in agreement with previous studies (Krishna et al., 2014; Wu et al., 2020). Finally, although the P concentrations ranged from 1500 to 3000 mg/kg in fresh, Bo1, and Bo2 samples, the available phosphorus in bauxite residue is usually around 6.7 mg/kg (Santini et al., 2015a; Wu et al., 2020).

These extreme conditions, together with the absence of pre-existing life in the freshly produced bauxite residues, indicate that BRDAs behave as primary successional environments, similar to other well-studied natural systems (Guo et al., 2014; Schmidt et al., 2008). Primary succession is characterized by an initial short phase where the microbial community assembly is driven by stochastic dispersal processes, followed by a longer deterministic phase dominated by environmental selection and competition between species (Ortiz-Álvarez et al., 2018). In engineered environments like BRDAs, the dispersal of microorganisms is limited, and environmental selection pressure is usually stronger due to the extreme chemical characteristics of these sites. Therefore, a shorter stochastic phase and a stronger deterministic assembly is expected in such systems (Santini et al., 2015a). Given the environmental stress imposed by the geochemical properties of fresh bauxite residues, the first pioneer microorganisms to colonize BRDAs are likely to be haloalkaliphilic species able to accumulate nutrients.
to facilitate the incorporation of new species (Santini et al., 2015a; Schmidt et al., 2014; Sun et al., 2018).

**4.2 Geochemistry-dependent microbial community assembly in the early stages of primary succession**

The bacterial community composition of the recently deposited bauxite residues (Bo1 and Bo2) was similar to that observed in soda lakes, alkaline residues, and other unamended BRDAs (Bondici et al., 2013; Chakraborty et al., 2021; Kalwasińska et al., 2017; Ohlsson et al., 2019), characterized by the dominance of Actinobacteria (Actinobacteria, Nitriliruptoria, Acidimicrobiia) and Proteobacteria (especially Gammaproteobacteria) and relatively high proportions (1 to 16 %) of Bacteroidetes (Cytophagia, Rhodothermi, Flavobacteria), Firmicutes (Bacilli, Clostridia), and other alkalophilic classes (Opitutae, Deinococci) (Fig. 5-6).

The phyla Actinobacteria, Proteobacteria and Firmicutes contain some of the prokaryotes best adapted to high alkalinity and salinity. Alkaliphilic and alkalinotolerant microorganisms maintain their intracellular pH homeostasis by accumulating H⁺ in their cytoplasm through Na⁺/H⁺ antiporters (Mamo, 2020). Our results showed that Halobacteria (Archaea), Gammaproteobacteria, Nitriliruptoria and Rhodothermi were more abundant in Bo1 (Fig. 6-7) and correlated with EC (Fig. 10), corroborating the preference of these classes for high-salinity niches (Zhang et al., 2019). Extreme halophiles accumulate intracytoplasmic KCl to regulate their osmotic potential, which constrains them to be obligate halophiles and limits their dispersal (Chen et al., 2020; Shameer, 2016; Vaidya et al., 2018). Salinity (EC) decreased significantly from Bo1 to Bo2 (Fig. 1, Table S2), allowing the proliferation of a wider range of halophilic and halotolerant species such as Opitutae, Acidimicrobiia or the genus Halomonas (Gammaproteobacteria) (Fig. 6-7). Halotolerant bacteria can flexibly adapt to habitats with different salinities by synthesizing and accumulating compatible solutes in their cytoplasm, which in turn requires a more intensive use of energy (Gunde-Cimerman et al., 2018). Interestingly, the number of edges and network complexity was higher in Bo1 compared with Bo2 (Fig. 8), indicating that saline stress promoted microbial interaction as reported in other studies (Ji et al., 2019; Wang et al., 2019).

