



HAL
open science

Contrasted microbial community colonization of a bauxite residue deposit marked by a complex geochemical context

Luis Alberto Macías-Pérez, Clément Levard, Mohamed Barakat, Bernard Angeletti, Daniel Borschneck, Laurent Poizat, Wafa Achouak, Melanie Auffan

► **To cite this version:**

Luis Alberto Macías-Pérez, Clément Levard, Mohamed Barakat, Bernard Angeletti, Daniel Borschneck, et al.. Contrasted microbial community colonization of a bauxite residue deposit marked by a complex geochemical context. *Journal of Hazardous Materials*, 2022, 424, Part B, pp.127470. 10.1016/j.jhazmat.2021.127470 . hal-03468413

HAL Id: hal-03468413

<https://cnrs.hal.science/hal-03468413v1>

Submitted on 16 Dec 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

1 ***“Contrasted Microbial community colonization of a bauxite residue deposit***
2 ***marked by a complex geochemical context.”***

3 Luis Alberto Macías-Pérez ^{a,b}, Clément Levard ^a, Mohamed Barakat ^b, Bernard
4 Angeletti ^a, Daniel Borschneck ^a, Laurent Poizat ^d, Wafa Achouak ^{b*}, Mélanie Auffan ^{a,c*}

5 *These authors contributed equally to the study

6 **Author contact information (FAMILY NAME, Name):**

7 1) MACÍAS-PÉREZ, Luis Alberto ^{a,b} (Corresponding author)

8 Address: CEREGE, Technopôle de l'Arbois-Méditerranée, BP80, 13545 Aix-en-
9 Provence, France

10 Email: macias@cerege.fr

11 Telephone number: +33 766887887 (France)

12 Fax: +33 (0)4 13 94 91 30 (France)

13 2) LEVARD, Clément ^a

14 Email: levard@cerege.fr

15 3) BARAKAT, Mohamed ^b

16 Email: mohamed.barakat@cea.fr

17 4) ANGELETTI, Bernard ^a

18 Email: angeletti@cerege.fr

19 5) BORSCHNECK, Daniel ^a

20 Email: borschneck@cerege.fr

21 6) POIZAT, Laurent ^d

22 Email: laurent.poizat@alteo-alumina.com

23 7) ACHOUAK, Wafa ^b

24 Email: wafa.achouak@cea.fr

25 8) AUFFAN, Mélanie ^{a,c}

26 Email: auffan@cerege.fr

27

28 **Author affiliations:**

29 a. Aix Marseille Université, CNRS, IRD, INRAE, Collège de France, CEREGE,
30 13545, Aix-en-Provence, France

31 b. Aix Marseille Univ, CEA, CNRS, BIAM, LEMIRE, Laboratory of Microbial Ecology
32 of the Rhizosphere, ECCOREV FR 3098, F-13108, St-Paul-lez-Durance, France

33 c. Civil and Environmental Engineering, Duke University, Durham, NC 27708, USA

34 d. ALTEO Alumina, 13120, Gardanne, France

35

36 **Keywords:** Bauxite residue, Primary succession, Microbial communities,
37 Physicochemical characterization, Critical metals.

38

39 **Abstract**

40 Bauxite residue is the alkaline byproduct generated during alumina extraction and is
41 commonly landfilled in open-air deposits. The growth in global alumina production
42 have raised environmental concerns about these deposits since no large-scale reuses
43 exist to date. Microbial-driven techniques including bioremediation and critical metal
44 bio-recovery are now considered sustainable and cost-effective methods to revalorize
45 bauxite residues. However, the establishment of microbial communities and their
46 active role in these strategies are still poorly understood. We thus determined the
47 geochemical composition of different bauxite residues produced in southern France
48 and explored the development of bacterial and fungal communities using Illumina
49 high-throughput sequencing. Physicochemical parameters were influenced differently
50 by the deposit age and the bauxite origin. Taxonomical analysis revealed an early-
51 stage microbial community dominated by haloalkaliphilic microorganisms and strongly
52 influenced by chemical gradients. Microbial richness, diversity and network complexity
53 increased significantly with the deposit age, reaching an equilibrium community
54 composition similar to typical soils after decades of natural weathering. Our results
55 suggested that salinity, pH, and toxic metals affected the bacterial community
56 structure, while fungal community composition showed no clear correlations with
57 chemical variations.

58

59

60

61

62

63

64

65

66

67

68

69 1. INTRODUCTION

70 Bauxite residue is a solid by-product generated during the production of alumina from
71 bauxite. Alumina extraction is most often performed by the Bayer process, where
72 bauxite is digested with large quantities of sodium hydroxide at temperatures between
73 150 and 250 °C (Evans, 2016). Depending on the process parameters and the origin
74 of the bauxite, 0.7 to 2 tons of bauxite residues are produced per ton of alumina
75 (International Aluminium Institute and European Aluminium, 2015). Over the last 10
76 years, the average annual production of alumina was 116 ± 15 Mt (World Aluminium,
77 2020), which corresponds to ~200 Mt of bauxite residue produced per year. In the
78 absence of economically profitable large-scale applications, bauxite residues are
79 commonly landfilled in large open-air deposit areas (BRDA), reaching a total of 4.5 Bt
80 bauxite residue in storage at the present time (Dentoni et al., 2021).

81 Due to the physicochemical characteristics of bauxite residue i.e., high alkalinity, high
82 salinity, high metal content and lack of nutrients, BRDAs pose a challenge to most
83 living organisms (Di Carlo et al., 2020; Santini et al., 2015b) and represent a potential
84 source of contamination for the surrounding terrestrial and aquatic ecosystems
85 (Bouchoucha et al., 2019; Ren et al., 2018). Consequently, efforts have been made to
86 remediate these deposits, usually by the addition of amendments (e.g. gypsum and
87 organic matter) that attenuate their harsh conditions prior to the revegetation of the
88 area (Bray et al., 2018; Khaitan et al., 2010). BRDAs have also been considered as a
89 promising secondary source of valuable and critical metals, as these elements are
90 concentrated by a factor of 2 in bauxite residues compared to the initial bauxite ore
91 (Panda et al., 2021; Ujaczki et al., 2017; Vind et al., 2018). Critical metals are chemical
92 elements characterized by their high economic relevance and supply risk (European
93 Commission, 2020) and include metals crucial for information and energy technologies
94 such as Co, Mg, Ba, V, Ge, Nb, Sr, Ga, platinumoids (PGMs), and rare earth elements
95 (REEs: Lanthanides (Ln), Y, Sc).

96 Recently, microbial-driven approaches have gained attention as viable and cost-
97 effective methods for the management and valorization of urban and industrial wastes,
98 including bauxite residue (Lyu et al., 2021; Panda et al., 2021; Santini et al., 2019).
99 During in-situ remediation processes, native microbial communities have been shown
100 to decrease the pH and salinity of the bauxite residue, as well as play an essential role
101 in soil formation and plant growth (Di Carlo et al., 2019; Tian et al., 2020; Wu et al.,

102 2019). Microbial metal recovery strategies are also promising for the dissolution and
103 recovery of elements of interest (e.g., Fe, Al, critical metals) from waste (Baniasadi et
104 al., 2019; Dev et al., 2020; Maes et al., 2016). Out of all these techniques, bioleaching
105 and microbial electrochemistry are among the most-studied methods (Dominguez-
106 Benetton et al., 2018; Srichandan et al., 2019). In bauxite residues, these techniques
107 are still at an early experimentation stage, although some promising results for REE
108 recovery by bioleaching have been reported (Kiskira et al., 2021; Qu et al., 2019;
109 Zhang et al., 2020).

110 More generally, microbial-driven strategies for waste management imply the use of
111 microbial communities native to contaminated sites, which tend to thrive better in such
112 harsh environments due to their unique metabolisms developed through natural
113 selection (Ghosh et al., 2018; Ma et al., 2019; Roy et al., 2018; Sajjad et al., 2020). In
114 addition, microbial communities are known to perform complex functions and are more
115 robust to environmental fluctuations compared to pure cultures (Perez-Garcia et al.,
116 2016; Wang et al., 2020; Zhang et al., 2008). Metagenomics studies based on Next
117 Generation Sequencing have significantly expanded the identification and dynamics
118 of microbial communities involved in different bioprocesses. However, the evolution of
119 microbial communities in metal biorecovery experiments is still limited to a few weeks
120 (Ma et al., 2017; Sajjad et al., 2020; Wang et al., 2020). Regarding bauxite residues,
121 the identification of pioneer microbial communities and their dynamics during primary
122 succession have been highlighted as research needs to improve microbial-driven
123 bioremediation (Santini et al., 2015a). Nevertheless, to date, studies on microbial
124 diversity in bauxite residue have focused on the responses of bacterial communities
125 to remediation strategies, overlooking their active role in the process (Banning et al.,
126 2011; Fourier et al., 2020; Ke, 2021; Krishna et al., 2014; Schmalenberger, 2013; Wu
127 et al., 2020).

128 This study aimed at exploring the microbial dynamics during primary succession in a
129 BRDA from Southern France using an integrated physicochemical and biological
130 approach. The specific objectives of this study were to (a) evaluate the effect of the
131 deposit age and the ore origin on the geochemical characteristics of bauxite residue;
132 (b) identify the first microbial communities colonizing BRDAs and their role in the
133 establishment of new species; (c) determine the microbial community structure in the
134 equilibrium stage of primary succession in BRDAs and the main geochemical factors

135 driving it; (d) explore possible implications of this study for microbial-driven
136 bioremediation and critical metal recovery.

137

138 **2. MATERIALS AND METHODS**

139 2.1 Site description and sampling

140 The bauxite residues come from a refinery operating the Bayer aluminum extraction
141 process since 1894 in Gardanne, Southern France. The sampling took place at the
142 Mange-Garri bauxite residue disposal area (BRDA) in Bouc-Bel-Air, Southern France.
143 This region is characterized by a hot-summer Mediterranean climate, with average
144 annual precipitation and temperature of 485 mm and 16.3°C (Copernicus Climate Data
145 Store, 2021). We selected four distinct areas of the BRDA based on their deposit age
146 (1, 2, 90, and 100 years) and bauxite origin (lateritic bauxite from Boké, Guinea (Bo)
147 and karstic bauxite from Provence, France (Pr)), named Bo1, Bo2, Pr90 and Pr100
148 (**Fig. 1**). At the Bo1 and Bo2 sites, bauxite residues were produced from Guinean
149 bauxite and were landfilled one and two years before sampling respectively. These
150 deposits are occasionally watered to prevent dust dispersion and were not amended
151 at the time of sampling. The Pr90 and Pr100 sites contain bauxite residues that were
152 produced from Provençal bauxite and were deposited around 90 and 100 years ago
153 respectively. These sites were amended with a soil layer in the 1960s and a slight
154 coverage with low-lying vegetation can be seen.

155 At each of the four sites, samples were collected in triplicate, gathering approximately
156 500 g of bauxite residue at a depth of 20 to 30 cm in sterile plastic bags. At sites Pr90
157 and Pr100, samples were collected at the edge of the deposit, where bauxite residues
158 were distant from the added soil layer. In addition, freshly produced bauxite residue (4
159 replicates) was also selected to analyze the initial bio-geochemical characteristics
160 before landfill. Samples were divided into two groups based on the following analyses.
161 Samples undergoing physicochemical analysis were oven-dried at 70°C, grounded in
162 a mortar, passed through a 130 µm sieve, and stored in metal-free plastic tubes.
163 Samples for microbial community analyses were stored at 4°C.

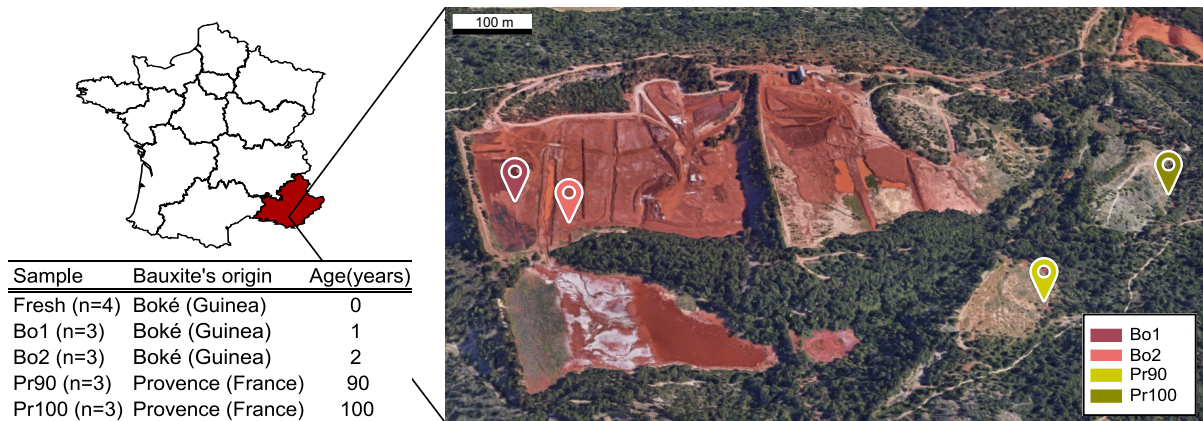


Figure 1. Sampling site and samples description.

