

Contrasted microbial community colonization of a bauxite residue deposit marked by a complex geochemical context

Luis Alberto Macías-Pérez, Clément Levard, Mohamed Barakat, Bernard Angeletti, Daniel Borschneck, Laurent Poizat, Wafa Achouak, Melanie Auffan

▶ To cite this version:

Luis Alberto Macías-Pérez, Clément Levard, Mohamed Barakat, Bernard Angeletti, Daniel Borschneck, et al.. Contrasted microbial community colonization of a bauxite residue deposit marked by a complex geochemical context. Journal of Hazardous Materials, 2022, 424, Part B, pp.127470. 10.1016/j.jhazmat.2021.127470. hal-03468413

HAL Id: hal-03468413 https://cnrs.hal.science/hal-03468413v1

Submitted on 16 Dec 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

- 1 "Contrasted Microbial community colonization of a bauxite residue deposit
- 2 marked by a complex geochemical context."
- 3 Luis Alberto Macías-Pérez ^{a,b}, Clément Levard ^a, Mohamed Barakat ^b, Bernard
- 4 Angeletti^a, Daniel Borschneck^a, Laurent Poizat^d, Wafa Achouak^{b*}, Mélanie Auffan^{a,c*}
- 5 *These authors contributed equally to the study
- 6 Author contact information (FAMILY NAME, Name):
- 7 1) MACÍAS-PÉREZ, Luis Alberto ^{a,b} (Corresponding author)
- 8 Address: CEREGE, Technopôle de l'Arbois-Méditerranée, BP80, 13545 Aix-en-
- 9 Provence, France
- 10 Email: <u>macias@cerege.fr</u>
- 11 Telephone number: +33 766887887 (France)
- 12 Fax: +33 (0)4 13 94 91 30 (France)
- 13 2) LEVARD, Clément ^a
 14 Email: levard@cerege.fr
- 15 3) BARAKAT, Mohamed ^b
- 16 Email: mohamed.barakat@cea.fr
- 17 4) ANGELETTI, Bernard ^a
- 18 Email: <u>angeletti@cerege.fr</u>
- 19 5) BORSCHNECK, Daniel ^a
- 20 Email: <u>borschneck@cerege.fr</u>
- 21 6) POIZAT, Laurent ^d
- 22 Email: <u>laurent.poizat@alteo-alumina.com</u>
- 23 7) ACHOUAK, Wafa ^b
- 24 Email: wafa.achouak@cea.fr
- 25 8) AUFFAN, Mélanie ^{a,c}
- 26 Email: <u>auffan@cerege.fr</u>
- 27

28 Author affiliations:

- a. Aix Marseille Université, CNRS, IRD, INRAE, Collège de France, CEREGE,
 13545, Aix-en-Provence, France
- b. Aix Marseille Univ, CEA, CNRS, BIAM, LEMIRE, Laboratory of Microbial Ecology
 of the Rhizosphere, ECCOREV FR 3098, F-13108, St-Paul-lez-Durance, France
- 33 c. Civil and Environmental Engineering, Duke University, Durham, NC 27708, USA
- d. ALTEO Alumina, 13120, Gardanne, France
- 35
- 36 Keywords: Bauxite residue, Primary succession, Microbial communities,
- 37 Physicochemical characterization, Critical metals.
- 38

39 Abstract

Bauxite residue is the alkaline byproduct generated during alumina extraction and is commonly landfilled in open-air deposits. The growth in global alumina production have raised environmental concerns about these deposits since no large-scale reuses exist to date. Microbial-driven techniques including bioremediation and critical metal bio-recovery are now considered sustainable and cost-effective methods to revalorize bauxite residues. However, the establishment of microbial communities and their active role in these strategies are still poorly understood. We thus determined the geochemical composition of different bauxite residues produced in southern France and explored the development of bacterial and fungal communities using Illumina high-throughput sequencing. Physicochemical parameters were influenced differently by the deposit age and the bauxite origin. Taxonomical analysis revealed an early-stage microbial community dominated by haloalkaliphilic microorganisms and strongly influenced by chemical gradients. Microbial richness, diversity and network complexity increased significantly with the deposit age, reaching an equilibrium community composition similar to typical soils after decades of natural weathering. Our results suggested that salinity, pH, and toxic metals affected the bacterial community structure, while fungal community composition showed no clear correlations with chemical variations.

69 **1. INTRODUCTION**

70 Bauxite residue is a solid by-product generated during the production of alumina from 71 bauxite. Alumina extraction is most often performed by the Bayer process, where 72 bauxite is digested with large quantities of sodium hydroxide at temperatures between 73 150 and 250 °C (Evans, 2016). Depending on the process parameters and the origin 74 of the bauxite, 0.7 to 2 tons of bauxite residues are produced per ton of alumina (International Aluminium Institute and European Aluminium, 2015). Over the last 10 75 76 years, the average annual production of alumina was 116 ± 15 Mt (World Aluminium, 77 2020), which corresponds to ~200 Mt of bauxite residue produced per year. In the 78 absence of economically profitable large-scale applications, bauxite residues are 79 commonly landfilled in large open-air deposit areas (BRDA), reaching a total of 4.5 Bt 80 bauxite residue in storage at the present time (Dentoni et al., 2021).

81 Due to the physicochemical characteristics of bauxite residue i.e., high alkalinity, high 82 salinity, high metal content and lack of nutrients, BRDAs pose a challenge to most 83 living organisms (Di Carlo et al., 2020; Santini et al., 2015b) and represent a potential 84 source of contamination for the surrounding terrestrial and aquatic ecosystems 85 (Bouchoucha et al., 2019; Ren et al., 2018). Consequently, efforts have been made to 86 remediate these deposits, usually by the addition of amendments (e.g. gypsum and 87 organic matter) that attenuate their harsh conditions prior to the revegetation of the area (Bray et al., 2018; Khaitan et al., 2010). BRDAs have also been considered as a 88 89 promising secondary source of valuable and critical metals, as these elements are 90 concentrated by a factor of 2 in bauxite residues compared to the initial bauxite ore 91 (Panda et al., 2021; Ujaczki et al., 2017; Vind et al., 2018). Critical metals are chemical elements characterized by their high economic relevance and supply risk (European 92 93 Commission, 2020) and include metals crucial for information and energy technologies 94 such as Co, Mg, Ba, V, Ge, Nb, Sr, Ga, platinoids (PGMs), and rare earth elements 95 (REEs: Lanthanides (Ln), Y, Sc).

96 Recently, microbial-driven approaches have gained attention as viable and cost-97 effective methods for the management and valorization of urban and industrial wastes, 98 including bauxite residue (Lyu et al., 2021; Panda et al., 2021; Santini et al., 2019). 99 During in-situ remediation processes, native microbial communities have been shown 100 to decrease the pH and salinity of the bauxite residue, as well as play an essential role 101 in soil formation and plant growth (Di Carlo et al., 2019; Tian et al., 2020; Wu et al., 102 2019). Microbial metal recovery strategies are also promising for the dissolution and 103 recovery of elements of interest (e.g., Fe, AI, critical metals) from waste (Baniasadi et 104 al., 2019; Dev et al., 2020; Maes et al., 2016). Out of all these techniques, bioleaching 105 and microbial electrochemistry are among the most-studied methods (Dominguez-106 Benetton et al., 2018; Srichandan et al., 2019). In bauxite residues, these techniques 107 are still at an early experimentation stage, although some promising results for REE 108 recovery by bioleaching have been reported (Kiskira et al., 2021; Qu et al., 2019; 109 Zhang et al., 2020).

110 More generally, microbial-driven strategies for waste management imply the use of 111 microbial communities native to contaminated sites, which tend to thrive better in such 112 harsh environments due to their unique metabolisms developed through natural 113 selection (Ghosh et al., 2018; Ma et al., 2019; Roy et al., 2018; Sajjad et al., 2020). In 114 addition, microbial communities are known to perform complex functions and are more 115 robust to environmental fluctuations compared to pure cultures (Perez-Garcia et al., 116 2016; Wang et al., 2020; Zhang et al., 2008). Metagenomics studies based on Next Generation Sequencing have significantly expanded the identification and dynamics 117 of microbial communities involved in different bioprocesses. However, the evolution of 118 119 microbial communities in metal biorecovery experiments is still limited to a few weeks 120 (Ma et al., 2017; Sajjad et al., 2020; Wang et al., 2020). Regarding bauxite residues, 121 the identification of pioneer microbial communities and their dynamics during primary 122 succession have been highlighted as research needs to improve microbial-driven 123 bioremediation (Santini et al., 2015a). Nevertheless, to date, studies on microbial 124 diversity in bauxite residue have focused on the responses of bacterial communities 125 to remediation strategies, overlooking their active role in the process (Banning et al., 2011; Fourrier et al., 2020; Ke, 2021; Krishna et al., 2014; Schmalenberger, 2013; Wu 126 127 et al., 2020).

This study aimed at exploring the microbial dynamics during primary succession in a BRDA from Southern France using an integrated physicochemical and biological approach. The specific objectives of this study were to (a) evaluate the effect of the deposit age and the ore origin on the geochemical characteristics of bauxite residue; (b) identify the first microbial communities colonizing BRDAs and their role in the establishment of new species; (c) determine the microbial community structure in the equilibrium stage of primary succession in BRDAs and the main geochemical factors driving it; (d) explore possible implications of this study for microbial-drivenbioremediation and critical metal recovery.

137

138 2. MATERIALS AND METHODS

139 2.1 Site description and sampling

140 The bauxite residues come from a refinery operating the Bayer aluminum extraction 141 process since 1894 in Gardanne, Southern France. The sampling took place at the 142 Mange-Garri bauxite residue disposal area (BRDA) in Bouc-Bel-Air, Southern France. 143 This region is characterized by a hot-summer Mediterranean climate, with average 144 annual precipitation and temperature of 485 mm and 16.3°C (Copernicus Climate Data 145 Store, 2021). We selected four distinct areas of the BRDA based on their deposit age 146 (1, 2, 90, and 100 years) and bauxite origin (lateritic bauxite from Boké, Guinea (Bo) 147 and karstic bauxite from Provence, France (Pr)), named Bo1, Bo2, Pr90 and Pr100 148 (Fig. 1). At the Bo1 and Bo2 sites, bauxite residues were produced from Guinean 149 bauxite and were landfilled one and two years before sampling respectively. These deposits are occasionally watered to prevent dust dispersion and were not amended 150 151 at the time of sampling. The Pr90 and Pr100 sites contain bauxite residues that were 152 produced from Provençal bauxite and were deposited around 90 and 100 years ago 153 respectively. These sites were amended with a soil layer in the 1960s and a slight 154 coverage with low-lying vegetation can be seen.

155 At each of the four sites, samples were collected in triplicate, gathering approximately 156 500 g of bauxite residue at a depth of 20 to 30 cm in sterile plastic bags. At sites Pr90 157 and Pr100, samples were collected at the edge of the deposit, where bauxite residues 158 were distant from the added soil layer. In addition, freshly produced bauxite residue (4 159 replicates) was also selected to analyze the initial bio-geochemical characteristics 160 before landfill. Samples were divided into two groups based on the following analyses. Samples undergoing physicochemical analysis were oven-dried at 70°C, grounded in 161 162 a mortar, passed through a 130 µm sieve, and stored in metal-free plastic tubes. 163 Samples for microbial community analyses were stored at 4°C.





164

165 2.2 Physicochemical analyses

166 2.2.1 Elemental composition

167 First, pH and electrical conductivity (EC) were measured at a solid/MilliQ water ratio 168 of 1:5. Then, the elemental composition of the bauxite residues was determined after alkaline fusion of the samples (Rivera et al., 2019). Briefly, 1 g of each sample was 169 digested with 750 mg of lithium tetraborate (Li₂B₄O₇) at 1000°C for 30 min (WiseTherm 170 F/FH 0-1200°C) and then immediately dissolved in 40 mL of HNO₃ 1N. Prior to 171 elemental analysis, the samples were dissolved 200 times. Subsequently, inductively 172 coupled plasma mass spectrometry (ICP-MS, Perkin Elmer Nexlon 300X) was used 173 174 to analyze elements with concentrations up to 1000 µg/L, while elements with concentrations above 1000 µg/L were analyzed by inductively coupled plasma optical 175 emission spectroscopy (ICP-OES, Perkin Elmer 4300 DV). All analyses were 176 177 conducted for major and trace elements commonly found in bauxite residues, namely Fe, Ti, Al, Ca, Zr, Cr, Mn, Zn, Th, Ni, Pb, Cu, Sn, Rb, Cs as well as critical raw elements 178 179 (as defined by the European Commission (2020): P, Mg, V, Lanthanides (Ln), Sr, Y, Nb, Ba, Sc, Ga, Hf, Ge, Co, Sb, W, Ta). Total carbon (TC) and nitrogen were 180 181 determined by dry combustion using an Elemental Analyzer (Flash EA, Thermo 182 Scientific). To measure the total organic carbon (TOC), samples were treated with HCl 183 to remove carbonates before analysis.

184 2.2.2 Mineralogical composition

185 The mineralogical analyses of bauxite residue were carried out by X-ray diffraction 186 (XRD) using a PANalytical X'Pert Pro (Malvern Panalytical, UK) diffractometer 187 equipped with a cobalt tube ($\lambda = 1.79$ Å) running at 40kV and 40mA. Samples were

deposited on low background silicon plates and analyzed from 8° to 80° (20) with a 188 189 step size of 0.033° and a total counting time of 7 hours. Samples were also spun at 15 rpm to improve statistics. Phase identification was performed using the X'pert 190 191 Highscore plus software (PANalytical) together with the PDF-2 ICDD database 192 (International Center for Diffraction data, Powder Diffraction Files 2). Profex software 193 (Doebelin and Kleeberg, 2015) was used for Rietveld refinement to semi-quantify the 194 proportions of the minerals in the bauxite residues. The following parameters were 195 refined: zero point shift, sample displacement, cell parameters, preferred orientation 196 and peak broadening resulting from the size of the crystallite and the micro strain.