As pioneer colonizers, microorganisms are critical in the biogeochemical cycles and the ensuing development of soil (Schmidt et al., 2008; Zeng et al., 2016). TOC
concentration showed no significant trend in early-stage samples (Fig. 2, Table S2), suggesting that the ratio in TOC input and consumption was balanced. In primary successional environments, the main sources of TOC are airborne allochthonous organic matter and C fixed by autotrophic microbial species (Ciccazzo et al., 2016). Our results also highlighted the significant abundance of heterotrophic bacteria in the early-stage samples, especially in Bo1, enriched in Actinobacteria and Bacteroidetes (Fig. 6-7). MCIA revealed that Cytophagia (Bacteroidetes) and Nitriliruptoria (Actinobacteria) correlated negatively with TOC (Fig. 10) and members of these classes were identified as keystone species in Bo1 (Fig. 8), confirming their relevance in oligotrophic habitats (Foreman et al., 2007; Gonzalez-Pimentel et al., 2018). Heterotrophic communities may represent the earliest stage of microbial assembly, degrading allochthonous organic compounds and supplying essential nutrients (TOC, N, P) for the subsequent development of autotrophic species (Hodkinson et al., 2002). Our results also indicated the presence of chemoautotrophic Proteobacteria typically involved in the primary production of alkaline environments (Fig. 7), including iron-oxidizing bacteria within Alphaproteobacteria (Rhodobacteraceae) and Betaproteobacteria (Comamonadaceae) (Jamieson et al., 2018; Kumaraswamy et al., 2006; Straub et al., 1996) and sulfur-oxidizing Chromatiales (Gammaproteobacteria) (Yuan et al., 2021). In tune with previous studies (Harantová et al., 2017; Schmidt et al., 2008), the abundance of non-symbiotic N-fixing bacteria, represented by members of Clostridia, Opitutae, Comamonadaceae (Betaproteobacteria), Rhodospirillales (Alphaproteobacteria) and Bacilli (Ciccazzo et al., 2016; Hingole and Pathak, 2013; Navarro-Noya et al., 2016), increased significantly after the first years of primary succession (Fig. 5-6). These bacteria could contribute to the slight increase in TN observed after the first years of bauxite residue weathering (Wu et al., 2020), undetectable in this study as it is expected to be less than 100 mg/kg. Nitrifying bacteria (Nitrosomonas and Nitrospiraceae) were also more prevalent in Bo2, confirming that, although the amount of ammonia derived from bacterial fixation is low at early stages of primary succession, it must be sufficient to fuel nitrification (Ollivier et al., 2011; Zeng et al., 2016). Moreover, our results suggest that the increase in N-cycle bacteria could be due to a significant decrease in EC and Cr in Bo2 (Fig. 2, Fig. 10, Table S2), as N-cycle enzymes are particularly sensitive to high salinity (Claros et al., 2010; Herbst, 1998) and heavy metal contamination (Kim et al., 2016; Oliveira and Pampulha, 2006).
4.3 Primary succession reaches an equilibrium in BRDA over few decades

The last stage of primary succession arises after long periods of time when the habitat becomes less harsh and environmental selection loses its strength in the assembly of communities, reaching an equilibrium (Ferrenberg et al., 2013). This maturity is expected to require more time to be developed in extreme geochemical systems as BRDAs (Santini et al., 2015a). In our study, Pr90 and Pr100 samples, which have been naturally weathered for at least 90 years, did not show significant dissimilarities in either physicochemical parameters (Fig. 3a, Tables S1 and S2) or bacterial and fungal community composition (Fig. 4b; Fig. S5), suggesting that primary succession had reached the equilibrium in these sites. In Pr90 and Pr100, pH (8.8 ± 0.3) and salinity (EC = 0.29 ± 0.11 mS cm⁻¹) were significantly lower than in early-stage samples (Fig. 2, Table S1), revealing a significant dependence with the age of the samples. During bauxite residue natural weathering, salinity is generally assumed to decrease due to the dissolution of alkaline minerals by rainfall (Cusack et al., 2019; Zhu et al., 2016), while the decrease in pH is normally attributed to atmospheric and microbial carbonation (Cusack et al., 2019; Khaitan et al., 2010; Schmalenberger, 2013; X. Kong et al., 2017). Moreover, Pr90 and Pr100 showed a significantly higher concentration of macronutrients such as TOC, TN and P, which are accumulated during the bauxite residue natural aging due to the action of the different microorganisms that colonize the BRDAs (Santini et al., 2015a). This negative correlation between pH-salinity and age-nutrients is characteristic of restored and amended BRDAs (Courtney et al., 2014; Wu et al., 2021). Finally, our results also indicated a significant and steady decrease in Cr and V associated with the natural aging of the samples (Fig. 2, Table S2). However, the concentration of these metals also seems to be determined by the bauxite type and, therefore, the effect of the ore origin should not be neglected (Gentzmann et al., 2021).