164

165 2.2 Physicochemical analyses

166 2.2.1 Elemental composition

167 First, pH and electrical conductivity (EC) were measured at a solid/MilliQ water ratio
 168 of 1:5. Then, the elemental composition of the bauxite residues was determined after
 169 alkaline fusion of the samples (Rivera et al., 2019). Briefly, 1 g of each sample was
 170 digested with 750 mg of lithium tetraborate ($\text{Li}_2\text{B}_4\text{O}_7$) at 1000°C for 30 min (WiseTherm
 171 F/FH 0-1200°C) and then immediately dissolved in 40 mL of HNO_3 1N. Prior to
 172 elemental analysis, the samples were dissolved 200 times. Subsequently, inductively
 173 coupled plasma mass spectrometry (ICP-MS, Perkin Elmer Nexlon 300X) was used
 174 to analyze elements with concentrations up to $1000 \mu\text{g/L}$, while elements with
 175 concentrations above $1000 \mu\text{g/L}$ were analyzed by inductively coupled plasma optical
 176 emission spectroscopy (ICP-OES, Perkin Elmer 4300 DV). All analyses were
 177 conducted for major and trace elements commonly found in bauxite residues, namely
 178 Fe, Ti, Al, Ca, Zr, Cr, Mn, Zn, Th, Ni, Pb, Cu, Sn, Rb, Cs as well as critical raw elements
 179 (as defined by the European Commission (2020): P, Mg, V, Lanthanides (Ln), Sr, Y,
 180 Nb, Ba, Sc, Ga, Hf, Ge, Co, Sb, W, Ta). Total carbon (TC) and nitrogen were
 181 determined by dry combustion using an Elemental Analyzer (Flash EA, Thermo
 182 Scientific). To measure the total organic carbon (TOC), samples were treated with HCl
 183 to remove carbonates before analysis.

184 2.2.2 Mineralogical composition

185 The mineralogical analyses of bauxite residue were carried out by X-ray diffraction
 186 (XRD) using a PANalytical X'Pert Pro (Malvern Panalytical, UK) diffractometer
 187 equipped with a cobalt tube ($\lambda = 1.79 \text{ \AA}$) running at 40kV and 40mA. Samples were

188 deposited on low background silicon plates and analyzed from 8° to 80° (2θ) with a
189 step size of 0.033° and a total counting time of 7 hours. Samples were also spun at
190 15 rpm to improve statistics. Phase identification was performed using the X'pert
191 Highscore plus software (PANalytical) together with the PDF-2 ICDD database
192 (International Center for Diffraction data, Powder Diffraction Files 2). Profex software
193 (Doebelin and Kleeberg, 2015) was used for Rietveld refinement to semi-quantify the
194 proportions of the minerals in the bauxite residues. The following parameters were
195 refined: zero point shift, sample displacement, cell parameters, preferred orientation
196 and peak broadening resulting from the size of the crystallite and the micro strain.

197 2.3 DNA extraction, PCR amplification and sequencing

198 The total DNA was extracted from 5 g of each bauxite residue samples using the
199 FastDNA® Spin Kit for soil (MP Biomedicals, USA) following the manufacturer's
200 protocol. Extracted DNA was used as template in separate PCR reactions amplifying
201 the bacterial 16S and the fungal ITS rRNA gene sequences. For bacterial diversity
202 analysis, we used the primers 341F (5'-CCTAYGGGRBGCASCAG-3') and 806R (5'-
203 GGACTACNNGGGTATCTAAT-3'), targeting the V3 and V4 variable regions of the
204 bacterial 16S rRNA (Caporaso et al., 2011; Muyzer et al., 1993). For fungal diversity
205 analysis, we used the primers fITS7 (5'-GTGARTCATCGAATCTTTG-3') and ITS4 (5'-
206 TCCTCCGCTTATTGATATGC-3'), targeting the ITS2 region (Ihrmark et al., 2012).
207 The amplification conditions were as follows: initial denaturation at 95 °C for 2 min;
208 followed by 34 cycles of denaturation at 95 °C for 30 s, annealing at 55 °C for 30 s,
209 and extension at 72 °C for 1 min; and a final extension at 72 °C for 5 min (Muller et
210 al., 2021). PCR products were then purified using ProNex® Size-Selective Purification
211 System (Promega, USA) and sequenced on Illumina MiSeq platform (Biofidal, Vaulx-
212 en-Velin, France). The raw sequence reads generated from this study have been
213 deposited in the National Center for Biotechnology Information (NCBI) Sequence
214 Read Archive under the accession number PRJNA748554.

215 2.4 Sequencing data processing

216 Microbiome bioinformatics were performed by the open-source software QIIME2,
217 version 2019.10 (<https://qiime2.org>) (Bolyen et al., 2019). Raw reads were quality-
218 filtered, denoised and chimera-checked using DADA2 (Callahan et al., 2016). DADA2
219 uses a parametric model to infer true biological sequences from reads. The model

220 relies on input read abundances (true reads are likely to be more abundant) and the
221 pairwise similarity between sequences. The taxonomic annotation of the resulting
222 sequence variants (ASVs) was assigned using the feature-classifier command with
223 default parameters in QIIME2 and sequences were matched against the Greengenes
224 13_8 database (McDonald et al., 2012). Finally, scaling with ranked subsampling
225 (SRS) curves (Beule and Karlovsky, 2020) were drawn to determine whether the
226 sequencing depth was sufficient to represent the true diversity of the samples.

227 2.5 Biodiversity parameters and microbial biomarker discovery

228 Alpha diversity was explored through observed ASVs, Chao1, and Shannon. The
229 observed ASVs and the Chao1 estimator were selected to identify community
230 richness, and Shannon index was used to assess community diversity (Callahan et
231 al., 2017; Hill et al., 2003). Beta diversity analysis was used to evaluate distribution
232 patterns in samples based on bacterial and fungal ASV composition (Anderson et al.,
233 2011; Callahan et al., 2017). For this purpose, a principal coordinate analysis (PCoA)
234 based on weighted UniFrac distances (Lozupone et al., 2007) was conducted. All
235 alpha and beta diversity metrics were calculated using QIIME 2 after normalization to
236 59388 and 4778 sequences per sample for bacteria and fungi, respectively. Linear
237 discriminant analysis effect sizes algorithm (LEfSe) (Segata et al., 2011) was
238 performed on the Galaxy platform (<https://huttenhower.sph.harvard.edu/galaxy/>) to
239 identify bacterial biomarkers characterizing the samples. LEfSe couples Kruskal–
240 Wallis tests for measuring statistical significance with quantitative tests for biological
241 consistency (Wilcoxon rank sum test).

242 2.6 Statistical analysis

243 All statistical analyses were performed with the open-source software R (R Core
244 Team, 2020) using the packages “dplyr” (Wickham et al., 2021), “vegan” (Oksanen et
245 al., 2020), “car” (Fox and Weisberg, 2019), “ggpubr” (Kassambara, 2020), and “rstatix”
246 (Kassambara, 2021). Figures were produced with the package “ggplot2” (Wickham,
247 2009). To study the ore-dependent differences in chemical and microbiological
248 compositions of bauxite residue, unpaired two-sided T-tests and Wilcoxon rank sum
249 tests were conducted respectively. Furthermore, Games-Howell post-hoc tests were
250 used to assess the age-dependent variations in chemical and microbial compositions.
251 Pearson correlations between each variable and the deposit age were used to further

252 explore these relationships. Also, significant differences in alpha diversity indices were
253 tested by Wilcoxon rank sum tests.

254 Principal component analysis (PCA) was used to identify the variables that explain
255 most of the variation in chemical composition. In addition, unsupervised hierarchical
256 clustering was applied to the PCA to group the samples according to their chemical
257 similarity. Both PCA and hierarchical clustering were performed using the packages
258 “FactoMineR” (Lê et al., 2008) and “factoextra” (Kassambara and Mundt, 2020).
259 PERMANOVA via the *adonis* function was conducted in both PCA and PCoA to test
260 for the chemical and microbiological dissimilarities based on the deposit age and the
261 bauxite origin. To study the factor age alone, nested PERMANOVA were calculated
262 using the parameter *strata* to exclude the effect of bauxite origin. Furthermore, multiple
263 co-inertia analysis (MCIA) was performed using the package “omicade4” (Meng et al.,
264 2014) to determine the relationships between the four datasets used in this study
265 (chemical characteristics, mineralogy, 16S rRNA sequences and ITS sequences).

266 2.7 Co-occurrence network construction

267 Co-occurrence analyses were implemented for a better understanding of bacterial and
268 fungal interactions in the four bauxite residues. Co-occurrence networks were
269 constructed based on pairwise Pearson correlations calculated between bacterial and
270 fungal ASVs by using the base R function *cor* (Berry and Widder, 2014; Williams et
271 al., 2014). To avoid including false positives in the network due to spurious or random
272 interactions, the ASV table we permuted 100 times and a *p*-value for each possible
273 pairwise interaction was calculated to test its validity. The *p*-values were then adjusted
274 using the Benjamini-Hochberg procedure (Benjamini and Hochberg, 1995), and only
275 edges with a *p*-value below 0.01 that corresponded to an absolute correlation higher
276 than 0.4 were retained. To describe the topology properties of the networks, a set of
277 network indexes including graph density, average degree of nodes, transitivity,
278 modularity, average geodesic distance, betweenness centrality, and density were
279 calculated with the R package “igraph” (Csardi and Nepusz, 2006). The Wilcoxon test
280 was employed to assess significant differences in topological parameters between
281 networks. The connectivity of the network nodes was determined by their within-
282 module connectivity (Z_i) and among-module connectivity (P_i). Nodes were then
283 classified into four categories, according to Poudel et al. (2016): peripherals ($Z_i < 2.5$
284 and $P_i < 0.62$, nodes with few links to other species), connectors ($P_i > 0.62$, nodes

285 that connect modules), module hubs ($Z_i > 2.5$, highly connected nodes within
286 modules), and network hubs ($Z_i > 2.5$ and $P_i > 0.62$, highly connected nodes among
287 and within modules).

288

289 3. RESULTS

290 3.1. pH, EC and geochemical composition

291 The pH of the bauxite residue ranged from 8.70 ± 0.34 to 11.90 ± 0.01 , with the highest
292 value in fresh samples and the lowest in Pr100 (**Fig. 2a**, Table S1). The EC of the
293 bauxite residue ranged from 0.29 ± 0.15 mS/cm to 2.53 ± 0.04 mS/cm, and again,
294 fresh samples showed the highest value and Pr100 the lowest (**Fig. 2a**, Table S1).

295 The elements with the highest concentrations in all bauxite residue samples were Fe
296 (268894 ± 45706 mg/kg), Ti (51313 ± 17317 mg/kg), Al (33093 ± 5662 mg/kg), Ca
297 (22425 ± 4952 mg/kg), and TC (17162 ± 9439 mg/kg) (**Fig. 2b**, Table S2). Fe
298 represents at least 22 ± 3 % of the dried mass in all samples. The main critical
299 elements present in all samples were Ti (51313 ± 17317 mg/kg), P (2743 ± 1274
300 mg/kg), Mg (1877 ± 987 mg/kg), V (849 ± 423 mg/kg), Ln (775 ± 339 mg/kg), Sr (240
301 ± 142 mg/kg), Y (159 ± 59 mg/kg), Nb (99 ± 16 mg/kg), Sc (73 ± 12 mg/kg), Ba ($73 \pm$
302 15 mg/kg), and Ga (56 ± 24 mg/kg) (**Fig. 2c-d**, Table S2). TN concentrations were
303 only above the machine's limit of detection (100 mg/kg) in the Provence samples, with
304 values of 287 ± 183 mg/kg for Pr90 and 470 ± 158 mg/kg for Pr100 (Table S2).

305 Statistical analysis showed that Ti, Cr, V, and Ga concentrations were significantly
306 higher in samples coming from Boké bauxite ($p < 0.05$, Table S2), while samples from
307 Provence bauxite showed greater abundances of TC, TOC, P, Mg, Ln, Mn, Sr, Y, Sc,
308 Pb, Ni, and Co ($p < 0.05$, Table S2). Within these two groups, some elements
309 concentrations changed over time. In the Boké samples, a significant decrease in Fe,
310 Ca, Cr, V, Ga, Cu, Ge, and W was observed as sample age increased ($p < 0.05$, Table
311 S2) while Mg, and Sr concentrations increased with age ($p < 0.05$, Table S2). In the
312 Provence samples, Mg decreased significantly with age, while P and TOC
313 concentrations increased ($p < 0.05$, Table S2). These relationships between the age
314 of the sample and their chemical properties were further confirmed by Pearson
315 correlations ($R \geq 0.9$, $p < 0.05$, Fig. S1).

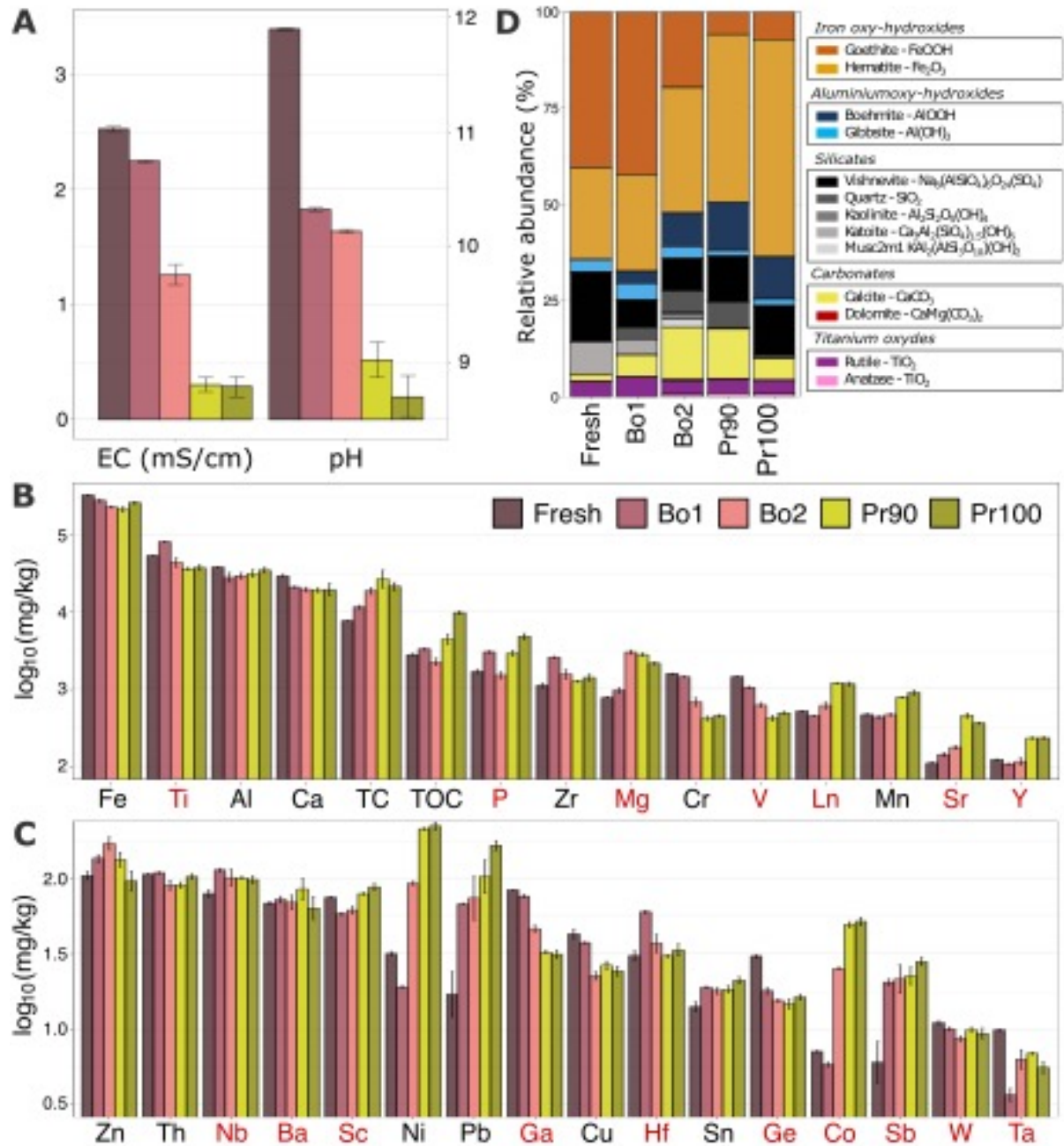


Figure 2. Physicochemical properties of bauxite residue samples. A) pH and EC, **B)** Major elements concentrations, **C)** Trace elements concentrations, and **D)** Mineralogy. Critical elements are colored red.