197 2.3 DNA extraction, PCR amplification and sequencing

198 The total DNA was extracted from 5 g of each bauxite residue samples using the 199 FastDNA® Spin Kit for soil (MP Biomedicals, USA) following the manufacturer's 200 protocol. Extracted DNA was used as template in separate PCR reactions amplifying 201 the bacterial 16S and the fungal ITS rRNA gene sequences. For bacterial diversity 202 analysis, we used the primers 341F (5'-CCTAYGGGRBGCASCAG-3') and 806R (5'-203 GGACTACNNGGGTATCTAAT-3'), targeting the V3 and V4 variable regions of the 204 bacterial 16S rRNA (Caporaso et al., 2011; Muyzer et al., 1993). For fungal diversity 205 analysis, we used the primers fITS7 (5'-GTGARTCATCGAATCTTTG-3') and ITS4 (5'-206 TCCTCCGCTTATTGATATGC-3'), targeting the ITS2 region (Ihrmark et al., 2012). 207 The amplification conditions were as follows: initial denaturation at 95 °C for 2 min; 208 followed by 34 cycles of denaturation at 95 °C for 30 s, annealing at 55 °C for 30 s, 209 and extension at 72 °C for 1 min; and a final extension at 72 °C for 5 min (Muller et 210 al., 2021). PCR products were then purified using ProNex® Size-Selective Purification System (Promega, USA) and sequenced on Illumina MiSeq platform (Biofidal, Vaulx-211 212 en-Velin, France). The raw sequence reads generated from this study have been 213 deposited in the National Center for Biotechnology Information (NCBI) Sequence 214 Read Archive under the accession number PRJNA748554.

215 2.4 Sequencing data processing

Microbiome bioinformatics were performed by the open-source software QIIME2, version 2019.10 (<u>https://qiime2.org</u>) (Bolyen et al., 2019). Raw reads were qualityfiltered, denoised and chimera-checked using DADA2 (Callahan et al., 2016). DADA2 uses a parametric model to infer true biological sequences from reads. The model relies on input read abundances (true reads are likely to be more abundant) and the pairwise similarity between sequences. The taxonomic annotation of the resulting sequence variants (ASVs) was assigned using the feature-classifier command with default parameters in QIIME2 and sequences were matched against the Greengenes 13_8 database (McDonald et al., 2012). Finally, scaling with ranked subsampling (SRS) curves (Beule and Karlovsky, 2020) were drawn to determine whether the sequencing depth was sufficient to represent the true diversity of the samples.

227 2.5 Biodiversity parameters and microbial biomarker discovery

228 Alpha diversity was explored through observed ASVs, Chao1, and Shannon. The 229 observed ASVs and the Chao1 estimator were selected to identify community 230 richness, and Shannon index was used to assess community diversity (Callahan et 231 al., 2017; Hill et al., 2003). Beta diversity analysis was used to evaluate distribution 232 patterns in samples based on bacterial and fungal ASV composition (Anderson et al., 233 2011; Callahan et al., 2017). For this purpose, a principal coordinate analysis (PCoA) 234 based on weighted UniFrac distances (Lozupone et al., 2007) was conducted. All 235 alpha and beta diversity metrics were calculated using QIIME 2 after normalization to 236 59388 and 4778 sequences per sample for bacteria and fungi, respectively. Linear 237 discriminant analysis effect sizes algorithm (LEfSe) (Segata et al., 2011) was 238 performed on the Galaxy platform (<u>https://huttenhower.sph.harvard.edu/galaxy/</u>) to 239 identify bacterial biomarkers characterizing the samples. LEfSe couples Kruskal-240 Wallis tests for measuring statistical significance with quantitative tests for biological consistency (Wilcoxon rank sum test). 241

242 2.6 Statistical analysis

243 All statistical analyses were performed with the open-source software R (R Core Team, 2020) using the packages "dplyr" (Wickham et al., 2021), "vegan" (Oksanen et 244 al., 2020), "car" (Fox and Weisberg, 2019), "ggpubr" (Kassambara, 2020), and "rstatix" 245 246 (Kassambara, 2021). Figures were produced with the package "ggplot2" (Wickham, 247 2009). To study the ore-dependent differences in chemical and microbiological 248 compositions of bauxite residue, unpaired two-sided T-tests and Wilcoxon rank sum 249 tests were conducted respectively. Furthermore, Games-Howell post-hoc tests were 250 used to assess the age-dependent variations in chemical and microbial compositions. Pearson correlations between each variable and the deposit age were used to further 251

explore these relationships. Also, significant differences in alpha diversity indices weretested by Wilcoxon rank sum tests.

254 Principal component analysis (PCA) was used to identify the variables that explain 255 most of the variation in chemical composition. In addition, unsupervised hierarchical 256 clustering was applied to the PCA to group the samples according to their chemical 257 similarity. Both PCA and hierarchical clustering were performed using the packages 258 "FactoMineR" (Lê et al., 2008) and "factoextra" (Kassambara and Mundt, 2020). 259 PERMANOVA via the adonis function was conducted in both PCA and PCoA to test for the chemical and microbiological dissimilarities based on the deposit age and the 260 261 bauxite origin. To study the factor age alone, nested PERMANOVA were calculated using the parameter strata to exclude the effect of bauxite origin. Furthermore, multiple 262 263 co-inertia analysis (MCIA) was performed using the package "omicade4" (Meng et al., 264 2014) to determine the relationships between the four datasets used in this study 265 (chemical characteristics, mineralogy, 16S rRNA sequences and ITS sequences).

266 2.7 Co-occurrence network construction

Co-occurrence analyses were implemented for a better understanding of bacterial and 267 268 fungal interactions in the four bauxite residues. Co-occurrence networks were constructed based on pairwise Pearson correlations calculated between bacterial and 269 270 fungal ASVs by using the base R function cor (Berry and Widder, 2014; Williams et 271 al., 2014). To avoid including false positives in the network due to spurious or random 272 interactions, the ASV table we permuted 100 times and a *p*-value for each possible 273 pairwise interaction was calculated to test its validity. The *p*-values were then adjusted 274 using the Benjamini-Hochberg procedure (Benjamini and Hochberg, 1995), and only 275 edges with a *p*-value below 0.01 that corresponded to an absolute correlation higher than 0.4 were retained. To describe the topology properties of the networks, a set of 276 277 network indexes including graph density, average degree of nodes, transitivity, 278 modularity, average geodesic distance, betweenness centrality, and density were 279 calculated with the R package "igraph" (Csardi and Nepusz, 2006). The Wilcoxon test 280 was employed to assess significant differences in topological parameters between 281 networks. The connectivity of the network nodes was determined by their within-282 module connectivity (Zi) and among-module connectivity (Pi). Nodes were then 283 classified into four categories, according to Poudel et al. (2016): peripherals (Zi < 2.5 284 and Pi < 0.62, nodes with few links to other species), connectors (Pi > 0.62, nodes that connect modules), module hubs (Zi > 2.5, highly connected nodes within
modules), and network hubs (Zi > 2.5 and Pi > 0.62, highly connected nodes among
and within modules).

288

289 **3. RESULTS**

3.1.pH, EC and geochemical composition

The pH of the bauxite residue ranged from 8.70 ± 0.34 to 11.90 ± 0.01 , with the highest value in fresh samples and the lowest in Pr100 (**Fig. 2a**, Table S1). The EC of the bauxite residue ranged from 0.29 ± 0.15 mS/cm to 2.53 ± 0.04 mS/cm, and again, fresh samples showed the highest value and Pr100 the lowest (**Fig. 2a**, Table S1).

295 The elements with the highest concentrations in all bauxite residue samples were Fe (268894 ± 45706 mg/kg), Ti (51313 ± 17317 mg/kg), Al (33093 ± 5662 mg/kg), Ca 296 (22425 ± 4952 mg/kg), and TC (17162 ± 9439 mg/kg) (Fig. 2b, Table S2). Fe 297 298 represents at least 22 ± 3 % of the dried mass in all samples. The main critical 299 elements present in all samples were Ti (51313 ± 17317 mg/kg), P (2743 ± 1274 300 mg/kg), Mg (1877 ± 987 mg/kg), V (849 ± 423 mg/kg), Ln (775 ± 339 mg/kg), Sr (240 301 ± 142 mg/kg), Y (159 ± 59 mg/kg), Nb (99 ± 16 mg/kg), Sc (73 ± 12 mg/kg), Ba (73 ± 302 15 mg/kg), and Ga (56 ± 24 mg/kg) (Fig. 2c-d, Table S2). TN concentrations were 303 only above the machine's limit of detection (100 mg/kg) in the Provence samples, with 304 values of 287 \pm 183 mg/kg for Pr90 and 470 \pm 158 mg/kg for Pr100 (Table S2).

Statistical analysis showed that Ti, Cr, V, and Ga concentrations were significantly 305 306 higher in samples coming from Boké bauxite (p < 0.05, Table S2), while samples from Provence bauxite showed greater abundances of TC, TOC, P, Mg, Ln, Mn, Sr, Y, Sc, 307 Pb, Ni, and Co (p < 0.05, Table S2). Within these two groups, some elements 308 309 concentrations changed over time. In the Boké samples, a significant decrease in Fe, Ca, Cr, V, Ga, Cu, Ge, and W was observed as sample age increased (*p* < 0.05, Table 310 311 S2) while Mg, and Sr concentrations increased with age (p < 0.05, Table S2). In the Provence samples, Mg decreased significantly with age, while P and TOC 312 313 concentrations increased (p < 0.05, Table S2). These relationships between the age 314 of the sample and their chemical properties were further confirmed by Pearson 315 correlations ($R \ge 0.9$, p < 0.05, Fig. S1).



Figure 2. Physicochemical properties of bauxite residue samples. A) pH and EC, B) Major elements concentrations, C) Trace elements concentrations, and D) Mineralogy. Critical elements are colored red.

316

317 **3.1.2 Element correlation**

- 318 PCA was used to detect patterns in the chemical dataset and describe linear relations
- between the analyzed physicochemical parameters (**Fig. 3a**). The first three principal
- 320 component axis (PC1, PC2, and PC3) explained 84.3% of the data variability between
- 321 samples. PC1 explained 53.0% of the variance and included two groups: the first one

322 composed of pH, EC, Ga, Cr, V, Cu, Fe, Ge, Ca, Ti, and W whereas the second one comprised TOC, Co, Ni, Sr, Ln, Mg, Pb, Mn, Y, Sb, Cs, and Sn. PC2 explained 18.4% 323 324 of the variance among samples and was mainly composed of Zr, Sc, Hf, Ta, Al, Zn, 325 Nb, and Ti. Finally, PC3 (not shown) explained 12.9% of variance and was primarily 326 composed of P, Th, Rb, Zn, Cs, and Sn. PERMANOVA analysis corroborated that the 327 factors "age" (p < 0.01) and "bauxite origin" (p < 0.05) significantly explained the 328 differences in chemical properties across samples. The factor "age" (tested by nested PERMANOVA to exclude the effect of the bauxite origin) explained 65.8% of the 329 330 variance in chemical parameters while the factor "bauxite origin" explained 28.18%. 331 The hierarchical cluster analysis (Fig. S2) provided statistical confirmation of the 332 tendencies observed in the PCA and bauxite residue samples were clustered based 333 on their chemical similarity. The first cluster contained the fresh samples, the second 334 cluster Bo1 samples, the third cluster Bo2 samples, and the fourth cluster contained 335 both Pr90 and Pr100 samples.

336 To further understand the effect of the factor "age" on residues generated from the 337 same bauxite ore, a second PCA was conducted considering only the Boké samples 338 (fresh, Bo1 and Bo2) (Fig. 3b). The first two dimensions explained 82.5% of the 339 variability and the factor "age" separated the samples significantly (PERMANOVA, p 340 < 0.01). Co, Sr, Pb, Mg, and Sb were mostly associated with Bo2, while pH, EC, W, 341 Ca, Ge, Fe, Cu, V, Cr, and Ga had their highest values in fresh samples and decreased 342 progressively with age. In contrast, TOC, Y, Mn, Ln, Ni, Sn, Al, and Ta did not show 343 any age-dependent trends during the first years of storage (Table S2). Globaly, this 344 statistical analysis is in line with the expected behavior of the elements, such as the colocalization of Cr and V and their correlation with Fe (Markus Gräfe et al., 2011) as 345 well as the REEs affinity for Mn phases (Vind et al., 2018). 346



Figure 3. Principal component analysis (PCA) of chemical parameters in A) all and B) Boké bauxite residue samples. Individuals are represented by dots and colored by sample. Individuals are grouped according to hierarchical clustering results. Variables are represented by arrows and colored by chemical nature. Only variables with a $cos^2 > 0.5$ are shown.

349 **3.1.3 Mineralogical composition**

350 XRD analysis (Fig. 2d, Table S3) showed that bauxite residues were principally 351 composed of iron and aluminum oxides, hydroxides, and silicates. Goethite 352 (FeO(OH)), hematite $(Fe_2O_3),$ gibbsite and vishnevite $(AI(OH)_3),$ (Na₈(AlSiO₄)₆O₂₄(SO₄)) are the dominant phases and accounted for at least 70 % of 353 354 the crystalline fraction in all the samples. Rutile (TiO₂) and calcite (CaCO₃) were also 355 detected in all samples, in the range of 3-5 % and 2-14% respectively. The hematite, 356 anatase (TiO₂) and boehmite (AlO(OH)) were significantly more present in Provence 357 samples, and interestingly, neither anatase nor boehmite were found in fresh samples 358 (Table S3). In Boké samples, a significant decrease in goethite and Katoite 359 $(Ca_3Al_2(SiO_4)_{1.5}(OH)_6)$ was observed as the age of the samples increased.