Both alpha and beta bacterial diversity increased significantly in the equilibrium-stage samples (Fig. 4), in line with the results observed in restored BRDAs (Krishna et al., 2014; Wu et al., 2021, 2020). Beta-diversity analysis confirmed that “age” was the main factor explaining the dissimilarities of microbial community between Bo1/Bo2 and Pr90/Pr100 and not the ore origin. In addition, the changes in beta-diversity could be explained by pH, EC and some metal concentration gradients, all of them strongly anti-correlated with the age of the samples (Fig. 3). Circumneutral pH, lower salinity and
metal concentration, and increase in macronutrients support the survival of a more diverse microbial community dominated by bacteria commonly found in ordinary soils and freshwater like Alphaproteobacteria, Betaproteobacteria, Planctomycetes, Gemmatimonadetes, and Acidobacteria. Acidobacteria are one of the most abundant terrestrial bacterial taxa, reaching 52% from the total bacterial community in certain soils (Kielak et al., 2016). Their abundance is positively correlated with low pH (Jones et al., 2009), and their accumulation over time in BRDAs is commonly associated with the restoration of chemical conditions in bauxite residue (Santini et al., 2015b; Wu et al., 2021). The equilibrium-stage bauxite residue also harbors a more complex microbial network, accounting for more highly connected taxa (nodes) than early-stage residues. Nutrient availability and higher microbial richness have been shown to favor microbial network complexity and the emergence of new ecological functions, making the ecosystem more stable (Qiu et al., 2021; Wagg et al., 2019).

In contrast, fungal biodiversity and community structure did not appear to be significantly affected by chemical gradients (Table S4, Fig. S5). The fungal community structure in early-stage bauxite residues was dominated by Ascomycota (80.1 ± 9.8% of fungal communities), and Basidiomycota (> 14%). This distribution seems to be the usual in natural and engineered haloalkaline environments, although the analysis of fungal communities in these habitats is limited to few studies (Grum-Grzhimaylo et al., 2016; Salano et al., 2017; Santini et al., 2015b). In the equilibrium-stage samples, the relative abundance of Ascomycota decreased (52.1 ± 24.2%) while Basidiomycota (mostly Agaricomycetes) increased (35.0 ± 23.3%). This pattern has been observed in restored BRDAs and was attributed to a decrease in total alkalinity (Santini et al., 2015b). In this study however, MCIA revealed a poor correlation between fungal communities and chemical parameters (RV score = 0.36). It is known that the fungal community assembly is more influenced by stochastic processes in primary successional environments (Schmidt et al., 2014). Unlike bacteria, climate is often considered the main environmental factor affecting the fungal community composition, rather than soil chemical properties (Egidi et al., 2019). Interestingly, all fungal keystones identified in Bo1 (Fig. 8) belonged to the class Dothideomycetes (Pleosporaceae, Macrophoma, Cladosporium), within the Ascomycota. Dothideomycetes constitute almost the entire fungal community in microbial mats from soda lakes and are also found in microbial mats from hypersaline and iron-rich habitats.
In microbial mats, fungi play a crucial role in nutrient recycling by decomposing complex carbohydrates into simpler compounds that fuel chemoheterotrophic species (Carreira et al., 2020).