316

317 3.1.2 Element correlation

318 PCA was used to detect patterns in the chemical dataset and describe linear relations
 319 between the analyzed physicochemical parameters (**Fig. 3a**). The first three principal
 320 component axis (PC1, PC2, and PC3) explained 84.3% of the data variability between
 321 samples. PC1 explained 53.0% of the variance and included two groups: the first one

322 composed of pH, EC, Ga, Cr, V, Cu, Fe, Ge, Ca, Ti, and W whereas the second one
323 comprised TOC, Co, Ni, Sr, Ln, Mg, Pb, Mn, Y, Sb, Cs, and Sn. PC2 explained 18.4%
324 of the variance among samples and was mainly composed of Zr, Sc, Hf, Ta, Al, Zn,
325 Nb, and Ti. Finally, PC3 (not shown) explained 12.9% of variance and was primarily
326 composed of P, Th, Rb, Zn, Cs, and Sn. PERMANOVA analysis corroborated that the
327 factors “age” ($p < 0.01$) and “bauxite origin” ($p < 0.05$) significantly explained the
328 differences in chemical properties across samples. The factor “age” (tested by nested
329 PERMANOVA to exclude the effect of the bauxite origin) explained 65.8% of the
330 variance in chemical parameters while the factor “bauxite origin” explained 28.18%.
331 The hierarchical cluster analysis (Fig. S2) provided statistical confirmation of the
332 tendencies observed in the PCA and bauxite residue samples were clustered based
333 on their chemical similarity. The first cluster contained the fresh samples, the second
334 cluster Bo1 samples, the third cluster Bo2 samples, and the fourth cluster contained
335 both Pr90 and Pr100 samples.

336 To further understand the effect of the factor “age” on residues generated from the
337 same bauxite ore, a second PCA was conducted considering only the Boké samples
338 (fresh, Bo1 and Bo2) (**Fig. 3b**). The first two dimensions explained 82.5% of the
339 variability and the factor “age” separated the samples significantly (PERMANOVA, p
340 < 0.01). Co, Sr, Pb, Mg, and Sb were mostly associated with Bo2, while pH, EC, W,
341 Ca, Ge, Fe, Cu, V, Cr, and Ga had their highest values in fresh samples and decreased
342 progressively with age. In contrast, TOC, Y, Mn, Ln, Ni, Sn, Al, and Ta did not show
343 any age-dependent trends during the first years of storage (Table S2). Globally, this
344 statistical analysis is in line with the expected behavior of the elements, such as the
345 colocalization of Cr and V and their correlation with Fe (Markus Gräfe et al., 2011) as
346 well as the REEs affinity for Mn phases (Vind et al., 2018).

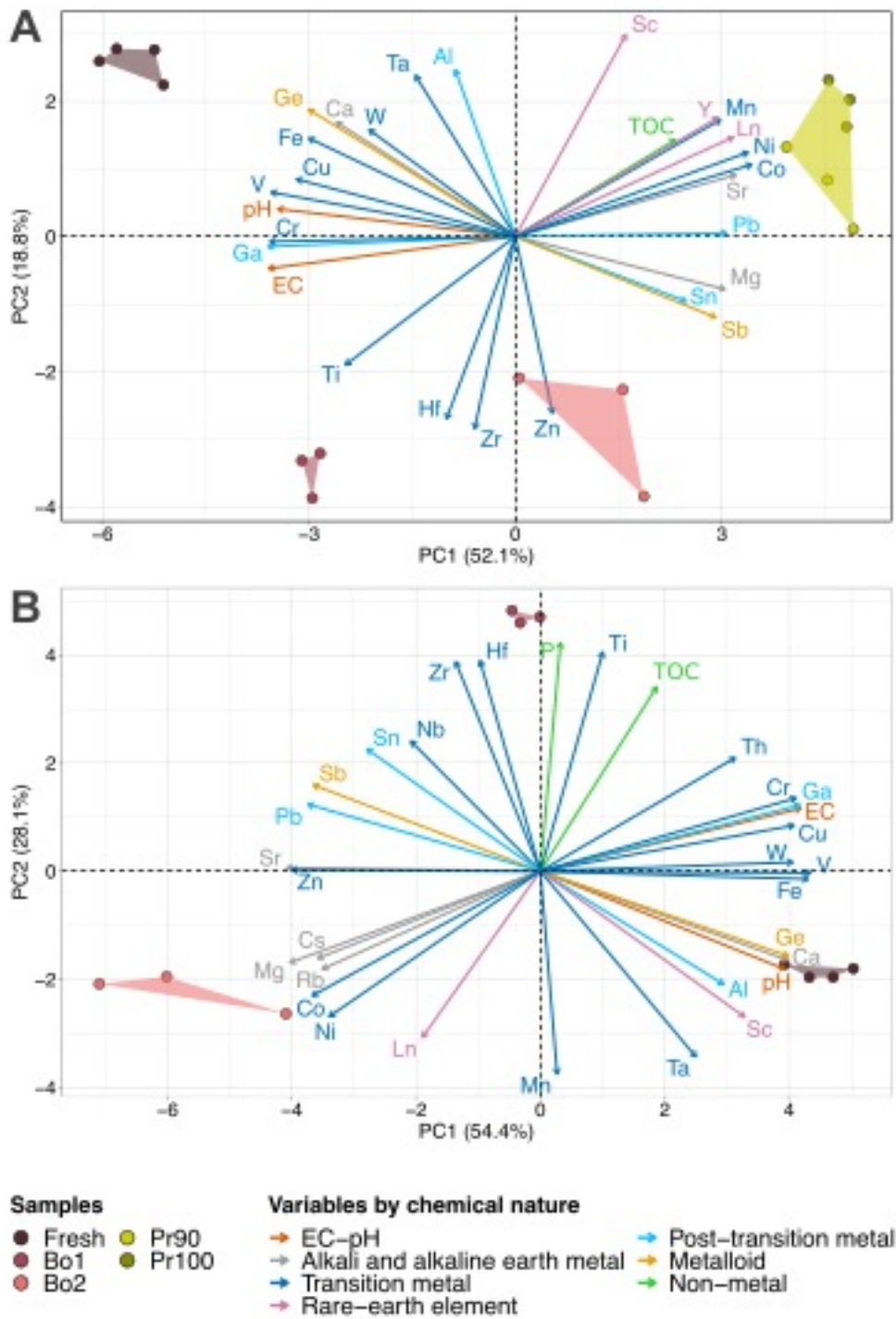


Figure 3. Principal component analysis (PCA) of chemical parameters in A) all and B) Boké bauxite residue samples. Individuals are represented by dots and colored by sample. Individuals are grouped according to hierarchical clustering results. Variables are represented by arrows and colored by chemical nature. Only variables with a $\cos^2 > 0.5$ are shown.

347

348

349 **3.1.3 Mineralogical composition**

350 XRD analysis (**Fig. 2d**, Table S3) showed that bauxite residues were principally
351 composed of iron and aluminum oxides, hydroxides, and silicates. Goethite
352 (FeO(OH)), hematite (Fe₂O₃), gibbsite (Al(OH)₃), and vishnevite
353 (Na₈(AlSiO₄)₆O₂₄(SO₄)) are the dominant phases and accounted for at least 70 % of
354 the crystalline fraction in all the samples. Rutile (TiO₂) and calcite (CaCO₃) were also
355 detected in all samples, in the range of 3-5 % and 2-14% respectively. The hematite,
356 anatase (TiO₂) and boehmite (AlO(OH)) were significantly more present in Provence
357 samples, and interestingly, neither anatase nor boehmite were found in fresh samples
358 (Table S3). In Boké samples, a significant decrease in goethite and Katoite
359 (Ca₃Al₂(SiO₄)_{1.5}(OH)₆) was observed as the age of the samples increased.

360 **3.2 Microbial community diversity, structure and dynamics**

361 **3.2.1 Alpha diversity**

362 The influence of bauxite residue's age and geochemical composition on microbial
363 communities was assessed using 16S and ITS metabarcoding. As expected, no DNA
364 could be extracted from the fresh residue samples. For Bo1, Bo2, Pr90, and Pr100,
365 the sequencing achieved a coverage of more than 55000 16S rRNA sequences and
366 4700 ITS sequences per sample. The SRS curves (Fig. S3) indicated that the
367 sequencing depth was sufficient to identify the majority of ASVs within bacterial and
368 fungal communities in all the samples (except fresh samples). Alpha-diversity derived
369 from the number of ASVs, Chao1 richness, and Shannon's index showed differences
370 between samples regarding bacterial and fungal communities (Table S4). A total
371 number of 7994 bacterial ASVs were identified, with 350 to 1318 ASVs per
372 sample. The number of bacterial ASVs in the Provence samples (Pr90 and Pr100)
373 was 6928, significantly greater than the 2867 ASVs found in the Boké samples (Bo1
374 and Bo2). The Provence samples also showed the highest richness and Shannon
375 values for bacteria, compared with Boké samples (**Fig. 4a**). For fungi, 892 ASVs
376 ranging from 45 to 156 ASVs per sample were obtained, with again more ASVs and
377 Chao1 richness in Provence samples. However, there were no significant changes in
378 Shannon's index between both group of samples, indicating that the specific diversity
379 of the fungal communities was similar in Provence and Boké samples.

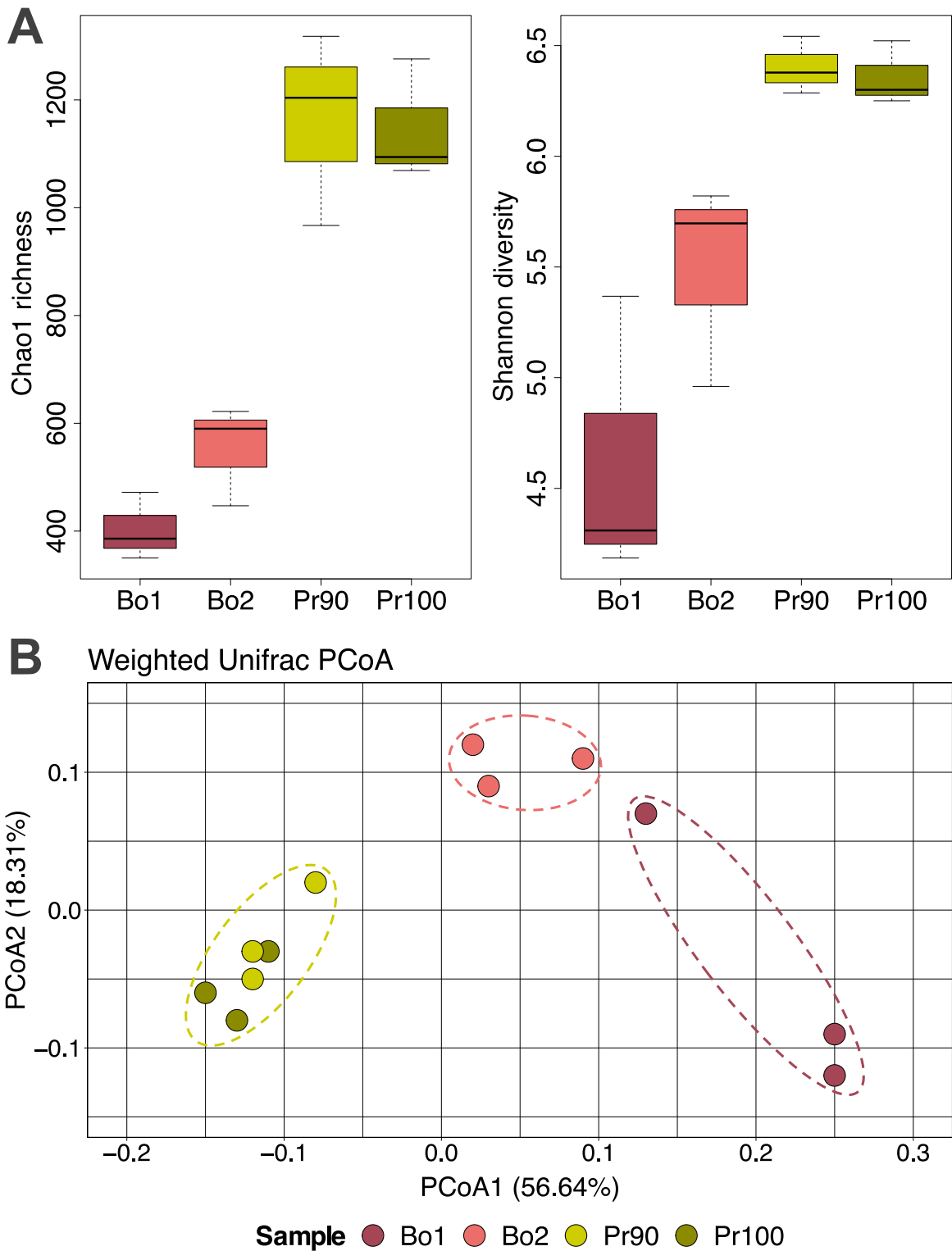


Figure 4. Bacterial alpha and beta diversity of bauxite residues. A) Chao1 richness and Shannon diversity index. B) Principal coordinate analysis (PCoA) based on weighted UniFrac distances between bauxite residue samples.