360 3.2 Microbial community diversity, structure and dynamics

361 **3.2.1 Alpha diversity**

362 The influence of bauxite residue's age and geochemical composition on microbial 363 communities was assessed using 16S and ITS metabarcoding. As expected, no DNA could be extracted from the fresh residue samples. For Bo1, Bo2, Pr90, and Pr100, 364 365 the sequencing achieved a coverage of more than 55000 16S rRNA sequences and 366 4700 ITS sequences per sample. The SRS curves (Fig. S3) indicated that the 367 sequencing depth was sufficient to identify the majority of ASVs within bacterial and 368 fungal communities in all the samples (except fresh samples). Alpha-diversity derived 369 from the number of ASVs, Chao1 richness, and Shannon's index showed differences between samples regarding bacterial and fungal communities (Table S4). A total 370 371 number of 7994 bacterial ASVs were identified, with 350 to 1318 ASVs per 372 sample. The number of bacterial ASVs in the Provence samples (Pr90 and Pr100) 373 was 6928, significantly greater than the 2867 ASVs found in the Boké samples (Bo1 374 and Bo2). The Provence samples also showed the highest richness and Shannon 375 values for bacteria, compared with Boké samples (Fig. 4a). For fungi, 892 ASVs 376 ranging from 45 to 156 ASVs per sample were obtained, with again more ASVs and 377 Chao1 richness in Provence samples. However, there were no significant changes in 378 Shannon's index between both group of samples, indicating that the specific diversity 379 of the fungal communities was similar in Provence and Boké samples.



Figure 4. Bacterial alpha and beta diversity of bauxite residues. A) Chao1 richness and Shannon diversity index. B) Principal coordinate analysis (PCoA) based on weighted UniFrac distances between bauxite residue samples.

381 3.2.2 Microbial community structure

382 PCoA ordinations based on weighted UniFrac distances (Fig. 4b) revealed that the 383 bacterial community structure differed between the bauxite residue samples. Samples 384 were significantly separated based on bauxite origin (PERMANOVA, p < 0.05), and nested PERMANOVA confirmed the difference in bacterial community structure 385 386 across age (p < 0.001). The first two axes of PCoA explained 74.95% of the community 387 dissimilarity. Axis 1 was most correlated with EC (R = 0.97, p = 1.1e-07), followed by Co (*R* = 0.96, *p* = 9.8e-07), Ni (*R* = 0.95, *p* = 1.4e-06), Cr (*R* = 0.94, *p* = 1.1e-07), V 388 (R = 0.92, p = 2.5e-05), Ti (R = 0.9, p = 5.4e-05), and pH (R = 0.9, p = 8.1e-05); while 389 Axis 2 correlated with P (R = 0.75, p = 0.005). In contrast, the fungal community 390 391 structure did not seem to follow a clear pattern among samples according to PCoA 392 built on Bray-Curtis distances (Fig. S4).

393 The bacterial phyla that dominated all bauxite residue samples were Proteobacteria, 394 Actinobacteria, Chloroflexi, Bacteroidetes, Gemmatimonaidetes, Planctomycetes, 395 Verrucomicrobia, Acidobacteria and Firmicutes, accounting for more than 90% of the 396 total bacterial communities (Fig. 5). When comparing the samples from Provence and Boké, significant taxonomic differences were observed. Actinobacteria and 397 398 Gammaproteobacteria were the dominant taxa in the Boké samples with an average 399 abundance of 28.0 % and 21.3 % respectively, followed by Chloroflexi (9.8 %) and 400 Bacteroidetes (9.8 %). In the Provence samples, Actinobacteria (18.4 %) were still 401 dominant while Gammaproteobacteria (4.8 %) and Bacteroidetes (3.6 %) lost 402 prominence in favor of Alphaproteobacteria (12.0 %), Planctomycetes (9.6 %), Betaproteobacteria (6.3 %), and Deltaproteobacteria (3.2 %). Boké samples were 403 404 significantly enriched in Actinobacteria, Gammaproteobacteria, Bacteroidetes and Firmicutes, whereas samples from Provence showed greater abundances of 405 Alphaproteobacteria, Planctomycetes, Betaproteobacteria and Deltaproteobacteria (p 406 407 < 0.05). Acidobacteria (10.3 %), barely found in Boké samples, showed a strong 408 increase in Provence samples (p < 0.05). Chloroflexi, Gemmatimonaidetes and 409 Verrucomicrobia were nearly constant across samples.



Figure 5. Relative abundance of the most abundant bacterial taxa among bauxite residue samples. Proteobacteria are divided in classes.

410

411 At the level of bacterial classes, differences between Boké and Provence samples 412 were also found (Fig. 6). Significantly higher proportions of Gammaproteobacteria, 413 Nitriliruptoria, Thermomicrobia, Gemm3, Bacilli, Rhodothermi, TK17, Deinococci and Clostridia were identified in the Boké samples (p < 0.05). Within this group, variations 414 415 between samples of different ages were also observed. Bo1 was enriched in Thermomicrobia, Thermoleophilia, Acidimicrobiia and Bacilli, whereas Opitutae, 416 417 Verrucomicrobiae and Flavobacteriia showed greater abundances in Bo2 (p < 0.05). significantly 418 contrast, Provence samples were more enriched In in 419 Acidobacteria6. Alphaproteobacteria, Planctomycetia, Betaproteobacteria, 420 Thermoleophilia, Deltaproteobacteria, Gemmatimonadetes. Pedosphaerae. 421 Anaerolineae, Phycisphaerae, Spartobacteria, Gemm1, Chloracidobacteria, 422 Ellin6529, S085, Nitrospira, Saprospirae, TK10, Solibacteres and TM71 (p < 0.05). In 423 the Provence samples, the only differences seen over ages were Pedosphaerae and Actinobacteria, which were more abundant in Pr90 than in Pr100 (p < 0.05). 424



Figure 6. Z-score hierarchical clustering heat map representing the relative abundance of the top 40 bacterial classes in different bauxite residue samples.

425

426 **3.2.3 Taxonomic characteristics of bacterial and fungal communities**

427 To further explore these temporal scale differences, LEfSe analyses were conducted 428 to detect asymmetrically distributed bacterial taxa. Among the Boké samples, 9 429 indicator bacterial taxa were found in Bo1 against 37 in Bo2 (Fig. 7a). LEfSe confirmed 430 the highest abundance of three families of Bacilli in Bo1, as well as other taxa belonging to the phyla Actinobacteria and Proteobacteria, and the archaeal class 431 Halobacteria. In line with our previous results, Opitutae, Verrucomicrobiae and 432 433 Flavobacteriia were significantly more abundant in Bo2, together with other taxa 434 belonging to the phyla Actinobacteria, Proteobacteria, Firmicutes, Bacteroidetes, 435 Planctomycetes and Nitrospirae. For the Provence samples, only two indicator

- 436 bacterial taxa (classified as Bacteroidetes) were identified in Pr90, against 28 in Pr100
- 437 (classified as Actinobacteria, Proteobacteria, Planctomycetes, Bacteroidetes,
- 438 Verrucomicrobia, and Chloroflexi) (**Fig. 7b**).



Figure 7. LEfSe analysis of in the bauxite residue samples according to sample age. Histograms of LDA scores of 16S gene sequences in Boké (A) and Provence (C) samples. Only taxa with a LDA score (log10) above 2.0 and a p-value lower than 0.05 for Kruskal–Wallis tests are shown. B and D Cladograms are derived from LEfSe analysis. The central point denotes the root of the tree and expanded to each ring representing the next lower taxonomic level from phylum to genus.

440 **3.2.4 Microbial co-occurrence networks**

441 Co-occurrence networks were constructed using the ASV data of 16S rRNA and ITS 442 sequences. The Boké (Fig. 8) and Provence (Fig. 9) samples did not differ significantly 443 in network topological parameters such as average geodesic distance and modularity (Table S5). However, the network complexity (number of nodes and edges, degree of 444 445 nodes, and transitivity) was different between the Boké and Provence networks (Table 446 S5). The two Boké networks (Bo1 and Bo2) contained a similar number of nodes (366 447 and 369), although Bo1 showed more edges (516 versus 324), higher average degree (2.46 versus 1.76), and transitivity (0.15 versus 0.11). The two Provence networks 448 were more complex, as both Pr90 and Pr100 showed more nodes (603 and 571), more 449 450 edges (1254 and 1322), higher average degree (4.16 and 4.63), and higher transitivity 451 (0.18 and 0.12) compared to Boké.

452 Most of the nodes from the Boké and Provence networks were classified as 453 peripherals (specialists), and few nodes fell into module hubs (generalists) (Fig. 8-9). 454 Generalists are considered keystone taxa, as they are responsible for structuring the 455 different nodes and modules into a complete community, thus determining the 456 efficiency of energy metabolism and nutrient cycling in habitats (Wang et al., 2019). In 457 this study, more module hubs (generalists) were identified in Pr90 and Pr100 than in 458 Bo1 and Bo2. In Bo1 and Bo2 networks, the module hubs were mainly members of 459 Bacteroidetes, Actinobacteria, Proteobacteria, and Firmicutes for bacteria, and 460 Ascomycota for fungi. In contrast, the module hubs identified in Pr90 and Pr100 networks belonged to Proteobacteria, Actinobacteria, Chloroflexi, Verrucomicrobia, 461 462 Gemmatimonaidetes, Planctomycetes and Firmicutes for bacteria, and Ascomycota 463 and Basidiomycota for fungi.



Figure 8. Co-occurrence network and Zi-Pi plot of Bo1 (A) an Bo2 (B) bauxite residue samples. Each node represents a bacterial class (circles) or a fungal phyla (triangles). Nodes are colored by taxonomical affiliation and their size in the network is proportional to the node betweenness. The color of each link reflects positive (red) or negative (blue) interactions. The topological role of each node is defined by within-module connectivity (Zi) and among-module connectivity (Pi). According to values of Zi (2.5) and Pi (0.62), the roles of nodes are classified into four categories: Peripherals, Module hubs, Connectors and Network hubs.



Figure 9. Co-occurrence network and Zi-Pi plot of Pr90 (A) an Pr100 (B) bauxite residue samples. Each node represents a bacterial class (circles) or a fungal phyla (triangles). Nodes are colored by taxonomical affiliation and their size in the network is proportional to the node betweenness. The color of each link reflects positive (red) or negative (blue) interactions. The topological role of each node is defined by within-module connectivity (Zi) and among-module connectivity (Pi). According to values of Zi (2.5) and Pi (0.62), the roles of nodes are classified into four categories: Peripherals, Module hubs, Connectors and Network hubs.

469 **3.2.5 Multi co-inertia analysis**

Multiple co-inertia analysis (MCIA) was used to determine the relationships between 470 471 the chemistry, mineralogy, bacterial and fungal community composition in our bauxite 472 residue samples. Figure 10 displays the projection of the four datasets onto the first two principal components (PCs) of MCIA for Bo1, Bo2, Pr90 and Pr100 samples. PC1 473 474 and PC2 explained 54% and 19% of the total variation respectively (Fig. S6b). The 475 sample space (Fig. S6a) shows a good correlation between the different datasets in 476 all samples, which could be clearly differentiated. Early-stage bauxite residues (Bo1 477 and Bo2) were separated from equilibrium-stage samples (Pr90 and Pr100) along the 478 PC1. Pseudo-eigenvalues (Fig. S6c) indicated that the chemistry and bacterial dataset 479 contributed the most to the variation observed in PC1, while mineralogy was weighted 480 in PC2. The pair-wise RV coefficient, which is multivariate generalization of the 481 squared Pearson correlation coefficient, indicated higher global similarity between 482 chemistry and bacterial classes (RV score = 0.86) when compared to the similarities 483 between the chemistry and fungal classes (RV score = 0.36), and between the 484 mineralogy and microbial classes (RV score for bacterial classes = 0.41, RV score for 485 fungal classes = 0.22). Fungal classes contributed little to the variation of PC1 or PC2 486 and were mostly correlated with bacterial classes (RV score = 0.38).



O Chemistry ▲ Mineralogy ♦ Bacterial class ■ Fungal class

Figure 10. MCIA analysis of bauxite residue samples. The figure shows the coordinates of the different data sets represented by shapes (circle: chemical data set; triangle: mineralogical data set; square: fungal classes data set; diamond: bacterial classes data set). Only bacterial and fungal classes representing at least 5% of the abundance in at least one sample are shown. Uncl. = unclassified; Alpha = Alphaproteobacteria; Gamma = Gammaproteobacteria.

487 488

489 **4. DISCUSSION**

490 **4.1 BRDAs behave as primary successional environments**

491 The initial geochemical conditions of the fresh bauxite residue, characterized by high 492 alkalinity (pH = 11.90 \pm 0.01) and salinity (EC = 2.53 \pm 0.04 mS cm⁻¹) in line with previous studies (Cusack et al., 2019; Xue et al., 2016), make BRDAs hostile 493 494 environments for most domains of life. The high pH and salinity derive from the excess 495 of alkaline anions (hydroxides, carbonates, aluminates, aluminum hydroxides, 496 silicates) and Na⁺ produced during the Bayer process (Cusack et al., 2019; X. Kong 497 et al., 2017) and thus the bauxite origin does not seem to significantly influence the 498 value of these parameters. Fresh and recently deposited bauxite residues (Bo1 and

499 Bo2) also contained significant amounts of toxic metals for microorganisms, namely 500 Cr (1594 \pm 23 mg/kg), V (1452 \pm 10 mg/kg) and Al (38420 \pm 920 mg/kg). The high 501 alkalinity (pH > 10) of fresh, Bo1 and Bo2 bauxite residues may favor the Cr (III) 502 oxidation to Cr(IV) as well as the predominance of V(V), both highly mobile and toxic 503 for bacteria, fungi and plants (Chen et al., 2021; Economou-Eliopoulos et al., 2016; 504 Liang et al., 2021; Milačič et al., 2012; Xiao et al., 2017). Regarding aluminum, only 505 1% of the total AI content in bauxite residue has been reported to be soluble at pH values above 9, mainly as mobile [Al(OH)⁴]⁻ (Milačič et al., 2012). Nevertheless, its 506 507 impact on microbial and plant communities should not be neglected, as even small 508 proportions of the total AI correspond to potentially toxic concentrations. Moreover, 509 fresh and recently deposited bauxite residues were inherently poor in most 510 macronutrients required for microbial and plant growth, such as TOC, TN, and P 511 (Table S2). TOC accounted for less than 0.6% of the mass (0.28 % in fresh residue, 0.33 % in Bo1, and 0.23 % in Bo2) which is typically attributed to desert soils (Li et al., 512 513 2013; Zech et al., 2014). TN concentrations were below the detection limit (100 mg/kg) 514 in fresh, Bo1 and Bo2 samples in agreement with previous studies (Krishna et al., 515 2014; Wu et al., 2020). Finally, although the P concentrations ranged from 1500 to 516 3000 mg/kg in fresh, Bo1, and Bo2 samples, the available phosphorus in bauxite 517 residue is usually around 6.7 mg/kg (Santini et al., 2015a; Wu et al., 2020).