4.4 Implications for bioremediation and metal recovery

To date, only few studies have focused their efforts on an integrated chemical, physical, and biological approach for the characterization of BRDA in the optic of bioremediation or critical metal bio extraction strategies. Our results highlight the key role of pH and EC for natural microbial restoration and therefore should be considered primary targets in bioremediation. If natural weathering appears to be effective in reducing both pH and EC, it is also a slow process that could be artificially accelerated. In the last few years, microbially-driven pH neutralization using organic acids and CO$_2$ produced by fermentation of added carbon sources has gained attention as a promising technique for in-situ bioremediation of BRDAs (Santini et al., 2021; Wu et al., 2019). Santini et al. (2021, 2016) proved that the efficiency of this process could be improved by decreasing the initial pH and salinity and increasing the biodiversity of the microbial inoculum. However, they also identified the shortage of nutrients (N and P) in the medium term as a limiting factor for the development of this methodology. Our results suggest that lowering EC increased biodiversity and facilitated the development of bacteria related to the N cycle. Hence, a first treatment focused on reducing salinity (and pH if possible) before bacteria-mediated pH neutralization could yield considerably better results. Few studies have investigated the effect of lowering artificially pH and EC in the optic of bioremediation of BRDAs (Courtney et al., 2014; Jones et al., 2011; Wong and Ho, 1994). As a recent example, the recent study by Fourrier et al. (2021) has reported the positive effect of gypsum addition and repeated washing on pH neutralization and EC decrease. In addition, fungal contribution to early-stage microbial assembly seems to be underrated and their role in bauxite residue remediation needs more research. Regarding microbially-driven strategies for metal bio-extraction, our study is particularly informative for selecting appropriate bauxite residue and bacteria for the selective extraction of critical metals. In agreement with literature (Rivera et al., 2019; Vind et al., 2018), our results indicated that the concentration of REEs (Ln, Sc and Y) in bauxite residue is primarily determined by the ore origin, with karst bauxite (Pr90 and Pr100) being more enriched in REEs than laterite bauxite (Fresh, Bo1 and Bo2). In addition, Y and Ln were strongly
correlated as expected regarding the analogy between Y and heavy REEs while Sc behaves differently to Y and Ln and is likely incorporated in iron oxyhydroxide such as goethite (Levard et al., 2018). The conventional methods to recover these metals are not environmentally acceptable due to their excessive energy consumption and the production of hazardous residues, that require further treatment and high operating costs (Baniasadi et al., 2019). Metal bio-extraction is an interesting alternative to traditional metallurgy that has still been poorly investigated for critical metals. Among the ideas that could benefit from our study, alkaline-active exoenzymes produced by bacteria found in Bo1 and Bo2 offer interesting possibilities for the valorization of bauxite residue. Despite their ability to lower their intracellular pH and salinity, haloalkaliphilic species must still be able to secrete stable and operational enzymes at elevated pH and salinity (Mamo, 2020), making BRDAs potential sources of these enzymes. Some of the haloalkaliphilic bacteria known to synthesize alkaline-active enzymes, namely Gammaproteobacteria, Actinobacteria and Firmicutes (Kalwasińska et al., 2018; Litchfield, 2011; Maharaja et al., 2018; Shameer, 2016; Shivlata and Tulasi, 2015), are also capable to modify the chemical phase and mobility of various critical metals (Chidambaram et al., 2010; Presentato et al., 2020; Wee et al., 2014), including REEs. For example, Maleke et al. (2019) reported the direct reduction of Eu$^{3+}$ to Eu$^{2+}$ by Clostridium sp. (Firmicutes). In line with previous studies (Santini et al., 2015b), the majority of Gammaproteobacteria corresponded to uncharacterized lineages, highlighting the potential of this group as a phylogenetic hot spot for novel haloalkaliphilic taxa.

5. CONCLUSIONS

In summary, this paper presents an integrated physicochemical and biological approach to explore the composition and dynamics of bacterial and fungal communities in bauxite residues deposited at different times and produced from different bauxite ores. Our results revealed that both deposit age and ore origin affect the geochemistry of bauxite residue, although unequally. Salinity, pH, TOC, TN, and P values seem to depend predominantly on the natural aging of bauxite residue, while the content of REE is mainly influenced by the origin of the bauxite ore. Our results highlight the behavior of bauxite residue deposits as primary successional environments and bring new insights into the early stages of microbial community
assembly in these sites. The pioneer microbial community was dominated by haloalkaliphilic microorganisms, strongly influenced by chemical gradients. Autotrophic and heterotrophic microbial species contribute to the supply of nutrients necessary for the development of other species through C fixation and degradation of allochthonous organic compounds, respectively. After the first years of natural restoration, nitrogen-fixing bacteria increase their presence and contribute to enhance N bioavailability. Microbial richness, diversity and network complexity increased significantly with age of deposition, until primary succession equilibrium was reached decades later, characterized by a microbial community composition similar to that of typical soils and freshwater. Our results suggested that salinity, pH, nutrients, and toxic metals (mainly Cr and V) were the main factors explaining this change in microbial communities. These results confirm the key role of pH and salinity in the establishment of early microbial communities and highlight them as main targets for bioremediation. Moreover, our co-occurrence network data suggest an important role of fungal communities in structuring the early-stage microbial community during primary succession, potentially by recycling complex organic matter. We also identified bacteria with potential metal extraction abilities, such as secretion of alkaline-active enzymes that could modify the chemical phase of the metals present.


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