381 3.2.2 Microbial community structure

382 PCoA ordinations based on weighted UniFrac distances (**Fig. 4b**) revealed that the
383 bacterial community structure differed between the bauxite residue samples. Samples
384 were significantly separated based on bauxite origin (PERMANOVA, $p < 0.05$), and
385 nested PERMANOVA confirmed the difference in bacterial community structure
386 across age ($p < 0.001$). The first two axes of PCoA explained 74.95% of the community
387 dissimilarity. Axis 1 was most correlated with EC ($R = 0.97$, $p = 1.1e-07$), followed by
388 Co ($R = 0.96$, $p = 9.8e-07$), Ni ($R = 0.95$, $p = 1.4e-06$), Cr ($R = 0.94$, $p = 1.1e-07$), V
389 ($R = 0.92$, $p = 2.5e-05$), Ti ($R = 0.9$, $p = 5.4e-05$), and pH ($R = 0.9$, $p = 8.1e-05$); while
390 Axis 2 correlated with P ($R = 0.75$, $p = 0.005$). In contrast, the fungal community
391 structure did not seem to follow a clear pattern among samples according to PCoA
392 built on Bray-Curtis distances (Fig. S4).

393 The bacterial phyla that dominated all bauxite residue samples were Proteobacteria,
394 Actinobacteria, Chloroflexi, Bacteroidetes, Gemmatimonadetes, Planctomycetes,
395 Verrucomicrobia, Acidobacteria and Firmicutes, accounting for more than 90% of the
396 total bacterial communities (**Fig. 5**). When comparing the samples from Provence and
397 Boké, significant taxonomic differences were observed. Actinobacteria and
398 Gammaproteobacteria were the dominant taxa in the Boké samples with an average
399 abundance of 28.0 % and 21.3 % respectively, followed by Chloroflexi (9.8 %) and
400 Bacteroidetes (9.8 %). In the Provence samples, Actinobacteria (18.4 %) were still
401 dominant while Gammaproteobacteria (4.8 %) and Bacteroidetes (3.6 %) lost
402 prominence in favor of Alphaproteobacteria (12.0 %), Planctomycetes (9.6 %),
403 Betaproteobacteria (6.3 %), and Deltaproteobacteria (3.2 %). Boké samples were
404 significantly enriched in Actinobacteria, Gammaproteobacteria, Bacteroidetes and
405 Firmicutes, whereas samples from Provence showed greater abundances of
406 Alphaproteobacteria, Planctomycetes, Betaproteobacteria and Deltaproteobacteria (p
407 < 0.05). Acidobacteria (10.3 %), barely found in Boké samples, showed a strong
408 increase in Provence samples ($p < 0.05$). Chloroflexi, Gemmatimonadetes and
409 Verrucomicrobia were nearly constant across samples.

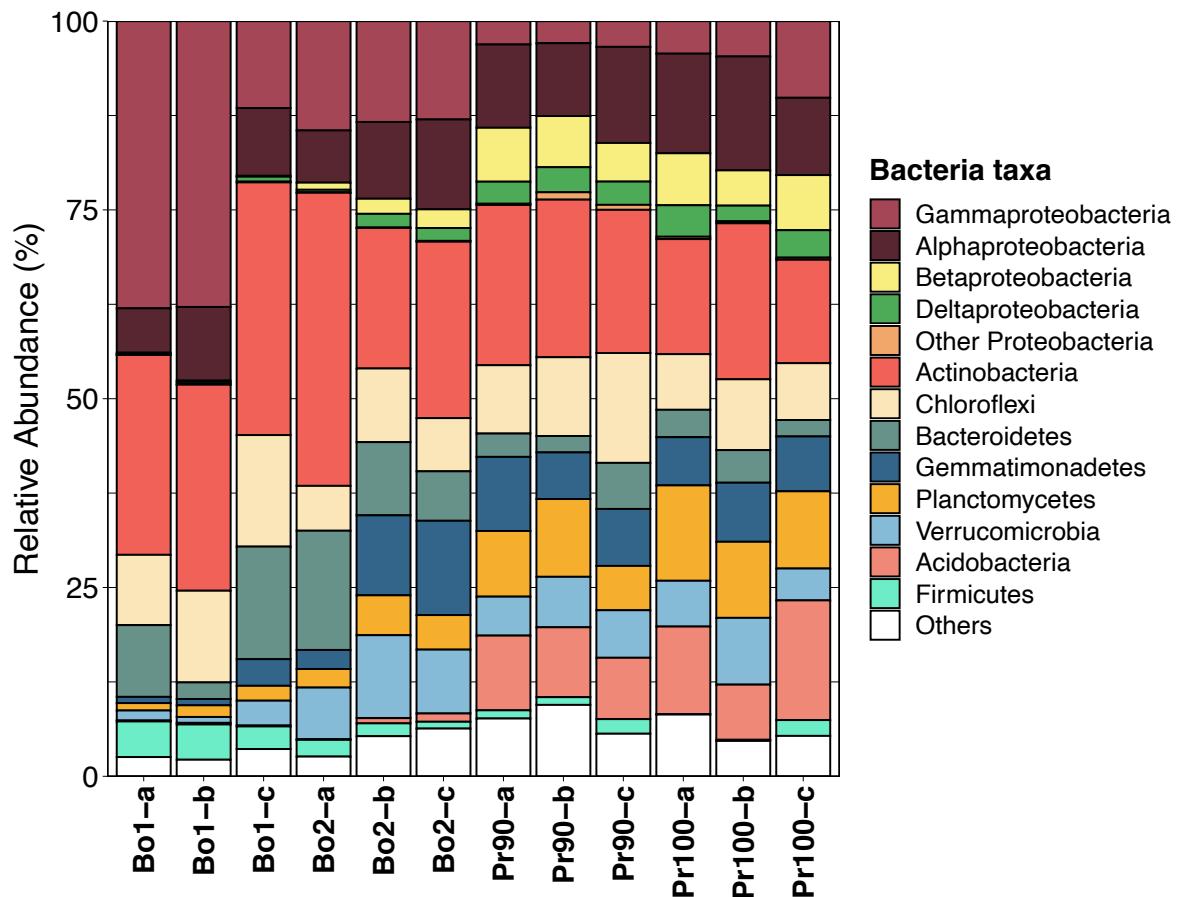


Figure 5. Relative abundance of the most abundant bacterial taxa among bauxite residue samples. Proteobacteria are divided in classes.

410

411 At the level of bacterial classes, differences between Boké and Provence samples
 412 were also found (**Fig. 6**). Significantly higher proportions of Gammaproteobacteria,
 413 Nitrospirae, Thermomicrobia, Gemmatimonadetes, Bacilli, Rhodothermi, TK17, Deinococci and
 414 Clostridia were identified in the Boké samples ($p < 0.05$). Within this group, variations
 415 between samples of different ages were also observed. Bo1 was enriched in
 416 Thermomicrobia, Thermoleophilia, Acidimicrobia and Bacilli, whereas Ophitales,
 417 Verrucomicrobiae and Flavobacteriia showed greater abundances in Bo2 ($p < 0.05$).
 418 In contrast, Provence samples were significantly more enriched in
 419 Alphaproteobacteria, Planctomycetes, Acidobacteria6, Betaproteobacteria,
 420 Thermoleophilia, Deltaproteobacteria, Gemmatimonadetes, Pedosphaerae,
 421 Anaerolineae, Phycisphaerae, Spartobacteria, Gemm1, Chloracidobacteria,
 422 Ellin6529, S085, Nitrospira, Saprospirae, TK10, Solibacteres and TM71 ($p < 0.05$). In
 423 the Provence samples, the only differences seen over ages were Pedosphaerae and
 424 Actinobacteria, which were more abundant in Pr90 than in Pr100 ($p < 0.05$).

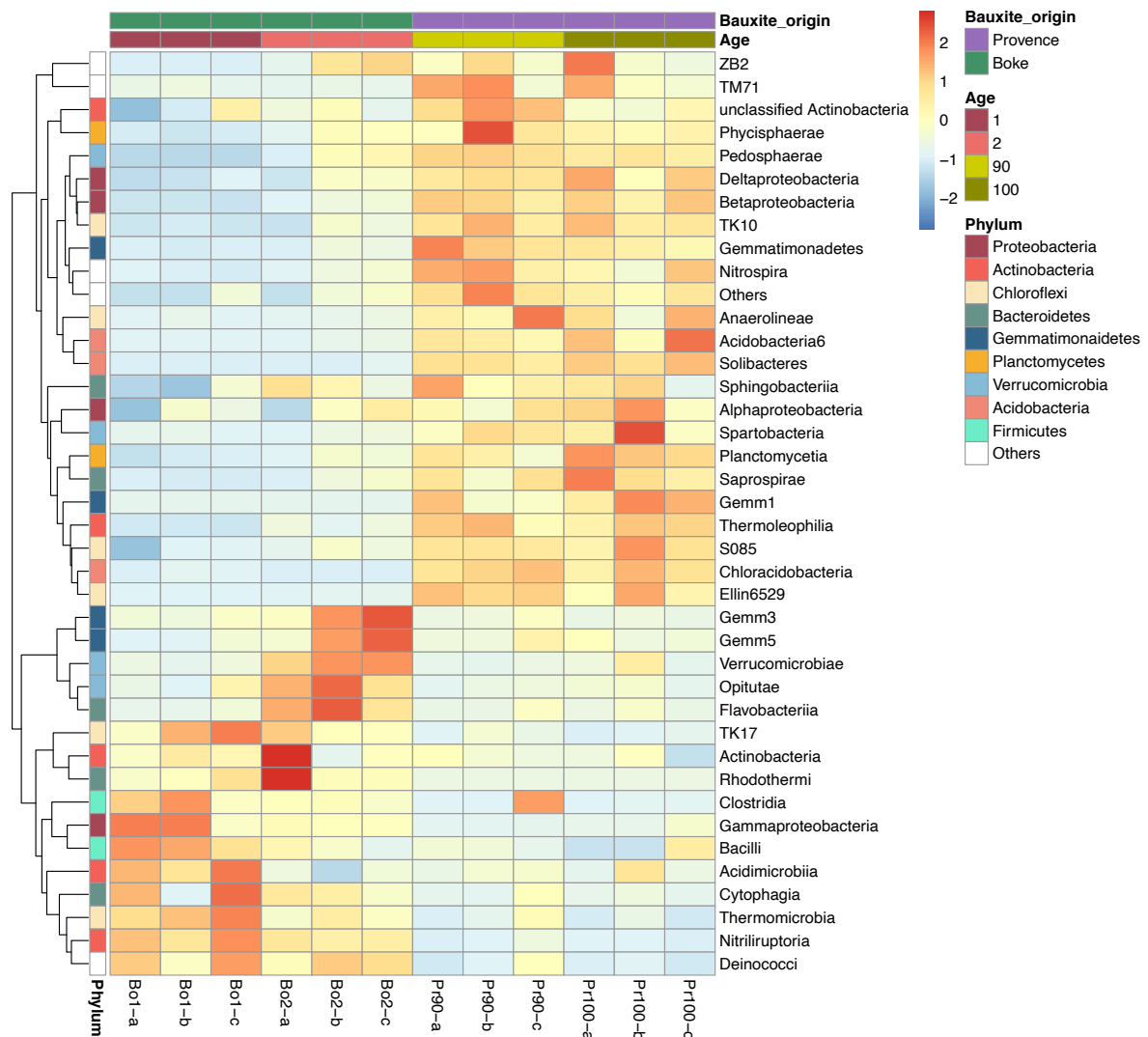


Figure 6. Z-score hierarchical clustering heat map representing the relative abundance of the top 40 bacterial classes in different bauxite residue samples.

425

426 3.2.3 Taxonomic characteristics of bacterial and fungal communities

427 To further explore these temporal scale differences, LEfSe analyses were conducted
 428 to detect asymmetrically distributed bacterial taxa. Among the Boké samples, 9
 429 indicator bacterial taxa were found in Bo1 against 37 in Bo2 (**Fig. 7a**). LEfSe confirmed
 430 the highest abundance of three families of Bacilli in Bo1, as well as other taxa
 431 belonging to the phyla Actinobacteria and Proteobacteria, and the archaeal class
 432 Halobacteria. In line with our previous results, Opitutae, Verrucomicrobiae and
 433 Flavobacteriia were significantly more abundant in Bo2, together with other taxa
 434 belonging to the phyla Actinobacteria, Proteobacteria, Firmicutes, Bacteroidetes,
 435 Planctomycetes and Nitrospirae. For the Provence samples, only two indicator

436 bacterial taxa (classified as Bacteroidetes) were identified in Pr90, against 28 in Pr100
437 (classified as Actinobacteria, Proteobacteria, Planctomycetes, Bacteroidetes,
438 Verrucomicrobia, and Chloroflexi) (**Fig. 7b**).