These extreme conditions, together with the absence of pre-existing life in the freshly 518 519 produced bauxite residues, indicate that BRDAs behave as primary successional 520 environments, similar to other well-studied natural systems (Guo et al., 2014; Schmidt 521 et al., 2008). Primary succession is characterized by an initial short phase where the 522 microbial community assembly is driven by stochastic dispersal processes, followed 523 by a longer deterministic phase dominated by environmental selection and competition between species (Ortiz-Álvarez et al., 2018). In engineered environments like BRDAs, 524 525 the dispersal of microorganisms is limited, and environmental selection pressure is 526 usually stronger due to the extreme chemical characteristics of these sites. Therefore, 527 a shorter stochastic phase and a stronger deterministic assembly is expected in such 528 systems (Santini et al., 2015a). Given the environmental stress imposed by the 529 geochemical properties of fresh bauxite residues, the first pioneer microorganisms to 530 colonize BRDAs are likely to be haloalkaliphilic species able to accumulate nutrients

to facilitate the incorporation of new species (Santini et al., 2015a; Schmidt et al., 2014;

532 Sun et al., 2018).

4.2 Geochemistry-dependent microbial community assembly in the early stages of primary succession

535 The bacterial community composition of the recently deposited bauxite residues (Bo1 536 and Bo2) was similar to that observed in soda lakes, alkaline residues, and other 537 unamended BRDAs (Bondici et al., 2013; Chakraborty et al., 2021; Kalwasińska et al., 2017; Ohlsson et al., 2019), characterized by the dominance of Actinobacteria 538 539 (Actinobacteria, Nitriliruptoria, Acidimicrobiia) and Proteobacteria (especially 540 Gammaproteobacteria) and relatively high proportions (1 to 16 %) of Bacteroidetes 541 (Cytophagia, Rhodothermi, Flavobacteriia), Firmicutes (Bacilli, Clostridia), and other 542 alkalophilic classes (Opitutae, Deinococci) (Fig. 5-6).

543 The phyla Actinobacteria, Proteobacteria and Firmicutes contain some of the 544 prokaryotes best adapted to high alkalinity and salinity. Alkaliphilic and alkalinotolerant 545 microorganisms maintain their intracellular pH homeostasis by accumulating H⁺ in 546 their cytoplasm trough Na⁺/H⁺ antiporters (Mamo, 2020). Our results showed that 547 Halobacteria (Archaea), Gammaproteobacteria, Nitriliruoptoria and Rhodothermi were more abundant in Bo1 (Fig. 6-7) and correlated with EC (Fig. 10), corroborating the 548 549 preference of these classes for high-salinity niches (Zhang et al., 2019). Extreme 550 halophiles accumulate intracytoplasmic KCI to regulate their osmotic potential, which 551 constrains them to be obligate halophiles and limits their dispersal (Chen et al., 2020; 552 Shameer, 2016; Vaidya et al., 2018). Salinity (EC) decreased significantly from Bo1 to 553 Bo2 (Fig. 1, Table S2), allowing the proliferation of a wider range of halophilic and 554 halotolerant species such as Opitutae, Acidimicrobiia or the genus Halomonas (Gammaproteobacteria) (Fig. 6-7). Halotolerant bacteria can flexibly adapt to habitats 555 556 with different salinities by synthesizing and accumulating compatible solutes in their 557 cytoplasm, which in turn requires a more intensive use of energy (Gunde-Cimerman 558 et al., 2018). Interestingly, the number of edges and network complexity was higher in 559 Bo1 compared with Bo2 (Fig. 8), indicating that saline stress promoted microbial 560 interaction as reported in other studies (Ji et al., 2019; Wang et al., 2019).

561 As pioneer colonizers, microorganisms are critical in the biogeochemical cycles and 562 the ensuing development of soil (Schmidt et al., 2008; Zeng et al., 2016). TOC 563 concentration showed no significant trend in early-stage samples (Fig. 2, Table S2), suggesting that the ratio in TOC input and consumption was balanced. In primary 564 565 successional environments, the main sources of TOC are airborne allochthonous 566 organic matter and C fixed by autotrophic microbial species (Ciccazzo et al., 2016). 567 Our results also highlighted the significant abundance of heterotrophic bacteria in the 568 early-stage samples, especially in Bo1, enriched in Actinobacteria and Bacteroidetes 569 (Fig. 6-7). MCIA revealed that Cytophagia (Bacteroidetes) and Nitriliruptoria (Actinobacteria) correlated negatively with TOC (Fig. 10) and members of these 570 571 classes were identified as keystone species in Bo1 (Fig. 8), confirming their relevance 572 in oligotrophic habitats (Foreman et al., 2007; Gonzalez-Pimentel et al., 2018). 573 Heterotrophic communities may represent the earliest stage of microbial assembly, 574 degrading allochthonous organic compounds and supplying essential nutrients (TOC, 575 N, P) for the subsequent development of autotrophic species (Hodkinson et al., 2002). 576 Our results also indicated the presence of chemoautotrophic Proteobacteria typically 577 involved in the primary production of alkaline environments (Fig. 7), including iron-578 oxidizing bacteria within Alphaproteobacteria (Rhodobacteraceae) and 579 Betaproteobacteria (Comamonadaceae) (Jamieson et al., 2018; Kumaraswamy et al., 580 2006; Straub et al., 1996) and sulfur-oxidizing Chromatiales (Gammaproteobacteria) 581 (Yuan et al., 2021). In tune with previous studies (Harantová et al., 2017; Schmidt et 582 al., 2008), the abundance of non-symbiotic N-fixing bacteria, represented by members 583 of Clostridia, Opitutae, Comamonadaceae (Betaproteobacteria), Rhodospirillales 584 (Alphaproteobacteria) and Bacilli (Ciccazzo et al., 2016; Hingole and Pathak, 2013; 585 Navarro-Noya et al., 2016), increased significantly after the first years of primary 586 succession (Fig. 5-6). These bacteria could contribute to the slight increase in TN 587 observed after the first years of bauxite residue weathering (Wu et al., 2020), 588 undetectable in this study as it is expected to be less than 100 mg/kg. Nitrifying 589 bacteria (Nitrosomonas and Nitrospiraceae) were also more prevalent in Bo2, 590 confirming that, although the amount of ammonia derived from bacterial fixation is low 591 at early stages of primary succession, it must be sufficient to fuel nitrification (Ollivier 592 et al., 2011; Zeng et al., 2016). Moreover, our results suggest that the increase in N-593 cycle bacteria could be due to a significant decrease in EC and Cr in Bo2 (Fig. 2, Fig. 594 **10**, Table S2), as N-cycle enzymes are particularly sensitive to high salinity (Claros et 595 al., 2010; Herbst, 1998) and heavy metal contamination (Kim et al., 2016; Oliveira and 596 Pampulha, 2006).

597 **4.3 Primary succession reaches an equilibrium in BRDA over few decades**

598 The last stage of primary succession arises after long periods of time when the habitat 599 becomes less harsh and environmental selection loses its strength in the assembly of 600 communities, reaching an equilibrium (Ferrenberg et al., 2013). This maturity is 601 expected to require more time to be developed in extreme geochemical systems as 602 BRDAs. (Santini et al., 2015a). In our study, Pr90 and Pr100 samples, which have 603 been naturally weathered for at least 90 years, did not show significant dissimilarities 604 in either physicochemical parameters (Fig. 3a, Tables S1 and S2) or bacterial and 605 fungal community composition (Fig. 4b; Fig. S5), suggesting that primary succession 606 had reached the equilibrium in these sites. In Pr90 and Pr100, pH (8.8 ± 0.3) and 607 salinity (EC = 0.29 ± 0.11 mS cm⁻¹) were significantly lower than in early-stage 608 samples (Fig. 2, Table S1), revealing a significant dependence with the age of the 609 samples. During bauxite residue natural weathering, salinity is generally assumed to 610 decrease due to the dissolution of alkaline minerals by rainfall (Cusack et al., 2019; 611 Zhu et al., 2016), while the decrease in pH is normally attributed to atmospheric and 612 microbial carbonation (Cusack et al., 2019; Khaitan et al., 2010; Schmalenberger, 613 2013; X. Kong et al., 2017). Moreover, Pr90 and Pr100 showed a significantly higher 614 concentration of macronutrients such as TOC, TN and P, which are accumulated 615 during the bauxite residue natural aging due to the action of the different microorganisms that colonize the BRDAs (Santini et al., 2015a). This negative 616 617 correlation between pH-salinity and age-nutrients is characteristic of restored and 618 amended BRDAs (Courtney et al., 2014; Wu et al., 2021). Finally, our results also 619 indicated a significant and steady decrease in Cr and V associated with the natural 620 aging of the samples (Fig. 2, Table S2). However, the concentration of these metals 621 also seems to be determined by the bauxite type and, therefore, the effect of the ore origin should not be neglected (Gentzmann et al., 2021). 622

Both alpha and beta bacterial diversity increased significantly in the equilibrium-stage samples (**Fig. 4**), in line with the results observed in restored BRDAs (Krishna et al., 2014; Wu et al., 2021, 2020). Beta-diversity analysis confirmed that "age" was the main factor explaining the dissimilarities of microbial community between Bo1/Bo2 and Pr90/Pr100 and not the ore origin. In addition, the changes in beta-diversity could be explained by pH, EC and some metal concentration gradients, all of them strongly anticorrelated with the age of the samples (**Fig. 3**). Circumneutral pH, lower salinity and 630 metal concentration, and increase in macronutrients support the survival of a more 631 diverse microbial community dominated by bacteria commonly found in ordinary soils and freshwater like Alphaproteobacteria, Betaproteobacteria, Planctomycetes, 632 633 Gemmatimonaidetes, and Acidobacteria. Acidobacteria are one of the most abundant 634 terrestrial bacterial taxa, reaching 52% from the total bacterial community in certain 635 soils (Kielak et al., 2016). Their abundance is positively correlated with low pH (Jones 636 et al., 2009), and their accumulation over time in BRDAs is commonly associated with 637 the restoration of chemical conditions in bauxite residue (Santini et al., 2015b; Wu et 638 al., 2021). The equilibrium-stage bauxite residue also harbors a more complex 639 microbial network, accounting for more highly connected taxa (nodes) than early-stage 640 residues. Nutrient availability and higher microbial richness have been shown to favor 641 microbial network complexity and the emergence of new ecological functions, making 642 the ecosystem more stable (Qiu et al., 2021; Wagg et al., 2019).

643 In contrast, fungal biodiversity and community structure did not appear to be 644 significantly affected by chemical gradients (Table S4, Fig. S5). The fungal community structure in early-stage bauxite residues was dominated by Ascomycota (80.1 \pm 9.8 645 646 % of fungal communities), and Basidiomycota (> 14 %). This distribution seems to be the usual in natural and engineered haloalkaline environments, although the analysis 647 648 of fungal communities in these habitats is limited to few studies (Grum-Grzhimaylo et 649 al., 2016; Salano et al., 2017; Santini et al., 2015b). In the equilibrium-stage samples, 650 the relative abundance of Ascomycota decreased (52.1 ± 24.2 %) while 651 Basidiomycota (mostly Agaricomycetes) increased (35.0 ± 23.3 %). This pattern has 652 been observed in restored BRDAs and was attributed to a decrease in total alkalinity (Santini et al., 2015b). In this study however, MCIA revealed a poor correlation 653 654 between fungal communities and chemical parameters (RV score = 0.36). It is known 655 that the fungal community assembly is more influenced by stochastic processes in 656 primary successional environments (Schmidt et al., 2014). Unlike bacteria, climate is 657 often considered the main environmental factor affecting the fungal community 658 composition, rather than soil chemical properties (Egidi et al., 2019). Interestingly, all 659 fungal keystones identified in Bo1 (Fig. 8) belonged to the class Dothideomycetes (Pleosporaceae, Cladosporium), 660 Macrophoma, within the Ascomycota. 661 Dothideomycetes constitute almost the entire fungal community in microbial mats from soda lakes and are also found in microbial mats from hypersaline and iron-rich habitats 662

(Gerea et al., 2012; Maza-Márquez et al., 2021; Salano et al., 2017). In microbial mats,
fungi play a crucial role in nutrient recycling by decomposing complex carbohydrates
into simpler compounds that fuel chemoheterotrophic species (Carreira et al., 2020).

666 **4.4 Implications for bioremediation and metal recovery**

To date, only few studies have focused their efforts on an integrated chemical,
 physical, and biological approach for the characterization of BRDA in the optic of
 bioremediation or critical metal bio extraction strategies.

670 Our results highlight the key role of pH and EC for natural microbial restoration and 671 therefore should be considered primary targets in bioremediation. If natural weathering 672 appears to be effective in reducing both pH and EC, it is also a slow process that could 673 be artificially accelerated. In the last few years, microbially-driven pH neutralization using organic acids and CO₂ produced by fermentation of added carbon sources has 674 675 gained attention as a promising technique for in-situ bioremediation of BRDAs (Santini 676 et al., 2021; Wu et al., 2019). Santini et al. (2021, 2016) proved that the efficiency of 677 this process could be improved by decreasing the initial pH and salinity and increasing 678 the biodiversity of the microbial inoculum. However, they also identified the shortage 679 of nutrients (N and P) in the medium term as a limiting factor for the development of 680 this methodology. Our results suggest that lowering EC increased biodiversity and 681 facilitated the development of bacteria related to the N cycle. Hence, a first treatment 682 focused on reducing salinity (and pH if possible) before bacteria-mediated pH 683 neutralization could yield considerably better results. Few studies have investigated 684 the effect of lowering artificially pH and EC in the optic of bioremediation of BRDAs 685 (Courtney et al., 2014; Jones et al., 2011; Wong and Ho, 1994). As a recent example, 686 the recent study by Fourrier et al. (2021) has reported the positive effect of gypsum addition and repeated washing on pH neutralization and EC decrease. In addition, 687 688 fungal contribution to early-stage microbial assembly seems to be underrated and their 689 role in bauxite residue remediation needs more research. Regarding microbially-690 driven strategies for metal bio-extraction, our study is particularly informative for 691 selecting appropriate bauxite residue and bacteria for the selective extraction of critical 692 metals. In agreement with literature (Rivera et al., 2019; Vind et al., 2018), our results 693 indicated that the concentration of REEs (Ln, Sc and Y) in bauxite residue is primarily 694 determined by the ore origin, with karst bauxite (Pr90 and Pr100) being more enriched 695 in REEs than laterite bauxite (Fresh, Bo1 and Bo2). In addition, Y and Ln were strongly

696 correlated as expected regarding the analogy between Y and heavy REEs while Sc 697 behaves differently to Y and Ln and is likely incorporated in iron oxyhydroxyde such 698 as goethite (Levard et al., 2018). The conventional methods to recover these metals 699 are not environmentally acceptable due to their excessive energy consumption and 700 the production of hazardous residues, that require further treatment and high operating 701 costs (Baniasadi et al., 2019). Metal bio-extraction is an interesting alternative to 702 traditional mettalurgy that has still been poorly investigated for critical metals. Among 703 the ideas that could benefit from our study, alkaline-active exoenzymes produced by 704 bacteria found in Bo1 and Bo2 offer interesting possibilities for the valorization of 705 bauxite residue. Despite their ability to lower their intracellular pH and salinity, 706 haloalkaliphilic species must still be able to secrete stable and operational enzymes 707 at elevated pH and salinity (Mamo, 2020), making BRDAs potential sources of these 708 enzymes. Some of the haloalkaliphilic bacteria known to synthesize alkaline-active 709 enzymes, namely Gammaproteobacteria, Actinobacteria and Firmicutes (Kalwasińska 710 et al., 2018; Litchfield, 2011; Maharaja et al., 2018; Shameer, 2016; Shivlata and 711 Tulasi, 2015), are also capable to modify the chemical phase and mobility of various 712 critical metals (Chidambaram et al., 2010; Presentato et al., 2020; Wee et al., 2014), 713 including REEs. For example, Maleke et al. (2019) reported the direct reduction of Eu³⁺ to Eu²⁺ by *Clostridium* sp. (Firmicutes). In line with previous studies (Santini et 714 715 al., 2015b), the majority of Gammaproteobacteria corresponded to uncharacterized 716 lineages, highlighting the potential of this group as a phylogenetic hot spot for novel 717 haloalkaliphilic taxa.

718

719 **5. CONCLUSIONS**

In summary, this paper presents an integrated physicochemical and biological 720 721 approach to explore the composition and dynamics of bacterial and fungal 722 communities in bauxite residues deposited at different times and produced from 723 different bauxite ores. Our results revealed that both deposit age and ore origin affect 724 the geochemistry of bauxite residue, although unequally. Salinity, pH, TOC, TN, and 725 P values seem to depend predominantly on the natural aging of bauxite residue, while 726 the content of REE is mainly influenced by the origin of the bauxite ore. Our results 727 highlight the behavior of bauxite residue deposits as primary successional 728 environments and bring new insights into the early stages of microbial community 729 assembly in these sites. The pioneer microbial community was dominated by 730 haloalkaliphilic microorganisms, strongly influenced by chemical gradients. Autotrophic and heterotrophic microbial species contribute to the supply of nutrients 731 732 necessary for the development of other species through C fixation and degradation of 733 allochthonous organic compounds, respectively. After the first years of natural 734 restoration, nitrogen-fixing bacteria increase their presence and contribute to enhance 735 N bioavailability. Microbial richness, diversity and network complexity increased 736 significantly with age of deposition, until primary succession equilibrium was reached 737 decades later, characterized by a microbial community composition similar to that of 738 typical soils and freshwater. Our results suggested that salinity, pH, nutrients, and toxic 739 metals (mainly Cr and V) were the main factors explaining this change in microbial 740 communities. These results confirm the key role of pH and salinity in the establishment 741 of early microbial communities and highlight them as main targets for bioremediation. Moreover, our co-occurrence network data suggest an important role of fungal 742 743 communities in structuring the early-stage microbial community during primary succession, potentially by recycling complex organic matter. We also identified 744 745 bacteria with potential metal extraction abilities, such as secretion of alkaline-active 746 enzymes that could modify the chemical phase of the metals present.

747

Author contributions: Conceptualization C.L., W.A., M.A.; Data curation L.A.M-P.;
Formal analysis L.A.M-P.; Funding acquisition C.L., W.A., M.A.; Investigation L.A.MP.; Methodology L.A.M-P., B.A., D.B., M.B., L.P.; Project administration C.L., W.A.,
M.A.; Supervision C.L., W.A., M.A.; Validation C.L., W.A., M.A., L.A.M-P., B.A., D.B.
M.B., L.P.; Writing - original draft L.A.M-P.; Writing - review & editing L.A.M-P., C.L.,
W.A., M.A. All authors have read and agreed to the published version of the
manuscript.

755

Funding: This project has received funding from the European Union's Horizon 2020
 research and innovation program under the Marie Skłodowska-Curie grant agreement
 No713750. Additionally, it has been performed with the financial supports of the
 Regional Council of Provence-Alpes-Côte d'Azur and A*MIDEX (n° ANR11-IDEX-

- 760 0001-02), funded by the Investissements d'Avenir project funded by the French
- 761 Government and managed by the French National Research Agency (ANR).
- 762
- 763 Acknowledgments: This work is also a contribution to the OSU-Institut Pythéas. The
- authors acknowledge the CNRS for the funding of the IRP iNOVE. The authors thank 764
- 765 the company ALTEO for providing access to the sampling sites.
- 766
- 767 Conflicts of Interest: The authors declare no conflict of interest.
- 768

769 6. REFERENCES

- 770 Anderson, M.J., Crist, T.O., Chase, J.M., Vellend, M., Inouye, B.D., Freestone, A.L., 771 Sanders, N.J., Cornell, H.V., Comita, L.S., Davies, K.F., Harrison, S.P., Kraft, N.J.B., 772 Stegen, J.C., Swenson, N.G., 2011. Navigating the multiple meanings of β diversity: a 773 roadmap for the practicing ecologist. Ecol. Lett. 14, 19-28. 774 https://doi.org/10.1111/j.1461-0248.2010.01552.x
- 775 Baniasadi, M., Vakilchap, F., Bahaloo-Horeh, N., Mousavi, S.M., Farnaud, S., 2019.
- 776 Advances in bioleaching as a sustainable method for metal recovery from e-waste: A review. J. Ind. Eng. Chem. 76, 75–90. https://doi.org/10.1016/j.jiec.2019.03.047 777
- 778 Banning, N.C., Phillips, I.R., Jones, D.L., Murphy, D.V., 2011. Development of Microbial 779 Diversity and Functional Potential in Bauxite Residue Sand under Rehabilitation. 780 Restor. Ecol. 19, 78-87. https://doi.org/10.1111/j.1526-100X.2009.00637.x
- 781 Benjamini, Y., Hochberg, Y., 1995. Controlling the False Discovery Rate: A Practical and 782 Powerful Approach to Multiple Testing. J. R. Stat. Soc. Ser. B Methodol. 57, 289-783 300. https://doi.org/10.1111/j.2517-6161.1995.tb02031.x
- 784 Berry, D., Widder, S., 2014. Deciphering microbial interactions and detecting keystone 785 species with co-occurrence networks. Front. Microbiol. 5, 219. 786 https://doi.org/10.3389/fmicb.2014.00219
- 787 Beule, L., Karlovsky, P., 2020. Improved normalization of species count data in ecology by 788 scaling with ranked subsampling (SRS): application to microbial communities. PeerJ 789 8, e9593. https://doi.org/10.7717/peerj.9593
- 790 Bolyen, E., Rideout, J.R., Dillon, M.R., Bokulich, N.A., Abnet, C.C., Al-Ghalith, G.A., 791 Alexander, H., Alm, E.J., Arumugam, M., Asnicar, F., Bai, Y., Bisanz, J.E., Bittinger, 792 K., Brejnrod, A., Brislawn, C.J., Brown, C.T., Callahan, B.J., Caraballo-Rodríguez, 793 A.M., Chase, J., Cope, E.K., Da Silva, R., Diener, C., Dorrestein, P.C., Douglas, 794 G.M., Durall, D.M., Duvallet, C., Edwardson, C.F., Ernst, M., Estaki, M., Fouquier, 795 J., Gauglitz, J.M., Gibbons, S.M., Gibson, D.L., Gonzalez, A., Gorlick, K., Guo, J., 796 Hillmann, B., Holmes, S., Holste, H., Huttenhower, C., Huttley, G.A., Janssen, S., 797 Jarmusch, A.K., Jiang, L., Kaehler, B.D., Kang, K.B., Keefe, C.R., Keim, P., Kelley, 798 S.T., Knights, D., Koester, I., Kosciolek, T., Kreps, J., Langille, M.G.I., Lee, J., Ley, 799 R., Liu, Y.-X., Loftfield, E., Lozupone, C., Maher, M., Marotz, C., Martin, B.D., 800 McDonald, D., McIver, L.J., Melnik, A.V., Metcalf, J.L., Morgan, S.C., Morton, J.T., 801
 - Naimey, A.T., Navas-Molina, J.A., Nothias, L.F., Orchanian, S.B., Pearson, T.,

802	Peoples, S.L., Petras, D., Preuss, M.L., Pruesse, E., Rasmussen, L.B., Rivers, A.,
803	Robeson, M.S., Rosenthal, P., Segata, N., Shaffer, M., Shiffer, A., Sinha, R., Song,
804	S.J., Spear, J.R., Swafford, A.D., Thompson, L.R., Torres, P.J., Trinh, P., Tripathi, A.,
805	Turnbaugh, P.J., Ul-Hasan, S., van der Hooft, J.J.J., Vargas, F., Vázquez-Baeza, Y.,
806	Vogtmann, E., von Hippel, M., Walters, W., Wan, Y., Wang, M., Warren, J., Weber,
807	K.C., Williamson, C.H.D., Willis, A.D., Xu, Z.Z., Zaneveld, J.R., Zhang, Y., Zhu, Q.,
808	Knight, R., Caporaso, J.G., 2019. Reproducible, interactive, scalable and extensible
809	microbiome data science using QIIME 2. Nat. Biotechnol. 37, 852–857.
810	https://doi.org/10.1038/s41587-019-0209-9
811	Bondici, V.F., Lawrence, J.R., Khan, N.H., Hill, J.E., Yergeau, E., Wolfaardt, G.M., Warner,
812	J., Korber, D.R., 2013. Microbial communities in low permeability, high pH uranium
813	mine tailings: characterization and potential effects. J. Appl. Microbiol. 114, 1671–
814	1686. https://doi.org/10.1111/jam.12180
815	Bouchoucha, M., Chekri, R., Leufroy, A., Jitaru, P., Millour, S., Marchond, N., Chafey, C.,
816	Testu, C., Zinck, J., Cresson, P., Mirallès, F., Mahe, A., Arnich, N., Sanaa, M.,
817	Bemrah, N., Guérin, T., 2019. Trace element contamination in fish impacted by
818	bauxite red mud disposal in the Cassidaigne canyon (NW French Mediterranean). Sci.
819	Total Environ. 690, 16–26. https://doi.org/10.1016/j.scitotenv.2019.06.474
820	Brav, A.W., Stewart, D.I., Courtney, R., Rout, S.P., Humphreys, P.N., Mayes, W.M., Burke,
821	I.T., 2018. Sustained Bauxite Residue Rehabilitation with Gypsum and Organic
822	Matter 16 years after Initial Treatment. Environ. Sci. Technol. 52, 152–161.
823	https://doi.org/10.1021/acs.est.7b03568
824	Callahan, B.J., McMurdie, P.J., Holmes, S.P., 2017. Exact sequence variants should replace
825	operational taxonomic units in marker-gene data analysis. ISME J. 11, 2639–2643.
826	https://doi.org/10.1038/ismej.2017.119
827	Callahan, B.J., McMurdie, P.J., Rosen, M.J., Han, A.W., Johnson, A.J.A., Holmes, S.P.,
828	2016. DADA2: High resolution sample inference from Illumina amplicon data. Nat.
829	Methods 13, 581–583. https://doi.org/10.1038/nmeth.3869
830	Caporaso, J.G., Lauber, C.L., Walters, W.A., Berg-Lyons, D., Lozupone, C.A., Turnbaugh,
831	P.J., Fierer, N., Knight, R., 2011. Global patterns of 16S rRNA diversity at a depth of
832	millions of sequences per sample. Proc. Natl. Acad. Sci. 108, 4516–4522.
833	https://doi.org/10.1073/pnas.1000080107
834	Carreira, C., Lønborg, C., Kühl, M., Lillebø, A., Sandaa, RA., Villanueva, L., Cruz, S.,
835	2020. Fungi and viruses as important players in microbial mats. FEMS Microbiol.
836	Ecol. 96. https://doi.org/10.1093/femsec/fiaa187
837	Chakraborty, J., Rajput, V., Sapkale, V., Kamble, S., Dharne, M., 2021. Spatio-temporal
838	resolution of taxonomic and functional microbiome of Lonar soda lake of India
839	reveals metabolic potential for bioremediation. Chemosphere 264, 128574.
840	https://doi.org/10.1016/j.chemosphere.2020.128574
841	Chen, DD., Tian, Y., Jiao, JY., Zhang, XT., Zhang, YG., Dong, ZY., Xiong, MJ.,
842	Xiao, M., Shu, WS., Li, WJ., 2020. Comparative genomics analysis of
843	Nitriliruptoria reveals the genomic differences and salt adaptation strategies.
844	Extremophiles 24, 249–264. https://doi.org/10.1007/s00792-019-01150-3
845	Chen, L., Liu, J., Hu, W., Gao, J., Yang, J., 2021. Vanadium in soil-plant system: Source,
846	fate, toxicity, and bioremediation. J. Hazard. Mater. 405, 124200.
847	https://doi.org/10.1016/j.jhazmat.2020.124200
848	Chidambaram, D., Hennebel, T., Taghavi, S., Mast, J., Boon, N., Verstraete, W., van der
849	Lelie, D., Fitts, J.P., 2010. Concomitant microbial generation of palladium
850	nanoparticles and hydrogen to immobilize chromate. Environ. Sci. Technol. 44,
851	7635–7640. https://doi.org/10.1021/es101559r