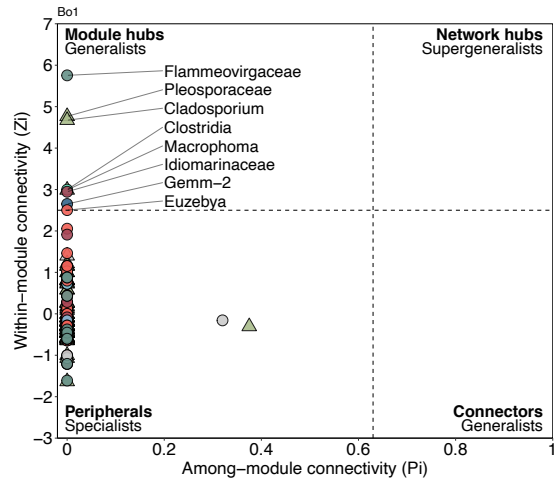
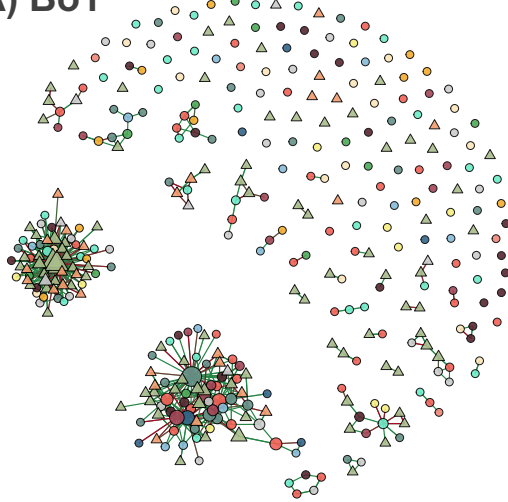
Figure 7. LEfSe analysis of in the bauxite residue samples according to sample age. Histograms of LDA scores of 16S gene sequences in Boké (A) and Provence (C) samples. Only taxa with a LDA score (log10) above 2.0 and a p-value lower than 0.05 for Kruskal–Wallis tests are shown. B and D Cladograms are derived from LEfSe analysis. The central point denotes the root of the tree and expanded to each ring representing the next lower taxonomic level from phylum to genus.

440 **3.2.4 Microbial co-occurrence networks**

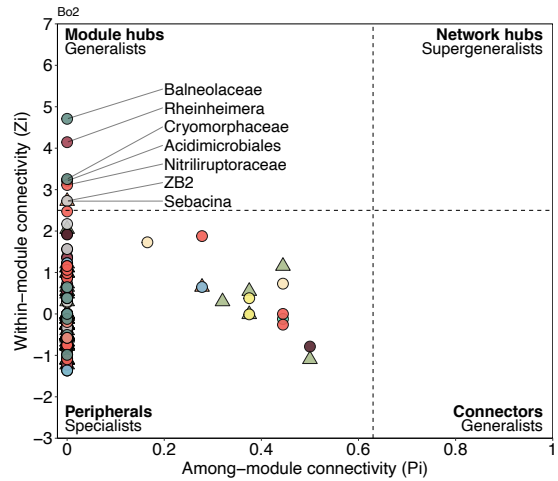
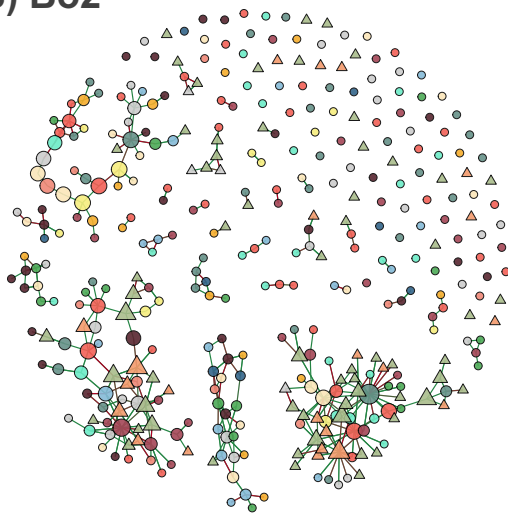
441 Co-occurrence networks were constructed using the ASV data of 16S rRNA and ITS
442 sequences. The Boké (**Fig. 8**) and Provence (**Fig. 9**) samples did not differ significantly
443 in network topological parameters such as average geodesic distance and modularity
444 (Table S5). However, the network complexity (number of nodes and edges, degree of
445 nodes, and transitivity) was different between the Boké and Provence networks (Table
446 S5). The two Boké networks (Bo1 and Bo2) contained a similar number of nodes (366
447 and 369), although Bo1 showed more edges (516 versus 324), higher average degree
448 (2.46 versus 1.76), and transitivity (0.15 versus 0.11). The two Provence networks
449 were more complex, as both Pr90 and Pr100 showed more nodes (603 and 571), more
450 edges (1254 and 1322), higher average degree (4.16 and 4.63), and higher transitivity
451 (0.18 and 0.12) compared to Boké.

452 Most of the nodes from the Boké and Provence networks were classified as
453 peripherals (specialists), and few nodes fell into module hubs (generalists) (**Fig. 8-9**).
454 Generalists are considered keystone taxa, as they are responsible for structuring the
455 different nodes and modules into a complete community, thus determining the
456 efficiency of energy metabolism and nutrient cycling in habitats (Wang et al., 2019). In
457 this study, more module hubs (generalists) were identified in Pr90 and Pr100 than in
458 Bo1 and Bo2. In Bo1 and Bo2 networks, the module hubs were mainly members of
459 Bacteroidetes, Actinobacteria, Proteobacteria, and Firmicutes for bacteria, and
460 Ascomycota for fungi. In contrast, the module hubs identified in Pr90 and Pr100
461 networks belonged to Proteobacteria, Actinobacteria, Chloroflexi, Verrucomicrobia,
462 Gemmatimonadetes, Planctomycetes and Firmicutes for bacteria, and Ascomycota
463 and Basidiomycota for fungi.

A) Bo1



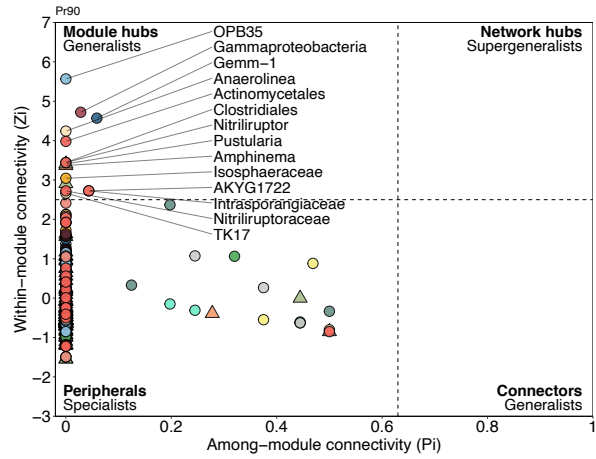
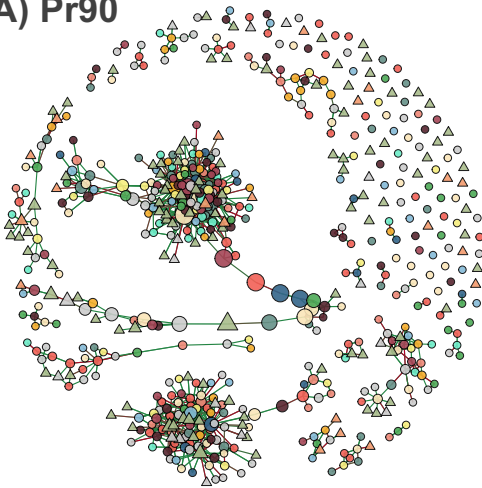
B) Bo2



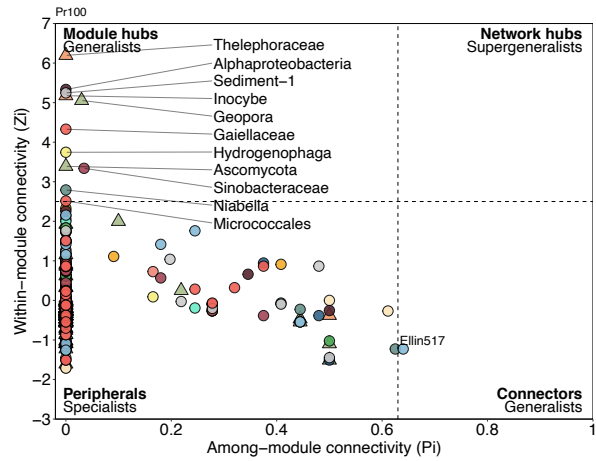
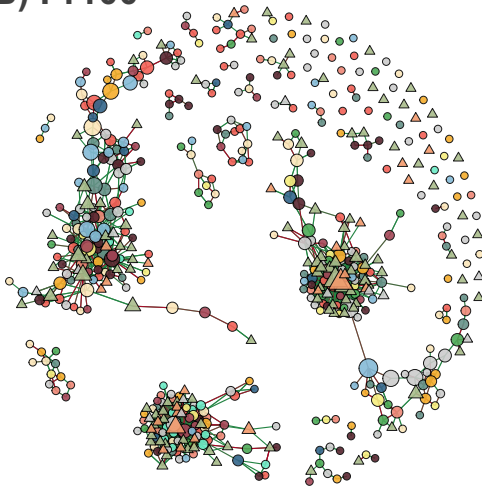
- | | | | |
|------------------------|--------------------|-------------------|-----------------|
| ● Gammaproteobacteria | ● Actinobacteria | ● Verrucomicrobia | ▲ Ascomycota |
| ● Alphaproteobacteria | ● Chloroflexi | ● Acidobacteria | ▲ Basidiomycota |
| ● Betaproteobacteria | ● Bacteroidetes | ● Firmicutes | ▲ Other Fungi |
| ● Deltaproteobacteria | ● Gemmatimonadetes | ● Other Bacteria | |
| ● Other Proteobacteria | ● Planctomycetes | | |

Figure 8. Co-occurrence network and Zi-Pi plot of Bo1 (A) and Bo2 (B) bauxite residue samples. Each node represents a bacterial class (circles) or a fungal phyla (triangles). Nodes are colored by taxonomical affiliation and their size in the network is proportional to the node betweenness. The color of each link reflects positive (red) or negative (blue) interactions. The topological role of each node is defined by within-module connectivity (Z_i) and among-module connectivity (P_i). According to values of Z_i (2.5) and P_i (0.62), the roles of nodes are classified into four categories: Peripherals, Module hubs, Connectors and Network hubs.

A) Pr90



B) Pr100



- | | | | |
|------------------------|--------------------|-------------------|-----------------|
| ● Gammaproteobacteria | ● Actinobacteria | ● Verrucomicrobia | ▲ Ascomycota |
| ● Alphaproteobacteria | ● Chloroflexi | ● Acidobacteria | ▲ Basidiomycota |
| ● Betaproteobacteria | ● Bacteroidetes | ● Firmicutes | ▲ Other Fungi |
| ● Deltaproteobacteria | ● Gemmatimonadetes | ● Other Bacteria | |
| ● Other Proteobacteria | ● Planctomycetes | | |

Figure 9. Co-occurrence network and Zi-Pi plot of Pr90 (A) an Pr100 (B) bauxite residue samples. Each node represents a bacterial class (circles) or a fungal phyla (triangles). Nodes are colored by taxonomical affiliation and their size in the network is proportional to the node betweenness. The color of each link reflects positive (red) or negative (blue) interactions. The topological role of each node is defined by within-module connectivity (Zi) and among-module connectivity (Pi). According to values of Zi (2.5) and Pi (0.62), the roles of nodes are classified into four categories: Peripherals, Module hubs, Connectors and Network hubs.

465

466

467

468

469 3.2.5 Multi co-inertia analysis

470 Multiple co-inertia analysis (MCIA) was used to determine the relationships between
471 the chemistry, mineralogy, bacterial and fungal community composition in our bauxite
472 residue samples. **Figure 10** displays the projection of the four datasets onto the first
473 two principal components (PCs) of MCIA for Bo1, Bo2, Pr90 and Pr100 samples. PC1
474 and PC2 explained 54% and 19% of the total variation respectively (Fig. S6b). The
475 sample space (Fig. S6a) shows a good correlation between the different datasets in
476 all samples, which could be clearly differentiated. Early-stage bauxite residues (Bo1
477 and Bo2) were separated from equilibrium-stage samples (Pr90 and Pr100) along the
478 PC1. Pseudo-eigenvalues (Fig. S6c) indicated that the chemistry and bacterial dataset
479 contributed the most to the variation observed in PC1, while mineralogy was weighted
480 in PC2. The pair-wise RV coefficient, which is multivariate generalization of the
481 squared Pearson correlation coefficient, indicated higher global similarity between
482 chemistry and bacterial classes (RV score = 0.86) when compared to the similarities
483 between the chemistry and fungal classes (RV score = 0.36), and between the
484 mineralogy and microbial classes (RV score for bacterial classes = 0.41, RV score for
485 fungal classes = 0.22). Fungal classes contributed little to the variation of PC1 or PC2
486 and were mostly correlated with bacterial classes (RV score = 0.38).