- Ciccazzo, S., Esposito, A., Borruso, L., Brusetti, L., 2016. Microbial communities and 852 853 primary succession in high altitude mountain environments. Ann. Microbiol. 66, 43-854 60. https://doi.org/10.1007/s13213-015-1130-1
- 855 Claros, J., Jiménez, E., Borrás, L., Aguado, D., Seco, A., Ferrer, J., Serralta, J., 2010. Short-856 term effect of ammonia concentration and salinity on activity of ammonia oxidizing 857 bacteria. Water Sci. Technol. 61, 3008–3016. https://doi.org/10.2166/wst.2010.217
- 858 Copernicus Climate Change Service Climate Data Store (CDS), 2021. Copernicus Climate 859 Change Service (C3S) (2017): ERA5: Fifth generation of ECMWF atmospheric reanalyses of the global climate. ECMWF. 860
- 861 Courtney, R., Harris, J., Pawlett, M., 2014. Microbial Community Composition in a 862 Rehabilitated Bauxite Residue Disposal Area: A Case Study for Improving Microbial Community Composition. Restor. Ecol. 22. https://doi.org/10.1111/rec.12143 863
- Csardi, G., Nepusz, T., 2006. The igraph software package for complex network research, 864 865 InterJournal.
- 866 Cusack, P.B., Courtney, R., Healy, M.G., O' Donoghue, L.M.T., Ujaczki, É., 2019. An 867 evaluation of the general composition and critical raw material content of bauxite 868 residue in a storage area over a twelve-year period. J. Clean. Prod. 208, 393-401. 869 https://doi.org/10.1016/j.jclepro.2018.10.083
- 870 Dentoni, V., Grosso, B., Pinna, F., 2021. Experimental Evaluation of PM Emission from Red 871 Mud Basins Exposed to Wind Erosion. Minerals 11, 405. 872 https://doi.org/10.3390/min11040405
- 873 Dev, S., Sachan, A., Dehghani, F., Ghosh, T., Briggs, B.R., Aggarwal, S., 2020. Mechanisms 874 of biological recovery of rare-earth elements from industrial and electronic wastes: A 875 review. Chem. Eng. J. 397, 124596. https://doi.org/10.1016/j.cej.2020.124596
- 876 Di Carlo, E., Boullemant, A., Courtney, R., 2020. Ecotoxicological risk assessment of 877 revegetated bauxite residue: Implications for future rehabilitation programmes. Sci. 878 Total Environ. 698, 134344. https://doi.org/10.1016/j.scitotenv.2019.134344
- 879 Di Carlo, E., Chen, C.R., Haynes, R.J., Phillips, I.R., Courtney, R., 2019. Soil quality and 880 vegetation performance indicators for sustainable rehabilitation of bauxite residue 881 disposal areas: a review. Soil Res. 57, 419. https://doi.org/10.1071/SR18348
- 882 Doebelin, N., Kleeberg, R., 2015. Profex: a graphical user interface for the Rietveld 883 refinement program BGMN. J. Appl. Crystallogr. 48, 1573–1580. 884 https://doi.org/10.1107/S1600576715014685
- 885 Dominguez-Benetton, X., Varia, J.C., Pozo, G., Modin, O., Ter Heijne, A., Fransaer, J., 886 Rabaey, K., 2018. Metal recovery by microbial electro-metallurgy. Prog. Mater. Sci. 887 94, 435–461. https://doi.org/10.1016/j.pmatsci.2018.01.007
- 888 Economou-Eliopoulos, M., Frei, R., Megremi, I., 2016. Potential leaching of Cr(VI) from 889 laterite mines and residues of metallurgical products (red mud and slag): An integrated approach. J. Geochem. Explor. 162, 40-49. 890 891 https://doi.org/10.1016/j.gexplo.2015.12.007
- 892 Egidi, E., Delgado-Baquerizo, M., Plett, J.M., Wang, J., Eldridge, D.J., Bardgett, R.D., 893 Maestre, F.T., Singh, B.K., 2019. A few Ascomycota taxa dominate soil fungal 894 communities worldwide. Nat. Commun. 10, 2369. https://doi.org/10.1038/s41467-895 019-10373-z
- 896 European Commission, 2020. Critical Raw Materials for Strategic Technologies and Sectors 897 in the EU.
- 898 Evans, K., 2016. The History, Challenges, and New Developments in the Management and 899 Use of Bauxite Residue. J. Sustain. Metall. 2, 316-331.
- 900 https://doi.org/10.1007/s40831-016-0060-x

- 901 Ferrenberg, S., O'Neill, S.P., Knelman, J.E., Todd, B., Duggan, S., Bradley, D., Robinson, 902 T., Schmidt, S.K., Townsend, A.R., Williams, M.W., Cleveland, C.C., Melbourne, B.A., Jiang, L., Nemergut, D.R., 2013. Changes in assembly processes in soil 903 904 bacterial communities following a wildfire disturbance. ISME J. 7, 1102–1111. 905 https://doi.org/10.1038/ismej.2013.11 906 Foreman, C.M., Sattler, B., Mikucki, J.A., Porazinska, D.L., Priscu, J.C., 2007. Metabolic 907 activity and diversity of cryoconites in the Taylor Valley, Antarctica. J. Geophys. Res. 908 Biogeosciences 112. https://doi.org/10.1029/2006JG000358 Fourrier, C., Luglia, M., Hennebert, P., Foulon, J., Ambrosi, J.-P., Angeletti, B., Keller, C., 909 910 Criquet, S., 2020. Effects of increasing concentrations of unamended and gypsum
- 911modified bauxite residues on soil microbial community functions and structure A912mesocosm study. Ecotoxicol. Environ. Saf. 201, 110847.
- 913 https://doi.org/10.1016/j.ecoenv.2020.110847
- Fourrier, C., Luglia, M., Keller, C., Hennebert, P., Foulon, J., Ambrosi, J.-P., Angeletti, B.,
 Criquet, S., 2021. How Raw and Gypsum Modified Bauxite Residues Affect Seed
 Germination, Enzyme Activities, and Root Development of Sinapis alba. Water. Air.
 Soil Pollut. 232, 309. https://doi.org/10.1007/s11270-021-05232-x
- Fox, J., Weisberg, S., 2019. An R Companion to Applied Regression, Third. ed. Sage,
 Thousand Oaks, CA.
- Gentzmann, M.C., Schraut, K., Vogel, C., Gäbler, H.-E., Huthwelker, T., Adam, C., 2021.
 Investigation of scandium in bauxite residues of different origin. Appl. Geochem.
 126, 104898. https://doi.org/10.1016/j.apgeochem.2021.104898
- Gerea, A.L., Branscum, K.M., King, J.B., You, J., Powell, D.R., Miller, A.N., Spear, J.R.,
 Cichewicz, R.H., 2012. Secondary metabolites produced by fungi derived from a
 microbial mat encountered in an iron-rich natural spring. Tetrahedron Lett. 53, 4202–
 4205. https://doi.org/10.1016/j.tetlet.2012.05.156
- Ghosh, S., Bal, B., Das, A.P., 2018. Enhancing Manganese Recovery from Low-Grade Ores
 by Using Mixed Culture of Indigenously Isolated Bacterial Strains. Geomicrobiol. J.
 35, 242–246. https://doi.org/10.1080/01490451.2017.1362080
- Gonzalez-Pimentel, J.L., Miller, A.Z., Jurado, V., Laiz, L., Pereira, M.F.C., Saiz-Jimenez, C.,
 2018. Yellow coloured mats from lava tubes of La Palma (Canary Islands, Spain) are
 dominated by metabolically active Actinobacteria. Sci. Rep. 8, 1944.
 https://doi.org/10.1038/s41598-018-20393-2
- Grum-Grzhimaylo, A.A., Georgieva, M.L., Bondarenko, S.A., Debets, A.J.M., Bilanenko,
 E.N., 2016. On the diversity of fungi from soda soils. Fungal Divers. 76, 27–74.
 https://doi.org/10.1007/s13225-015-0320-2
- Gunde-Cimerman, N., Plemenitaš, A., Oren, A., 2018. Strategies of adaptation of
 microorganisms of the three domains of life to high salt concentrations. FEMS
 Microbiol. Rev. 42, 353–375. https://doi.org/10.1093/femsre/fuy009
- Guo, Y., Fujimura, R., Sato, Y., Suda, W., Kim, S., Oshima, K., Hattori, M., Kamijo, T.,
 Narisawa, K., Ohta, H., 2014. Characterization of Early Microbial Communities on
 Volcanic Deposits along a Vegetation Gradient on the Island of Miyake, Japan.
 Microbes Environ. 29, 38–49. https://doi.org/10.1264/jsme2.ME13142
- Harantová, L., Mudrák, O., Kohout, P., Elhottová, D., Frouz, J., Baldrian, P., 2017.
 Development of microbial community during primary succession in areas degraded
 by mining activities. Land Degrad. Dev. 28, 2574–2584.
 https://doi.org/10.1002/ldr.2817
- Herbst, D.B., 1998. Potential salinity limitations on nitrogen fixation in sediments from
 Mono Lake, California. Int. J. Salt Lake Res. 7, 261–274.
- 950 https://doi.org/10.1007/BF02441878

951	Hill, T.C.J., Walsh, K.A., Harris, J.A., Moffett, B.F., 2003. Using ecological diversity
952	measures with bacterial communities. FEMS Microbiol. Ecol. 43, 1–11.
953	https://doi.org/10.1111/j.1574-6941.2003.tb01040.x
954	Hingole, S., Pathak, A., 2013. Report on efficient salt stable Azospirillum a Lonar Soda Lake
955	isolate. Sci. Res. Report. 3, 200–203.
956	Hodkinson, I.D., Webb, N.R., Coulson, S.J., 2002. Primary community assembly on land –
957	the missing stages: why are the heterotrophic organisms always there first? J. Ecol.
958	90, 569–577. https://doi.org/10.1046/j.1365-2745.2002.00696.x
959	Ihrmark, K., Bödeker, I.T.M., Cruz-Martinez, K., Friberg, H., Kubartova, A., Schenck, J.,
960	Strid, Y., Stenlid, J., Brandström-Durling, M., Clemmensen, K.E., Lindahl, B.D.,
961	2012. New primers to amplify the fungal ITS2 region – evaluation by 454-sequencing
962	of artificial and natural communities. FEMS Microbiol. Ecol. 82, 666-677.
963	https://doi.org/10.1111/j.1574-6941.2012.01437.x
964	International Aluminium Institute (IAI), European Aluminium (EA), 2015. Bauxite Residue
965	Management: Best Practice [WWW Document]. URL www.world-aluminium.org
966	Jamieson, J., Prommer, H., Kaksonen, A.H., Sun, J., Siade, A.J., Yusov, A., Bostick, B.,
967	2018. Identifying and Ouantifying the Intermediate Processes during Nitrate-
968	Dependent Iron(II) Oxidation, Environ, Sci. Technol. 52, 5771–5781.
969	https://doi.org/10.1021/acs.est.8b01122
970	Ji, M., Kong, W., Yue, L., Wang, J., Deng, Y., Zhu, L., 2019, Salinity reduces bacterial
971	diversity, but increases network complexity in Tibetan Plateau lakes. FEMS
972	Microbiol. Ecol. 95. https://doi.org/10.1093/femsec/fiz190
973	Jones, B.E.H., Havnes, R.J., Phillips, I.R., 2011, Influence of organic waste and residue mud
974	additions on chemical, physical and microbial properties of bauxite residue sand.
975	Environ. Sci. Pollut. Res. 18, 199–211. https://doi.org/10.1007/s11356-010-0364-5
976	Jones, R.T., Robeson, M.S., Lauber, C.L., Hamady, M., Knight, R., Fierer, N., 2009, A
977	comprehensive survey of soil acidobacterial diversity using pyrosequencing and clone
978	library analyses. ISME J. 3, 442–453, https://doi.org/10.1038/ismei.2008.127
979	Kalwasińska, A., Felföldi, T., Szabó, A., Deia-Sikora, E., Kosobucki, P., Walczak, M., 2017.
980	Microbial communities associated with the anthropogenic, highly alkaline
981	environment of a saline soda lime. Poland. Antonie Van Leeuwenhoek 110, 945–962.
982	https://doi.org/10.1007/s10482-017-0866-v
983	Kalwasińska, A., Jankiewicz, U., Felföldi, T., Burkowska-But, A., Brzezinska, M.S., 2018.
984	Alkaline and Halophilic Protease Production by Bacillus luteus H11 and Its Potential
985	Industrial Applications, Food Technol, Biotechnol, 56, 553–561.
986	https://doi.org/10.17113/ftb.56.04.18.5553
987	Kassambara A 2021 rstatix Pine-Friendly Framework for Basic Statistical Tests
988	Kassambara A 2020 gopubr: "goplot?" Based Publication Ready Plots
989	Kassambara A Mundt F 2020 factoextra: Extract and Visualize the Results of
990	Multivariate Data Analyses
991	Ke W 2021 Appropriate human intervention stimulates the development of microbial
997	communities and soil formation at a long-term weathered bauxite residue disposal
003	area I Hazard Mater 10
994	Khaitan S. Dzombak D.A. Swallow P. Schmidt K. Fu, J. Lowry G.V. 2010 Field
995	Evaluation of Bauxite Residue Neutralization by Carbon Dioxide Vegetation and
906	Organic Amendments I Environ Eng 136 1045 1053
907	https://doi.org/10.1061/(ASCE)EE 10/3_7870.0000230
991 008	Kielak A M. Barreto C.C. Kowalchuk G.A. von Veen I.A. Kuramae F.F. 2016 The
900	Ecology of Acidobacteria: Moving beyond Genes and Genomes Front Microbiol 7
1000	https://doi.org/10.3380/fmich.2016.00744
1000	nups.//doi.org/10.5567/1111c0.2010.00/444

- 1001 Kim, Y.M., Park, H., Chandran, K., 2016. Nitrification inhibition by hexavalent chromium
 1002 Cr(VI) Microbial ecology, gene expression and off-gas emissions. Water Res. 92,
 1003 254–261. https://doi.org/10.1016/j.watres.2016.01.042
- Kiskira, K., Lymperopoulou, T., Tsakanika, L.-A., Pavlopoulos, C., Papadopoulou, K.,
 Ochsenkühn, K.-M., Lyberatos, G., Ochsenkühn-Petropoulou, M., 2021. Study of
 Microbial Cultures for the Bioleaching of Scandium from Alumina Industry By Products. Metals 11, 951. https://doi.org/10.3390/met11060951
- 1008 Krishna, P., Babu, A.G., Reddy, M.S., 2014. Bacterial diversity of extremely alkaline bauxite
 1009 residue site of alumina industrial plant using culturable bacteria and residue 16S
 1010 rRNA gene clones 12.
- 1011 Kumaraswamy, R., Sjollema, K., Kuenen, J.G., van Loosdrecht, M., Muyzer, G., 2006.
 1012 Paracoccus ferroxidans.
- 1013 Lê, S., Josse, J., Husson, F., 2008. FactoMineR: An R Package for Multivariate Analysis. J.
 1014 Stat. Softw. 25. https://doi.org/10.18637/jss.v025.i01
- Levard, C., Borschneck, D., Grauby, O., Rose, J., Ambrosi, J.P., 2018. Goethite, a tailormade host for the critical metal scandium: The FexSc(1-x)OOH solid solution.
 Geochem. Perspect. Lett. 9, 16–20. https://doi.org/10.7185/geochemlet.1832
- Li, K., Liu, R., Zhang, H., Yun, J., 2013. The Diversity and Abundance of Bacteria and
 Oxygenic Phototrophs in Saline Biological Desert Crusts in Xinjiang, Northwest
 China. Microb. Ecol. 66, 40–48. https://doi.org/10.1007/s00248-012-0164-1
- Liang, J., Huang, X., Yan, J., Li, Y., Zhao, Z., Liu, Y., Ye, J., Wei, Y., 2021. A review of the
 formation of Cr(VI) via Cr(III) oxidation in soils and groundwater. Sci. Total
 Environ. 774, 145762. https://doi.org/10.1016/j.scitotenv.2021.145762
- Litchfield, C.D., 2011. Potential for industrial products from the halophilic Archaea. J. Ind.
 Microbiol. Biotechnol. 38, 1635–1635. https://doi.org/10.1007/s10295-011-1021-9
- Lozupone, C.A., Hamady, M., Kelley, S.T., Knight, R., 2007. Quantitative and Qualitative β
 Diversity Measures Lead to Different Insights into Factors That Structure Microbial
 Communities. Appl. Environ. Microbiol. 73, 1576–1585.
 https://doi.org/10.1128/AEM.01996-06
- Lyu, F., Hu, Y., Wang, L., Sun, W., 2021. Dealkalization processes of bauxite residue: A
 comprehensive review. J. Hazard. Mater. 403, 123671.
 https://doi.org/10.1016/j.jhazmat.2020.123671
- Ma, L., Wang, H., Wu, J., Wang, Y., Zhang, D., Liu, X., 2019. Metatranscriptomics reveals
 microbial adaptation and resistance to extreme environment coupling with
 bioleaching performance. Bioresour. Technol. 280, 9–17.
 https://doi.org/10.1016/j.biortech.2019.01.117
- Ma, L., Wang, X., Feng, X., Liang, Y., Xiao, Y., Hao, X., Yin, H., Liu, H., Liu, X., 2017. Coculture microorganisms with different initial proportions reveal the mechanism of chalcopyrite bioleaching coupling with microbial community succession. Bioresour.
 Technol. 223, 121–130. https://doi.org/10.1016/j.biortech.2016.10.056
- Maes, S., Claus, M., Verbeken, K., Wallaert, E., De Smet, R., Vanhaecke, F., Boon, N.,
 Hennebel, T., 2016. Platinum recovery from industrial process streams by halophilic
 bacteria: Influence of salt species and platinum speciation. Water Res. 105, 436–443.
 https://doi.org/10.1016/j.watres.2016.09.023
- Maharaja, P., Nanthini, E., Swarnalatha, S., Sekaran, G., 2018. Studies on the Production of
 Salt-Tolerant Alkaline Protease Isolated from Proteus mirabilis and Its Degradation of
 Hyper-Saline Soak Liquor, in: Singh, V.P., Yadav, S., Yadava, R.N. (Eds.),
- 1048 Environmental Pollution, Water Science and Technology Library. Springer,
- 1049 Singapore, pp. 439–457. https://doi.org/10.1007/978-981-10-5792-2_35

- Maleke, M., Valverde, A., Gomez-Arias, A., Cason, E.D., Vermeulen, J.-G., Coetsee-Hugo,
 L., Swart, H., van Heerden, E., Castillo, J., 2019. Anaerobic reduction of europium by
 a Clostridium strain as a strategy for rare earth biorecovery. Sci. Rep. 9, 14339.
 https://doi.org/10.1038/s41598-019-50179-z
- Mamo, G., 2020. Challenges and Adaptations of Life in Alkaline Habitats, in: Mamo, G.,
 Mattiasson, B. (Eds.), Alkaliphiles in Biotechnology, Advances in Biochemical
 Engineering/Biotechnology. Springer International Publishing, Cham, pp. 85–133.
 https://doi.org/10.1007/10_2019_97
- Markus Gräfe, Matthew Landers, Ryan Tappero, Peter Austin, Bee Gan, Alton Grabsch,
 Craig Klauber, 2011. Combined Application of QEM-SEM and Hard X-ray
 Microscopy to Determine Mineralogical Associations and Chemical Speciation of
 Trace Metals. J. Environ. Qual. https://doi.org/10.2134/jeq2010.0214
- Maza-Márquez, P., Lee, M.D., Bebout, B.M., 2021. The Abundance and Diversity of Fungi
 in a Hypersaline Microbial Mat from Guerrero Negro, Baja California, México. J.
 Fungi Basel Switz. 7, 210. https://doi.org/10.3390/jof7030210
- McDonald, D., Price, M.N., Goodrich, J., Nawrocki, E.P., DeSantis, T.Z., Probst, A.,
 Andersen, G.L., Knight, R., Hugenholtz, P., 2012. An improved Greengenes
 taxonomy with explicit ranks for ecological and evolutionary analyses of bacteria and
 archaea. ISME J. 6, 610–618. https://doi.org/10.1038/ismej.2011.139
- Meng, C., Kuster, B., Culhane, A.C., Gholami, A.M., 2014. A multivariate approach to the
 integration of multi-omics datasets. BMC Bioinformatics 15, 162.
 https://doi.org/10.1186/1471-2105-15-162
- Milačič, R., Zuliani, T., Ščančar, J., 2012. Environmental impact of toxic elements in red
 mud studied by fractionation and speciation procedures. Sci. Total Environ. 426, 359–
 365. https://doi.org/10.1016/j.scitotenv.2012.03.080
- Muller, L.A.H., Ballhausen, M.-B., Andrade-Linares, D.R., Pinek, L., Golubeva, P., Rillig,
 M.C., 2021. Fungus-bacterium associations are widespread in fungal cultures isolated
 from a semi-arid natural grassland in Germany. FEMS Microbiol. Ecol. 97, fiab059.
 https://doi.org/10.1093/femsec/fiab059
- Muyzer, G., de Waal, E.C., Uitterlinden, A.G., 1993. Profiling of complex microbial
 populations by denaturing gradient gel electrophoresis analysis of polymerase chain
 reaction-amplified genes coding for 16S rRNA. Appl. Environ. Microbiol. 59, 695–
 700.
- 1083 Navarro-Noya, Y.E., Luna-Guido, M., Dendooven, L., 2016. Cultivable Nitrogen Fixing
 1084 Bacteria from Extremely Alkaline-Saline Soils. Adv. Microbiol. 6, 412–423.
 1085 https://doi.org/10.4236/aim.2016.66041
- Ohlsson, J.I., Osvatic, J.T., Becraft, E.D., Swingley, W.D., 2019. Microbial Community in
 Hyperalkaline Steel Slag-Fill Emulates Serpentinizing Springs. Diversity 11, 103.
 https://doi.org/10.3390/d11070103
- 1089 Oksanen, J., Blanchet, F.G., Friendly, M., Kindt, R., Legendre, P., McGlinn, D., Minchin,
 1090 P.R., O'Hara, R.B., Simpson, G.L., Solymos, P., Stevens, M.H.H., Szoecs, E.,
 1091 Wagner, H., 2020. vegan: Community Ecology Package.
- 1092 Oliveira, A., Pampulha, M.E., 2006. Effects of long-term heavy metal contamination on soil
 1093 microbial characteristics. J. Biosci. Bioeng. 102, 157–161.
 1094 https://doi.org/10.1263/jbb.102.157
- Ollivier, J., Töwe, S., Bannert, A., Hai, B., Kastl, E.-M., Meyer, A., Su, M.X., Kleineidam,
 K., Schloter, M., 2011. Nitrogen turnover in soil and global change. FEMS Microbiol.
 Ecol. 78, 3–16. https://doi.org/10.1111/j.1574-6941.2011.01165.x
- 1098 Ortiz-Álvarez, R., Fierer, N., de los Ríos, A., Casamayor, E.O., Barberán, A., 2018.
 1099 Consistent changes in the taxonomic structure and functional attributes of bacterial

1100	communities during primary succession. ISME J. 12, 1658–1667.
1101	https://doi.org/10.1038/s41396-018-0076-2
1102	Panda, S., Costa, R.B., Shah, S.S., Mishra, S., Bevilaqua, D., Akcil, A., 2021.
1103	Biotechnological trends and market impact on the recovery of rare earth elements
1104	from bauxite residue (red mud) – A review. Resour. Conserv. Recycl. 171, 105645.
1105	https://doi.org/10.1016/j.resconrec.2021.105645
1106	Perez-Garcia, O., Lear, G., Singhal, N., 2016. Metabolic Network Modeling of Microbial
1107	Interactions in Natural and Engineered Environmental Systems. Front. Microbiol. 7.
1108	https://doi.org/10.3389/fmicb.2016.00673
1109	Poudel, R., Jumpponen, A., Schlatter, D.C., Paulitz, T.C., Gardener, B.B.M., Kinkel, L.L.,
1110	Garrett, K.A., 2016. Microbiome Networks: A Systems Framework for Identifying
1111	Candidate Microbial Assemblages for Disease Management. Phytopathology® 106,
1112	1083–1096. https://doi.org/10.1094/PHYTO-02-16-0058-FI
1113	Presentato, A., Piacenza, E., Turner, R.J., Zannoni, D., Cappelletti, M., 2020. Processing of
1114	Metals and Metalloids by Actinobacteria: Cell Resistance Mechanisms and Synthesis
1115	of Metal(loid)-Based Nanostructures. Microorganisms 8, 2027.
1116	https://doi.org/10.3390/microorganisms8122027
1117	Qiu, L., Zhang, Q., Zhu, H., Reich, P.B., Banerjee, S., van der Heijden, M.G.A., Sadowsky,
1118	M.J., Ishii, S., Jia, X., Shao, M., Liu, B., Jiao, H., Li, H., Wei, X., 2021. Erosion
1119	reduces soil microbial diversity, network complexity and multifunctionality. ISME J.
1120	15, 2474–2489. https://doi.org/10.1038/s41396-021-00913-1
1121	Qu, Y., Li, H., Wang, X., Tian, W., Shi, B., Yao, M., Zhang, Y., 2019. Bioleaching of Major,
1122	Rare Earth, and Radioactive Elements from Red Mud by using Indigenous
1123	Chemoheterotrophic Bacterium Acetobacter sp. Minerals 9, 67.
1124	https://doi.org/10.3390/min9020067
1125	R Core Team, 2020. R: A language and environment for statistical computing. R Foundation
1126	for Statistical Computing, Vienna, Austria.
1127	Ren, J., Chen, J., Han, L., Wang, M., Yang, B., Du, P., Li, F., 2018. Spatial distribution of
1128	heavy metals, salinity and alkalinity in soils around bauxite residue disposal area. Sci.
1129	Total Environ. 628–629, 1200–1208. https://doi.org/10.1016/j.scitotenv.2018.02.149
1130	Rivera, R.M., Ounoughene, G., Malfliet, A., Vind, J., Panias, D., Vassiliadou, V.,
1131	Binnemans, K., Van Gerven, T., 2019. A Study of the Occurrence of Selected Rare-
1132	Earth Elements in Neutralized–Leached Bauxite Residue and Comparison with
1133	Untreated Bauxite Residue. J. Sustain. Metall. 5, 57–68.
1134	https://doi.org/10.1007/s40831-018-0206-0
1135	Roy, A., Dutta, A., Pal, S., Gupta, A., Sarkar, J., Chatterjee, A., Saha, A., Sarkar, P., Sar, P.,
1136	Kazy, S.K., 2018. Biostimulation and bioaugmentation of native microbial
1137	community accelerated bioremediation of oil refinery sludge. Bioresour. Technol.
1138	253, 22–32. https://doi.org/10.1016/j.biortech.2018.01.004
1139	Sajjad, W., Zheng, G., Ma, X., Xu, W., Ali, B., Rafiq, M., Zada, S., Irfan, M., Zeman, J.,
1140	2020. Dissolution of Cu and Zn-bearing ore by indigenous iron-oxidizing bacterial
1141	consortia supplemented with dried bamboo sawdust and variations in bacterial
1142	structural dynamics: A new concept in bioleaching. Sci. Total Environ. 709, 136136.
1143	https://doi.org/10.1016/j.scitotenv.2019.136136
1144	Salano, O.A., Makonde, H.M., Kasili, R.W., Wangai, L.N., Nawiri, M.P., Boga, H.I., 2017.
1145	Diversity and distribution of fungal communities within the hot springs of soda lakes
1146	in the Kenyan rift valley. Afr. J. Microbiol. Res. 11, 764–775.
1147	https://doi.org/10.5897/AJMR2017.8510