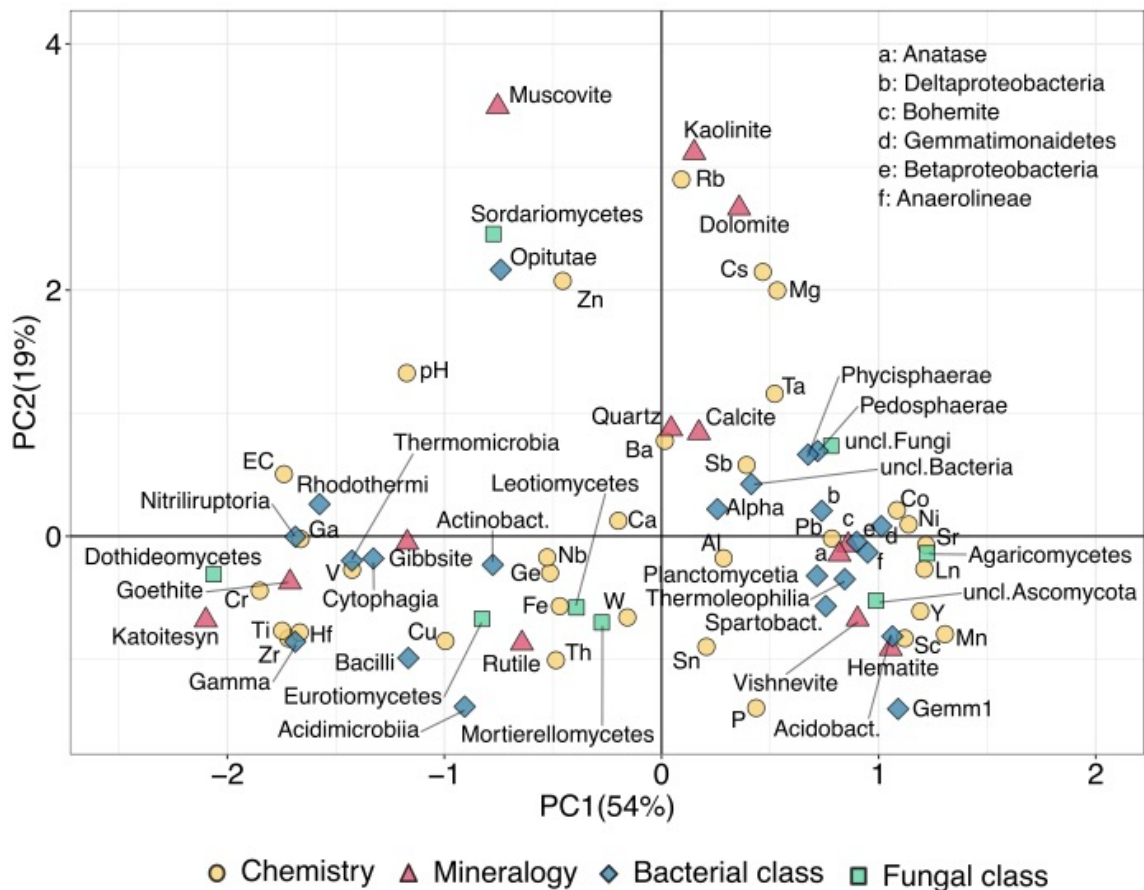


Figure 10. MCA analysis of bauxite residue samples. The figure shows the coordinates of the different data sets represented by shapes (circle: chemical data set; triangle: mineralogical data set; square: fungal classes data set; diamond: bacterial classes data set). Only bacterial and fungal classes representing at least 5% of the abundance in at least one sample are shown. Uncl. = unclassified; Alpha = Alphaproteobacteria; Gamma = Gammaproteobacteria.

487

488

489 4. DISCUSSION

490 4.1 BRDAs behave as primary successional environments

491 The initial geochemical conditions of the fresh bauxite residue, characterized by high
 492 alkalinity ($\text{pH} = 11.90 \pm 0.01$) and salinity ($\text{EC} = 2.53 \pm 0.04 \text{ mS cm}^{-1}$) in line with
 493 previous studies (Cusack et al., 2019; Xue et al., 2016), make BRDAs hostile
 494 environments for most domains of life. The high pH and salinity derive from the excess
 495 of alkaline anions (hydroxides, carbonates, aluminates, aluminum hydroxides,
 496 silicates) and Na^+ produced during the Bayer process (Cusack et al., 2019; X. Kong
 497 et al., 2017) and thus the bauxite origin does not seem to significantly influence the
 498 value of these parameters. Fresh and recently deposited bauxite residues (Bo1 and

499 Bo2) also contained significant amounts of toxic metals for microorganisms, namely
500 Cr (1594 ± 23 mg/kg), V (1452 ± 10 mg/kg) and Al (38420 ± 920 mg/kg). The high
501 alkalinity (pH > 10) of fresh, Bo1 and Bo2 bauxite residues may favor the Cr (III)
502 oxidation to Cr(IV) as well as the predominance of V(V), both highly mobile and toxic
503 for bacteria, fungi and plants (Chen et al., 2021; Economou-Eliopoulos et al., 2016;
504 Liang et al., 2021; Milačić et al., 2012; Xiao et al., 2017). Regarding aluminum, only
505 1% of the total Al content in bauxite residue has been reported to be soluble at pH
506 values above 9, mainly as mobile $[\text{Al}(\text{OH})_4]^-$ (Milačić et al., 2012). Nevertheless, its
507 impact on microbial and plant communities should not be neglected, as even small
508 proportions of the total Al correspond to potentially toxic concentrations. Moreover,
509 fresh and recently deposited bauxite residues were inherently poor in most
510 macronutrients required for microbial and plant growth, such as TOC, TN, and P
511 (Table S2). TOC accounted for less than 0.6% of the mass (0.28 % in fresh residue,
512 0.33 % in Bo1, and 0.23 % in Bo2) which is typically attributed to desert soils (Li et al.,
513 2013; Zech et al., 2014). TN concentrations were below the detection limit (100 mg/kg)
514 in fresh, Bo1 and Bo2 samples in agreement with previous studies (Krishna et al.,
515 2014; Wu et al., 2020). Finally, although the P concentrations ranged from 1500 to
516 3000 mg/kg in fresh, Bo1, and Bo2 samples, the available phosphorus in bauxite
517 residue is usually around 6.7 mg/kg (Santini et al., 2015a; Wu et al., 2020).

518 These extreme conditions, together with the absence of pre-existing life in the freshly
519 produced bauxite residues, indicate that BRDAs behave as primary successional
520 environments, similar to other well-studied natural systems (Guo et al., 2014; Schmidt
521 et al., 2008). Primary succession is characterized by an initial short phase where the
522 microbial community assembly is driven by stochastic dispersal processes, followed
523 by a longer deterministic phase dominated by environmental selection and competition
524 between species (Ortiz-Álvarez et al., 2018). In engineered environments like BRDAs,
525 the dispersal of microorganisms is limited, and environmental selection pressure is
526 usually stronger due to the extreme chemical characteristics of these sites. Therefore,
527 a shorter stochastic phase and a stronger deterministic assembly is expected in such
528 systems (Santini et al., 2015a). Given the environmental stress imposed by the
529 geochemical properties of fresh bauxite residues, the first pioneer microorganisms to
530 colonize BRDAs are likely to be haloalkaliphilic species able to accumulate nutrients

531 to facilitate the incorporation of new species (Santini et al., 2015a; Schmidt et al., 2014;
532 Sun et al., 2018).

533 **4.2 Geochemistry-dependent microbial community assembly in the early stages** 534 **of primary succession**

535 The bacterial community composition of the recently deposited bauxite residues (Bo1
536 and Bo2) was similar to that observed in soda lakes, alkaline residues, and other
537 unamended BRDAs (Bondici et al., 2013; Chakraborty et al., 2021; Kalwasińska et al.,
538 2017; Ohlsson et al., 2019), characterized by the dominance of Actinobacteria
539 (Actinobacteria, Nitriliruptoria, Acidimicrobiia) and Proteobacteria (especially
540 Gammaproteobacteria) and relatively high proportions (1 to 16 %) of Bacteroidetes
541 (Cytophagia, Rhodothermi, Flavobacteriia), Firmicutes (Bacilli, Clostridia), and other
542 alkalophilic classes (Opitutae, Deinococci) (**Fig. 5-6**).

543 The phyla Actinobacteria, Proteobacteria and Firmicutes contain some of the
544 prokaryotes best adapted to high alkalinity and salinity. Alkaliphilic and alkalitolerant
545 microorganisms maintain their intracellular pH homeostasis by accumulating H⁺ in
546 their cytoplasm through Na⁺/H⁺ antiporters (Mamo, 2020). Our results showed that
547 Halobacteria (Archaea), Gammaproteobacteria, Nitriliruptoria and Rhodothermi were
548 more abundant in Bo1 (**Fig. 6-7**) and correlated with EC (**Fig. 10**), corroborating the
549 preference of these classes for high-salinity niches (Zhang et al., 2019). Extreme
550 halophiles accumulate intracytoplasmic KCl to regulate their osmotic potential, which
551 constrains them to be obligate halophiles and limits their dispersal (Chen et al., 2020;
552 Shameer, 2016; Vaidya et al., 2018). Salinity (EC) decreased significantly from Bo1 to
553 Bo2 (**Fig. 1**, Table S2), allowing the proliferation of a wider range of halophilic and
554 halotolerant species such as Opitutae, Acidimicrobiia or the genus *Halomonas*
555 (Gammaproteobacteria) (**Fig. 6-7**). Halotolerant bacteria can flexibly adapt to habitats
556 with different salinities by synthesizing and accumulating compatible solutes in their
557 cytoplasm, which in turn requires a more intensive use of energy (Gunde-Cimerman
558 et al., 2018). Interestingly, the number of edges and network complexity was higher in
559 Bo1 compared with Bo2 (**Fig. 8**), indicating that saline stress promoted microbial
560 interaction as reported in other studies (Ji et al., 2019; Wang et al., 2019).

561 As pioneer colonizers, microorganisms are critical in the biogeochemical cycles and
562 the ensuing development of soil (Schmidt et al., 2008; Zeng et al., 2016). TOC

563 concentration showed no significant trend in early-stage samples (**Fig. 2**, Table S2),
564 suggesting that the ratio in TOC input and consumption was balanced. In primary
565 successional environments, the main sources of TOC are airborne allochthonous
566 organic matter and C fixed by autotrophic microbial species (Ciccazzo et al., 2016).
567 Our results also highlighted the significant abundance of heterotrophic bacteria in the
568 early-stage samples, especially in Bo1, enriched in Actinobacteria and Bacteroidetes
569 (**Fig. 6-7**). MCI A revealed that Cytophagia (Bacteroidetes) and Nitriliruptoria
570 (Actinobacteria) correlated negatively with TOC (**Fig. 10**) and members of these
571 classes were identified as keystone species in Bo1 (**Fig. 8**), confirming their relevance
572 in oligotrophic habitats (Foreman et al., 2007; Gonzalez-Pimentel et al., 2018).
573 Heterotrophic communities may represent the earliest stage of microbial assembly,
574 degrading allochthonous organic compounds and supplying essential nutrients (TOC,
575 N, P) for the subsequent development of autotrophic species (Hodkinson et al., 2002).
576 Our results also indicated the presence of chemoautotrophic Proteobacteria typically
577 involved in the primary production of alkaline environments (**Fig. 7**), including iron-
578 oxidizing bacteria within Alphaproteobacteria (Rhodobacteraceae) and
579 Betaproteobacteria (Comamonadaceae) (Jamieson et al., 2018; Kumaraswamy et al.,
580 2006; Straub et al., 1996) and sulfur-oxidizing Chromatiales (Gammaproteobacteria)
581 (Yuan et al., 2021). In tune with previous studies (Harantová et al., 2017; Schmidt et
582 al., 2008), the abundance of non-symbiotic N-fixing bacteria, represented by members
583 of Clostridia, Opitutae, Comamonadaceae (Betaproteobacteria), Rhodospirillales
584 (Alphaproteobacteria) and Bacilli (Ciccazzo et al., 2016; Hingole and Pathak, 2013;
585 Navarro-Noya et al., 2016), increased significantly after the first years of primary
586 succession (**Fig. 5-6**). These bacteria could contribute to the slight increase in TN
587 observed after the first years of bauxite residue weathering (Wu et al., 2020),
588 undetectable in this study as it is expected to be less than 100 mg/kg. Nitrifying
589 bacteria (Nitrosomonas and Nitrospiraceae) were also more prevalent in Bo2,
590 confirming that, although the amount of ammonia derived from bacterial fixation is low
591 at early stages of primary succession, it must be sufficient to fuel nitrification (Ollivier
592 et al., 2011; Zeng et al., 2016). Moreover, our results suggest that the increase in N-
593 cycle bacteria could be due to a significant decrease in EC and Cr in Bo2 (**Fig. 2**, **Fig.**
594 **10**, Table S2), as N-cycle enzymes are particularly sensitive to high salinity (Claros et
595 al., 2010; Herbst, 1998) and heavy metal contamination (Kim et al., 2016; Oliveira and
596 Pampulha, 2006).

597 **4.3 Primary succession reaches an equilibrium in BRDA over few decades**

598 The last stage of primary succession arises after long periods of time when the habitat
599 becomes less harsh and environmental selection loses its strength in the assembly of
600 communities, reaching an equilibrium (Ferrenberg et al., 2013). This maturity is
601 expected to require more time to be developed in extreme geochemical systems as
602 BRDAs. (Santini et al., 2015a). In our study, Pr90 and Pr100 samples, which have
603 been naturally weathered for at least 90 years, did not show significant dissimilarities
604 in either physicochemical parameters (**Fig. 3a**, Tables S1 and S2) or bacterial and
605 fungal community composition (**Fig. 4b**; Fig. S5), suggesting that primary succession
606 had reached the equilibrium in these sites. In Pr90 and Pr100, pH (8.8 ± 0.3) and
607 salinity ($EC = 0.29 \pm 0.11 \text{ mS cm}^{-1}$) were significantly lower than in early-stage
608 samples (**Fig. 2**, Table S1), revealing a significant dependence with the age of the
609 samples. During bauxite residue natural weathering, salinity is generally assumed to
610 decrease due to the dissolution of alkaline minerals by rainfall (Cusack et al., 2019;
611 Zhu et al., 2016), while the decrease in pH is normally attributed to atmospheric and
612 microbial carbonation (Cusack et al., 2019; Khaitan et al., 2010; Schmalenberger,
613 2013; X. Kong et al., 2017). Moreover, Pr90 and Pr100 showed a significantly higher
614 concentration of macronutrients such as TOC, TN and P, which are accumulated
615 during the bauxite residue natural aging due to the action of the different
616 microorganisms that colonize the BRDAs (Santini et al., 2015a). This negative
617 correlation between pH-salinity and age-nutrients is characteristic of restored and
618 amended BRDAs (Courtney et al., 2014; Wu et al., 2021). Finally, our results also
619 indicated a significant and steady decrease in Cr and V associated with the natural
620 aging of the samples (**Fig. 2**, Table S2). However, the concentration of these metals
621 also seems to be determined by the bauxite type and, therefore, the effect of the ore
622 origin should not be neglected (Gentzmann et al., 2021).

623 Both alpha and beta bacterial diversity increased significantly in the equilibrium-stage
624 samples (**Fig. 4**), in line with the results observed in restored BRDAs (Krishna et al.,
625 2014; Wu et al., 2021, 2020). Beta-diversity analysis confirmed that “age” was the
626 main factor explaining the dissimilarities of microbial community between Bo1/Bo2 and
627 Pr90/Pr100 and not the ore origin. In addition, the changes in beta-diversity could be
628 explained by pH, EC and some metal concentration gradients, all of them strongly anti-
629 correlated with the age of the samples (**Fig. 3**). Circumneutral pH, lower salinity and

630 metal concentration, and increase in macronutrients support the survival of a more
631 diverse microbial community dominated by bacteria commonly found in ordinary soils
632 and freshwater like Alphaproteobacteria, Betaproteobacteria, Planctomycetes,
633 Gemmatimonadetes, and Acidobacteria. Acidobacteria are one of the most abundant
634 terrestrial bacterial taxa, reaching 52% from the total bacterial community in certain
635 soils (Kielak et al., 2016). Their abundance is positively correlated with low pH (Jones
636 et al., 2009), and their accumulation over time in BRDAs is commonly associated with
637 the restoration of chemical conditions in bauxite residue (Santini et al., 2015b; Wu et
638 al., 2021). The equilibrium-stage bauxite residue also harbors a more complex
639 microbial network, accounting for more highly connected taxa (nodes) than early-stage
640 residues. Nutrient availability and higher microbial richness have been shown to favor
641 microbial network complexity and the emergence of new ecological functions, making
642 the ecosystem more stable (Qiu et al., 2021; Wagg et al., 2019).

643 In contrast, fungal biodiversity and community structure did not appear to be
644 significantly affected by chemical gradients (Table S4, Fig. S5). The fungal community
645 structure in early-stage bauxite residues was dominated by Ascomycota (80.1 ± 9.8
646 % of fungal communities), and Basidiomycota (> 14 %). This distribution seems to be
647 the usual in natural and engineered haloalkaline environments, although the analysis
648 of fungal communities in these habitats is limited to few studies (Grum-Grzhimaylo et
649 al., 2016; Salano et al., 2017; Santini et al., 2015b). In the equilibrium-stage samples,
650 the relative abundance of Ascomycota decreased (52.1 ± 24.2 %) while
651 Basidiomycota (mostly Agaricomycetes) increased (35.0 ± 23.3 %). This pattern has
652 been observed in restored BRDAs and was attributed to a decrease in total alkalinity
653 (Santini et al., 2015b). In this study however, MCIA revealed a poor correlation
654 between fungal communities and chemical parameters (RV score = 0.36). It is known
655 that the fungal community assembly is more influenced by stochastic processes in
656 primary successional environments (Schmidt et al., 2014). Unlike bacteria, climate is
657 often considered the main environmental factor affecting the fungal community
658 composition, rather than soil chemical properties (Egidi et al., 2019). Interestingly, all
659 fungal keystones identified in Bo1 (**Fig. 8**) belonged to the class Dothideomycetes
660 (Pleosporaceae, Macrophoma, Cladosporium), within the Ascomycota.
661 Dothideomycetes constitute almost the entire fungal community in microbial mats from
662 soda lakes and are also found in microbial mats from hypersaline and iron-rich habitats

663 (Gerea et al., 2012; Maza-Márquez et al., 2021; Salano et al., 2017). In microbial mats,
664 fungi play a crucial role in nutrient recycling by decomposing complex carbohydrates
665 into simpler compounds that fuel chemoheterotrophic species (Carreira et al., 2020).

666 **4.4 Implications for bioremediation and metal recovery**

667 To date, only few studies have focused their efforts on an integrated chemical,
668 physical, and biological approach for the characterization of BRDA in the optic of
669 bioremediation or critical metal bio extraction strategies.

670 Our results highlight the key role of pH and EC for natural microbial restoration and
671 therefore should be considered primary targets in bioremediation. If natural weathering
672 appears to be effective in reducing both pH and EC, it is also a slow process that could
673 be artificially accelerated. In the last few years, microbially-driven pH neutralization
674 using organic acids and CO₂ produced by fermentation of added carbon sources has
675 gained attention as a promising technique for in-situ bioremediation of BRDAs (Santini
676 et al., 2021; Wu et al., 2019). Santini et al. (2021, 2016) proved that the efficiency of
677 this process could be improved by decreasing the initial pH and salinity and increasing
678 the biodiversity of the microbial inoculum. However, they also identified the shortage
679 of nutrients (N and P) in the medium term as a limiting factor for the development of
680 this methodology. Our results suggest that lowering EC increased biodiversity and
681 facilitated the development of bacteria related to the N cycle. Hence, a first treatment
682 focused on reducing salinity (and pH if possible) before bacteria-mediated pH
683 neutralization could yield considerably better results. Few studies have investigated
684 the effect of lowering artificially pH and EC in the optic of bioremediation of BRDAs
685 (Courtney et al., 2014; Jones et al., 2011; Wong and Ho, 1994). As a recent example,
686 the recent study by Fourrier et al. (2021) has reported the positive effect of gypsum
687 addition and repeated washing on pH neutralization and EC decrease. In addition,
688 fungal contribution to early-stage microbial assembly seems to be underrated and their
689 role in bauxite residue remediation needs more research. Regarding microbially-
690 driven strategies for metal bio-extraction, our study is particularly informative for
691 selecting appropriate bauxite residue and bacteria for the selective extraction of critical
692 metals. In agreement with literature (Rivera et al., 2019; Vind et al., 2018), our results
693 indicated that the concentration of REEs (Ln, Sc and Y) in bauxite residue is primarily
694 determined by the ore origin, with karst bauxite (Pr90 and Pr100) being more enriched
695 in REEs than laterite bauxite (Fresh, Bo1 and Bo2). In addition, Y and Ln were strongly

696 correlated as expected regarding the analogy between Y and heavy REEs while Sc
697 behaves differently to Y and Ln and is likely incorporated in iron oxyhydroxide such
698 as goethite (Levard et al., 2018). The conventional methods to recover these metals
699 are not environmentally acceptable due to their excessive energy consumption and
700 the production of hazardous residues, that require further treatment and high operating
701 costs (Baniasadi et al., 2019). Metal bio-extraction is an interesting alternative to
702 traditional metallurgy that has still been poorly investigated for critical metals. Among
703 the ideas that could benefit from our study, alkaline-active exoenzymes produced by
704 bacteria found in Bo1 and Bo2 offer interesting possibilities for the valorization of
705 bauxite residue. Despite their ability to lower their intracellular pH and salinity,
706 haloalkaliphilic species must still be able to secrete stable and operational enzymes
707 at elevated pH and salinity (Mamo, 2020), making BRDAs potential sources of these
708 enzymes. Some of the haloalkaliphilic bacteria known to synthesize alkaline-active
709 enzymes, namely Gammaproteobacteria, Actinobacteria and Firmicutes (Kalwasińska
710 et al., 2018; Litchfield, 2011; Maharaja et al., 2018; Shameer, 2016; Shivilata and
711 Tulasi, 2015), are also capable to modify the chemical phase and mobility of various
712 critical metals (Chidambaram et al., 2010; Presentato et al., 2020; Wee et al., 2014),
713 including REEs. For example, Maleke et al. (2019) reported the direct reduction of
714 Eu^{3+} to Eu^{2+} by *Clostridium* sp. (Firmicutes). In line with previous studies (Santini et
715 al., 2015b), the majority of Gammaproteobacteria corresponded to uncharacterized
716 lineages, highlighting the potential of this group as a phylogenetic hot spot for novel
717 haloalkaliphilic taxa.

718

719 **5. CONCLUSIONS**

720 In summary, this paper presents an integrated physicochemical and biological
721 approach to explore the composition and dynamics of bacterial and fungal
722 communities in bauxite residues deposited at different times and produced from
723 different bauxite ores. Our results revealed that both deposit age and ore origin affect
724 the geochemistry of bauxite residue, although unequally. Salinity, pH, TOC, TN, and
725 P values seem to depend predominantly on the natural aging of bauxite residue, while
726 the content of REE is mainly influenced by the origin of the bauxite ore. Our results
727 highlight the behavior of bauxite residue deposits as primary successional
728 environments and bring new insights into the early stages of microbial community

729 assembly in these sites. The pioneer microbial community was dominated by
730 haloalkaliphilic microorganisms, strongly influenced by chemical gradients.
731 Autotrophic and heterotrophic microbial species contribute to the supply of nutrients
732 necessary for the development of other species through C fixation and degradation of
733 allochthonous organic compounds, respectively. After the first years of natural
734 restoration, nitrogen-fixing bacteria increase their presence and contribute to enhance
735 N bioavailability. Microbial richness, diversity and network complexity increased
736 significantly with age of deposition, until primary succession equilibrium was reached
737 decades later, characterized by a microbial community composition similar to that of
738 typical soils and freshwater. Our results suggested that salinity, pH, nutrients, and toxic
739 metals (mainly Cr and V) were the main factors explaining this change in microbial
740 communities. These results confirm the key role of pH and salinity in the establishment
741 of early microbial communities and highlight them as main targets for bioremediation.
742 Moreover, our co-occurrence network data suggest an important role of fungal
743 communities in structuring the early-stage microbial community during primary
744 succession, potentially by recycling complex organic matter. We also identified
745 bacteria with potential metal extraction abilities, such as secretion of alkaline-active
746 enzymes that could modify the chemical phase of the metals present.

747

748 **Author contributions:** Conceptualization C.L., W.A., M.A.; Data curation L.A.M-P.;
749 Formal analysis L.A.M-P.; Funding acquisition C.L., W.A., M.A.; Investigation L.A.M-
750 P.; Methodology L.A.M-P., B.A., D.B., M.B., L.P.; Project administration C.L., W.A.,
751 M.A.; Supervision C.L., W.A., M.A.; Validation C.L., W.A., M.A., L.A.M-P., B.A., D.B.
752 M.B., L.P.; Writing - original draft L.A.M-P.; Writing - review & editing L.A.M-P., C.L.,
753 W.A., M.A. All authors have read and agreed to the published version of the
754 manuscript.

755

756 **Funding:** This project has received funding from the European Union's Horizon 2020
757 research and innovation program under the Marie Skłodowska-Curie grant agreement
758 No713750. Additionally, it has been performed with the financial supports of the
759 Regional Council of Provence-Alpes-Côte d'Azur and A*MIDEX (n° ANR11-IDEX-

760 0001-02), funded by the Investissements d'Avenir project funded by the French
761 Government and managed by the French National Research Agency (ANR).

762

763 **Acknowledgments:** This work is also a contribution to the OSU-Institut Pythéas. The
764 authors acknowledge the CNRS for the funding of the IRP iNOVE. The authors thank
765 the company ALTEO for providing access to the sampling sites.

766

767 **Conflicts of Interest:** The authors declare no conflict of interest.

768

769 6. REFERENCES

- 770 Anderson, M.J., Crist, T.O., Chase, J.M., Vellend, M., Inouye, B.D., Freestone, A.L.,
771 Sanders, N.J., Cornell, H.V., Comita, L.S., Davies, K.F., Harrison, S.P., Kraft, N.J.B.,
772 Stegen, J.C., Swenson, N.G., 2011. Navigating the multiple meanings of β diversity: a
773 roadmap for the practicing ecologist. *Ecol. Lett.* 14, 19–28.
774 <https://doi.org/10.1111/j.1461-0248.2010.01552.x>
- 775 Baniasadi, M., Vakilchap, F., Bahaloo-Horeh, N., Mousavi, S.M., Farnaud, S., 2019.
776 Advances in bioleaching as a sustainable method for metal recovery from e-waste: A
777 review. *J. Ind. Eng. Chem.* 76, 75–90. <https://doi.org/10.1016/j.jiec.2019.03.047>
- 778 Banning, N.C., Phillips, I.R., Jones, D.L., Murphy, D.V., 2011. Development of Microbial
779 Diversity and Functional Potential in Bauxite Residue Sand under Rehabilitation.
780 *Restor. Ecol.* 19, 78–87. <https://doi.org/10.1111/j.1526-100X.2009.00637.x>
- 781 Benjamini, Y., Hochberg, Y., 1995. Controlling the False Discovery Rate: A Practical and
782 Powerful Approach to Multiple Testing. *J. R. Stat. Soc. Ser. B Methodol.* 57, 289–
783 300. <https://doi.org/10.1111/j.2517-6161.1995.tb02031.x>
- 784 Berry, D., Widder, S., 2014. Deciphering microbial interactions and detecting keystone
785 species with co-occurrence networks. *Front. Microbiol.* 5, 219.
786 <https://doi.org/10.3389/fmicb.2014.00219>
- 787 Beule, L., Karlovsky, P., 2020. Improved normalization of species count data in ecology by
788 scaling with ranked subsampling (SRS): application to microbial communities. *PeerJ*
789 8, e9593. <https://doi.org/10.7717/peerj.9593>
- 790 Bolyen, E., Rideout, J.R., Dillon, M.R., Bokulich, N.A., Abnet, C.C., Al-Ghalith, G.A.,
791 Alexander, H., Alm, E.J., Arumugam, M., Asnicar, F., Bai, Y., Bisanz, J.E., Bittinger,
792 K., Brejnrod, A., Brislawn, C.J., Brown, C.T., Callahan, B.J., Caraballo-Rodríguez,
793 A.M., Chase, J., Cope, E.K., Da Silva, R., Diener, C., Dorrestein, P.C., Douglas,
794 G.M., Durall, D.M., Duvallet, C., Edwardson, C.F., Ernst, M., Estaki, M., Fouquier,
795 J., Gauglitz, J.M., Gibbons, S.M., Gibson, D.L., Gonzalez, A., Gorlick, K., Guo, J.,
796 Hillmann, B., Holmes, S., Holste, H., Huttenhower, C., Huttley, G.A., Janssen, S.,
797 Jarmusch, A.K., Jiang, L., Kaehler, B.D., Kang, K.B., Keefe, C.R., Keim, P., Kelley,
798 S.T., Knights, D., Koester, I., Kosciulek, T., Kreps, J., Langille, M.G.I., Lee, J., Ley,
799 R., Liu, Y.-X., Loftfield, E., Lozupone, C., Maher, M., Marotz, C., Martin, B.D.,
800 McDonald, D., McIver, L.J., Melnik, A.V., Metcalf, J.L., Morgan, S.C., Morton, J.T.,
801 Naimey, A.T., Navas-Molina, J.A., Nothias, L.F., Orchanian, S.B., Pearson, T.,