- Santini, T., Kerr, J.L., Warren, L.A., 2015a. Microbially-driven strategies for bioremediation
 of bauxite residue. J. Hazard. Mater. 293, 131–157.
 https://doi.org/10.1016/j.jhazmat.2015.03.024
- Santini, T., Warren, L.A., Kendra, K.E., 2015b. Microbial Diversity in Engineered
 Haloalkaline Environments Shaped by Shared Geochemical Drivers Observed in
 Natural Analogues. Appl. Environ. Microbiol. 81, 5026–5036.
 https://doi.org/10.1128/AEM.01238-15
- Santini, T.C., Malcolm, L.I., Tyson, G.W., Warren, L.A., 2016. pH and Organic Carbon Dose
 Rates Control Microbially Driven Bioremediation Efficacy in Alkaline Bauxite
 Residue. Environ. Sci. Technol. 50, 11164–11173.
 https://doi.org/10.1021/acs.est.6b01973
- Santini, T.C., Wang, J.C., Warren, K.L., Pickering, G., Raudsepp, M.J., 2021. Simple
 Organic Carbon Sources and High Diversity Inocula Enhance Microbial
 Bioneutralization of Alkaline Bauxite Residues. Environ. Sci. Technol. 55, 3929–
 3939. https://doi.org/10.1021/acs.est.0c02534
- Santini, T.C., Warren, K., Raudsepp, M., Carter, N., Hamley, D., McCosker, C.,
 Couperthwaite, S., Southam, G., Tyson, G.W., Warren, L.A., 2019. Accelerating
 Bauxite Residue Remediation with Microbial Biotechnology, in: Chesonis, C. (Ed.),
 Light Metals 2019, The Minerals, Metals & Materials Series. Springer International
 Publishing, Cham, pp. 69–77. https://doi.org/10.1007/978-3-030-05864-7
- Publishing, Cham, pp. 69–77. https://doi.org/10.1007/978-3-030-05864-7_10
 Schmalenberger, A., 2013. Bacterial Communities Established in Bauxite Residues with
 Different Restoration Histories. Env. Sci Technol 10.
- Schmidt, S.K., Nemergut, D.R., Darcy, J.L., Lynch, R., 2014. Do bacterial and fungal
 communities assemble differently during primary succession? Mol. Ecol. 23, 254–
 258. https://doi.org/10.1111/mec.12589
- Schmidt, S.K., Reed, S.C., Nemergut, D.R., Stuart Grandy, A., Cleveland, C.C., Weintraub,
 M.N., Hill, A.W., Costello, E.K., Meyer, A.F., Neff, J.C., Martin, A.M., 2008. The
 earliest stages of ecosystem succession in high-elevation (5000 metres above sea
 level), recently deglaciated soils. Proc. R. Soc. B Biol. Sci. 275, 2793–2802.
 https://doi.org/10.1098/rspb.2008.0808
- Segata, N., Izard, J., Waldron, L., Gevers, D., Miropolsky, L., Garrett, W.S., Huttenhower,
 C., 2011. Metagenomic biomarker discovery and explanation. Genome Biol. 12, R60.
 https://doi.org/10.1186/gb-2011-12-6-r60
- Shameer, S., 2016. Haloalkaliphilic Bacillus species from solar salterns: an ideal prokaryote
 for bioprospecting studies. Ann. Microbiol. 66, 1315–1327.
 https://doi.org/10.1007/s13213-016-1221-7
- 1184 Shivlata, L., Tulasi, S., 2015. Thermophilic and alkaliphilic Actinobacteria: biology and 1185 potential applications. Front. Microbiol. 6. https://doi.org/10.3389/fmicb.2015.01014
- Srichandan, H., Mohapatra, R.K., Parhi, P.K., Mishra, S., 2019. Bioleaching approach for
 extraction of metal values from secondary solid wastes: A critical review.
 Hvdrometallurgy 189, 105122. https://doi.org/10.1016/i.hvdromet.2019.105122
- Straub, K.L., Benz, M., Schink, B., Widdel, F., 1996. Anaerobic, nitrate-dependent microbial
 oxidation of ferrous iron. Appl. Environ. Microbiol. 62, 1458–1460.
 https://doi.org/10.1128/aem.62.4.1458-1460.1996
- Sun, W., Xiao, E., Häggblom, M., Krumins, V., Dong, Y., Sun, X., Li, F., Wang, Q., Li, B.,
 Yan, B., 2018. Bacterial Survival Strategies in an Alkaline Tailing Site and the
 Physiological Mechanisms of Dominant Phylotypes As Revealed by Metagenomic
- 1195
 Analyses. Environ. Sci. Technol. 52, 13370–13380.
- 1196 https://doi.org/10.1021/acs.est.8b03853

- Tian, T., Liu, Z., Zhu, F., Hartley, W., Ye, Y., Xue, S., 2020. Improvement of aggregateassociated organic carbon and its stability in bauxite residue by substrate amendment
 addition. Land Degrad. Dev. 31, 2405–2416. https://doi.org/10.1002/ldr.3609
- Ujaczki, É., Zimmermann, Y., Gasser, C., Molnár, M., Feigl, V., Lenz, M., 2017. Red mud as secondary source for critical raw materials purification of rare earth elements by
 liquid/liquid extraction: Red mud as secondary source for critical raw materials. J.
 Chem. Technol. Biotechnol. 92, 2683–2690. https://doi.org/10.1002/jctb.5289
- Vaidya, S., Dev, K., Sourirajan, A., 2018. Distinct Osmoadaptation Strategies in the Strict
 Halophilic and Halotolerant Bacteria Isolated from Lunsu Salt Water Body of North
 West Himalayas. Curr. Microbiol. 75, 888–895. https://doi.org/10.1007/s00284-0181462-8
- 1208 Vind, J., Malfliet, A., Blanpain, B., Tsakiridis, P., Tkaczyk, A., Vassiliadou, V., Panias, D.,
 1209 2018. Rare Earth Element Phases in Bauxite Residue. Minerals 8, 77.
 1210 https://doi.org/10.3390/min8020077
- Wagg, C., Schlaeppi, K., Banerjee, S., Kuramae, E.E., van der Heijden, M.G.A., 2019.
 Fungal-bacterial diversity and microbiome complexity predict ecosystem functioning.
 Nat. Commun. 10, 4841. https://doi.org/10.1038/s41467-019-12798-y
- Wang, L., Yin, S., Wu, A., Chen, W., 2020. Synergetic bioleaching of copper sulfides using
 mixed microorganisms and its community structure succession. J. Clean. Prod. 245,
 118689. https://doi.org/10.1016/j.jclepro.2019.118689
- Wang, M., Chen, S., Chen, L., Wang, D., 2019. Responses of soil microbial communities and their network interactions to saline-alkaline stress in Cd-contaminated soils. Environ.
 Pollut. 252, 1609–1621. https://doi.org/10.1016/j.envpol.2019.06.082
- Wee, S.K., Burns, J.L., DiChristina, T.J., 2014. Identification of a molecular signature unique to metal-reducing Gammaproteobacteria. FEMS Microbiol. Lett. 350, 90–99. https://doi.org/10.1111/1574-6968.12304
- Wickham, H., 2009. ggplot2: Elegant Graphics for Data Analysis, 2197-5736. Springer Verlag, New York, NY.
- Wickham, H., François, R., Henry, L., Müller, K., 2021. dplyr: A Grammar of DataManipulation.
- Williams, R.J., Howe, A., Hofmockel, K.S., 2014. Demonstrating microbial co-occurrence
 pattern analyses within and between ecosystems. Front. Microbiol. 5, 358.
 https://doi.org/10.3389/fmicb.2014.00358
- Wong, J.W.C., Ho, G., 1994. Sewage sludge as organic ameliorant for revegetation of fine
 bauxite refining residue. Resour. Conserv. Recycl., Environmental biotechnology in
 waste treatment and recycling 11, 297–309. https://doi.org/10.1016/09213449(94)90097-3
- World Aluminium, 2020. World Aluminium [WWW Document]. URL www.world-aluminium.org (accessed 6.8.21).
- Wu, H., Chen, L., Zhu, F., Hartley, W., Zhang, Y., Xue, S., 2020. The dynamic development
 of bacterial community following long-term weathering of bauxite residue. J.
 Environ. Sci. 90, 321–330. https://doi.org/10.1016/j.jes.2019.12.001
- Wu, H., Liao, J., Zhu, F., Millar, G., Courtney, R., Xue, S., 2019. Isolation of an acid
 producing Bacillus sp. EEEL02: Potential for bauxite residue neutralization. J. Cent.
 South Univ. 26, 343–352. https://doi.org/10.1007/s11771-019-4006-x
- Wu, H., Tang, T., Zhu, F., Wei, X., Hartley, W., Xue, S., 2021. Long term natural restoration
 creates soil-like microbial communities in bauxite residue: A 50-year filed study.
 Land Degrad. Dev. 32, 1606–1617. https://doi.org/10.1002/ldr.3728

- 1245 X. Kong, Guo, Y., Xue, S., Hartley, W., Wu, C., Ye, Y., Cheng, Q., 2017. Natural evolution
 1246 of alkaline characteristics in bauxite residue. J. Clean. Prod. 143, 224–230.
 1247 https://doi.org/10.1016/j.jclepro.2016.12.125
- Xiao, X., Wang, M., Zhu, H., Guo, Z., Han, X., Zeng, P., 2017. Response of soil microbial
 activities and microbial community structure to vanadium stress. Ecotoxicol. Environ.
 Saf. 142, 200–206. https://doi.org/10.1016/j.ecoenv.2017.03.047
- Xue, S., Zhu, F., Kong, X., Wu, C., Huang, L., Huang, N., Hartley, W., 2016. A review of the characterization and revegetation of bauxite residues (Red mud). Environ. Sci. Pollut.
 Res. 23, 1120–1132. https://doi.org/10.1007/s11356-015-4558-8
- Yuan, Q., Wang, P., Wang, C., Chen, J., Wang, X., Liu, S., 2021. Indicator species and cooccurrence pattern of sediment bacterial community in relation to alkaline copper
 mine drainage contamination. Ecol. Indic. 120, 106884.
 https://doi.org/10.1016/j.ecolind.2020.106884
- Zech, W., Schad, P., Hintermaier-Erhard, G., 2014. Trockene Subtropen und Tropen, in:
 Zech, W., Schad, P., Hintermaier-Erhard, G. (Eds.), Böden der Welt: Ein Bildatlas.
 Springer, Berlin, Heidelberg, pp. 68–85. https://doi.org/10.1007/978-3-642-365751 7
- Zeng, J., Lou, K., Zhang, C.-J., Wang, J.-T., Hu, H.-W., Shen, J.-P., Zhang, L.-M., Han, L.L., Zhang, T., Lin, Q., Chalk, P.M., He, J.-Z., 2016. Primary Succession of Nitrogen
 Cycling Microbial Communities Along the Deglaciated Forelands of Tianshan
 Mountain, China. Front. Microbiol. 7, 1353.
 https://doi.org/10.3389/fmicb.2016.01353
- Zhang, D., Chen, H., Nie, Z., Xia, J., Li, E., Fan, X., Zheng, L., 2020. Extraction of Al and
 rare earths (Ce, Gd, Sc, Y) from red mud by aerobic and anaerobic bi-stage
 bioleaching. Chem. Eng. J. 401, 125914. https://doi.org/10.1016/j.cej.2020.125914
- Zhang, K., Shi, Y., Cui, X., Yue, P., Li, K., Liu, X., Tripathi, B.M., Chu, H., 2019. Salinity Is
 a Key Determinant for Soil Microbial Communities in a Desert Ecosystem. mSystems
 4, e00225-18. https://doi.org/10.1128/mSystems.00225-18
- 1273 Zhang, Y., Qin, W., Wang, J., Zhen, S., Yang, C., Zhang, J., Nai, S., Qiu, G., 2008.
 1274 Bioleaching of chalcopyrite by pure and mixed culture. Trans. Nonferrous Met. Soc.
 1275 China 18, 1491–1496. https://doi.org/10.1016/S1003-6326(09)60031-5
- 1276 Zhu, F., Xue, S., Hartley, W., Huang, L., Wu, C., Li, X., 2016. Novel predictors of soil
 1277 genesis following natural weathering processes of bauxite residues. Environ. Sci.
 1278 Pollut. Res. 23, 2856–2863. https://doi.org/10.1007/s11356-015-5537-9
- 1279