802 Peoples, S.L., Petras, D., Preuss, M.L., Priesse, E., Rasmussen, L.B., Rivers, A.,
803 Robeson, M.S., Rosenthal, P., Segata, N., Shaffer, M., Shiffer, A., Sinha, R., Song,
804 S.J., Spear, J.R., Swafford, A.D., Thompson, L.R., Torres, P.J., Trinh, P., Tripathi, A.,
805 Turnbaugh, P.J., Ul-Hasan, S., van der Hoof, J.J.J., Vargas, F., Vázquez-Baeza, Y.,
806 Vogtmann, E., von Hippel, M., Walters, W., Wan, Y., Wang, M., Warren, J., Weber,
807 K.C., Williamson, C.H.D., Willis, A.D., Xu, Z.Z., Zaneveld, J.R., Zhang, Y., Zhu, Q.,
808 Knight, R., Caporaso, J.G., 2019. Reproducible, interactive, scalable and extensible
809 microbiome data science using QIIME 2. *Nat. Biotechnol.* 37, 852–857.
810 <https://doi.org/10.1038/s41587-019-0209-9>

811 Bondici, V.F., Lawrence, J.R., Khan, N.H., Hill, J.E., Yergeau, E., Wolfaardt, G.M., Warner,
812 J., Korber, D.R., 2013. Microbial communities in low permeability, high pH uranium
813 mine tailings: characterization and potential effects. *J. Appl. Microbiol.* 114, 1671–
814 1686. <https://doi.org/10.1111/jam.12180>

815 Bouchoucha, M., Chekri, R., Leufroy, A., Jitaru, P., Millour, S., Marchond, N., Chafey, C.,
816 Testu, C., Zinck, J., Cresson, P., Mirallès, F., Mahe, A., Arnich, N., Sanaa, M.,
817 Bemrah, N., Guérin, T., 2019. Trace element contamination in fish impacted by
818 bauxite red mud disposal in the Cassidaigne canyon (NW French Mediterranean). *Sci.*
819 *Total Environ.* 690, 16–26. <https://doi.org/10.1016/j.scitotenv.2019.06.474>

820 Bray, A.W., Stewart, D.I., Courtney, R., Rout, S.P., Humphreys, P.N., Mayes, W.M., Burke,
821 I.T., 2018. Sustained Bauxite Residue Rehabilitation with Gypsum and Organic
822 Matter 16 years after Initial Treatment. *Environ. Sci. Technol.* 52, 152–161.
823 <https://doi.org/10.1021/acs.est.7b03568>

824 Callahan, B.J., McMurdie, P.J., Holmes, S.P., 2017. Exact sequence variants should replace
825 operational taxonomic units in marker-gene data analysis. *ISME J.* 11, 2639–2643.
826 <https://doi.org/10.1038/ismej.2017.119>

827 Callahan, B.J., McMurdie, P.J., Rosen, M.J., Han, A.W., Johnson, A.J.A., Holmes, S.P.,
828 2016. DADA2: High resolution sample inference from Illumina amplicon data. *Nat.*
829 *Methods* 13, 581–583. <https://doi.org/10.1038/nmeth.3869>

830 Caporaso, J.G., Lauber, C.L., Walters, W.A., Berg-Lyons, D., Lozupone, C.A., Turnbaugh,
831 P.J., Fierer, N., Knight, R., 2011. Global patterns of 16S rRNA diversity at a depth of
832 millions of sequences per sample. *Proc. Natl. Acad. Sci.* 108, 4516–4522.
833 <https://doi.org/10.1073/pnas.1000080107>

834 Carreira, C., Lønborg, C., Kühl, M., Lillebø, A., Sandaa, R.-A., Villanueva, L., Cruz, S.,
835 2020. Fungi and viruses as important players in microbial mats. *FEMS Microbiol.*
836 *Ecol.* 96. <https://doi.org/10.1093/femsec/fiaa187>

837 Chakraborty, J., Rajput, V., Sapkale, V., Kamble, S., Dharne, M., 2021. Spatio-temporal
838 resolution of taxonomic and functional microbiome of Lonar soda lake of India
839 reveals metabolic potential for bioremediation. *Chemosphere* 264, 128574.
840 <https://doi.org/10.1016/j.chemosphere.2020.128574>

841 Chen, D.-D., Tian, Y., Jiao, J.-Y., Zhang, X.-T., Zhang, Y.-G., Dong, Z.-Y., Xiong, M.-J.,
842 Xiao, M., Shu, W.-S., Li, W.-J., 2020. Comparative genomics analysis of
843 Nitriliruptoria reveals the genomic differences and salt adaptation strategies.
844 *Extremophiles* 24, 249–264. <https://doi.org/10.1007/s00792-019-01150-3>

845 Chen, L., Liu, J., Hu, W., Gao, J., Yang, J., 2021. Vanadium in soil-plant system: Source,
846 fate, toxicity, and bioremediation. *J. Hazard. Mater.* 405, 124200.
847 <https://doi.org/10.1016/j.jhazmat.2020.124200>

848 Chidambaram, D., Hennebel, T., Taghavi, S., Mast, J., Boon, N., Verstraete, W., van der
849 Lelie, D., Fitts, J.P., 2010. Concomitant microbial generation of palladium
850 nanoparticles and hydrogen to immobilize chromate. *Environ. Sci. Technol.* 44,
851 7635–7640. <https://doi.org/10.1021/es101559r>

852 Ciccazzo, S., Esposito, A., Borruso, L., Brusetti, L., 2016. Microbial communities and
853 primary succession in high altitude mountain environments. *Ann. Microbiol.* 66, 43–
854 60. <https://doi.org/10.1007/s13213-015-1130-1>

855 Claros, J., Jiménez, E., Borrás, L., Aguado, D., Seco, A., Ferrer, J., Serralta, J., 2010. Short-
856 term effect of ammonia concentration and salinity on activity of ammonia oxidizing
857 bacteria. *Water Sci. Technol.* 61, 3008–3016. <https://doi.org/10.2166/wst.2010.217>

858 Copernicus Climate Change Service Climate Data Store (CDS), 2021. Copernicus Climate
859 Change Service (C3S) (2017): ERA5: Fifth generation of ECMWF atmospheric
860 reanalyses of the global climate. ECMWF.

861 Courtney, R., Harris, J., Pawlett, M., 2014. Microbial Community Composition in a
862 Rehabilitated Bauxite Residue Disposal Area: A Case Study for Improving Microbial
863 Community Composition. *Restor. Ecol.* 22. <https://doi.org/10.1111/rec.12143>

864 Csardi, G., Nepusz, T., 2006. The igraph software package for complex network research,
865 *InterJournal*.

866 Cusack, P.B., Courtney, R., Healy, M.G., O’ Donoghue, L.M.T., Ujaczki, É., 2019. An
867 evaluation of the general composition and critical raw material content of bauxite
868 residue in a storage area over a twelve-year period. *J. Clean. Prod.* 208, 393–401.
869 <https://doi.org/10.1016/j.jclepro.2018.10.083>

870 Dentoni, V., Grosso, B., Pinna, F., 2021. Experimental Evaluation of PM Emission from Red
871 Mud Basins Exposed to Wind Erosion. *Minerals* 11, 405.
872 <https://doi.org/10.3390/min11040405>

873 Dev, S., Sachan, A., Dehghani, F., Ghosh, T., Briggs, B.R., Aggarwal, S., 2020. Mechanisms
874 of biological recovery of rare-earth elements from industrial and electronic wastes: A
875 review. *Chem. Eng. J.* 397, 124596. <https://doi.org/10.1016/j.cej.2020.124596>

876 Di Carlo, E., Boullemant, A., Courtney, R., 2020. Ecotoxicological risk assessment of
877 revegetated bauxite residue: Implications for future rehabilitation programmes. *Sci.*
878 *Total Environ.* 698, 134344. <https://doi.org/10.1016/j.scitotenv.2019.134344>

879 Di Carlo, E., Chen, C.R., Haynes, R.J., Phillips, I.R., Courtney, R., 2019. Soil quality and
880 vegetation performance indicators for sustainable rehabilitation of bauxite residue
881 disposal areas: a review. *Soil Res.* 57, 419. <https://doi.org/10.1071/SR18348>

882 Doebelin, N., Kleeberg, R., 2015. Profex: a graphical user interface for the Rietveld
883 refinement program BGMN. *J. Appl. Crystallogr.* 48, 1573–1580.
884 <https://doi.org/10.1107/S1600576715014685>

885 Dominguez-Benetton, X., Varia, J.C., Pozo, G., Modin, O., Ter Heijne, A., Fransaer, J.,
886 Rabaey, K., 2018. Metal recovery by microbial electro-metallurgy. *Prog. Mater. Sci.*
887 94, 435–461. <https://doi.org/10.1016/j.pmatsci.2018.01.007>

888 Economou-Eliopoulos, M., Frei, R., Megremi, I., 2016. Potential leaching of Cr(VI) from
889 laterite mines and residues of metallurgical products (red mud and slag): An
890 integrated approach. *J. Geochem. Explor.* 162, 40–49.
891 <https://doi.org/10.1016/j.gexplo.2015.12.007>

892 Egidi, E., Delgado-Baquerizo, M., Plett, J.M., Wang, J., Eldridge, D.J., Bardgett, R.D.,
893 Maestre, F.T., Singh, B.K., 2019. A few Ascomycota taxa dominate soil fungal
894 communities worldwide. *Nat. Commun.* 10, 2369. <https://doi.org/10.1038/s41467-019-10373-z>

896 European Commission, 2020. Critical Raw Materials for Strategic Technologies and Sectors
897 in the EU.

898 Evans, K., 2016. The History, Challenges, and New Developments in the Management and
899 Use of Bauxite Residue. *J. Sustain. Metall.* 2, 316–331.
900 <https://doi.org/10.1007/s40831-016-0060-x>

901 Ferrenberg, S., O'Neill, S.P., Knelman, J.E., Todd, B., Duggan, S., Bradley, D., Robinson,
902 T., Schmidt, S.K., Townsend, A.R., Williams, M.W., Cleveland, C.C., Melbourne,
903 B.A., Jiang, L., Nemergut, D.R., 2013. Changes in assembly processes in soil
904 bacterial communities following a wildfire disturbance. *ISME J.* 7, 1102–1111.
905 <https://doi.org/10.1038/ismej.2013.11>

906 Foreman, C.M., Sattler, B., Mikucki, J.A., Porazinska, D.L., Priscu, J.C., 2007. Metabolic
907 activity and diversity of cryoconites in the Taylor Valley, Antarctica. *J. Geophys. Res.*
908 *Biogeosciences* 112. <https://doi.org/10.1029/2006JG000358>

909 Fourier, C., Luglia, M., Hennebert, P., Foulon, J., Ambrosi, J.-P., Angeletti, B., Keller, C.,
910 Criquet, S., 2020. Effects of increasing concentrations of unamended and gypsum
911 modified bauxite residues on soil microbial community functions and structure – A
912 mesocosm study. *Ecotoxicol. Environ. Saf.* 201, 110847.
913 <https://doi.org/10.1016/j.ecoenv.2020.110847>

914 Fourier, C., Luglia, M., Keller, C., Hennebert, P., Foulon, J., Ambrosi, J.-P., Angeletti, B.,
915 Criquet, S., 2021. How Raw and Gypsum Modified Bauxite Residues Affect Seed
916 Germination, Enzyme Activities, and Root Development of *Sinapis alba*. *Water. Air.*
917 *Soil Pollut.* 232, 309. <https://doi.org/10.1007/s11270-021-05232-x>

918 Fox, J., Weisberg, S., 2019. *An R Companion to Applied Regression*, Third. ed. Sage,
919 Thousand Oaks, CA.

920 Gentzmann, M.C., Schraut, K., Vogel, C., Gäbler, H.-E., Huthwelker, T., Adam, C., 2021.
921 Investigation of scandium in bauxite residues of different origin. *Appl. Geochem.*
922 126, 104898. <https://doi.org/10.1016/j.apgeochem.2021.104898>

923 Gereá, A.L., Branscum, K.M., King, J.B., You, J., Powell, D.R., Miller, A.N., Spear, J.R.,
924 Cichewicz, R.H., 2012. Secondary metabolites produced by fungi derived from a
925 microbial mat encountered in an iron-rich natural spring. *Tetrahedron Lett.* 53, 4202–
926 4205. <https://doi.org/10.1016/j.tetlet.2012.05.156>

927 Ghosh, S., Bal, B., Das, A.P., 2018. Enhancing Manganese Recovery from Low-Grade Ores
928 by Using Mixed Culture of Indigenously Isolated Bacterial Strains. *Geomicrobiol. J.*
929 35, 242–246. <https://doi.org/10.1080/01490451.2017.1362080>

930 Gonzalez-Pimentel, J.L., Miller, A.Z., Jurado, V., Laiz, L., Pereira, M.F.C., Saiz-Jimenez, C.,
931 2018. Yellow coloured mats from lava tubes of La Palma (Canary Islands, Spain) are
932 dominated by metabolically active Actinobacteria. *Sci. Rep.* 8, 1944.
933 <https://doi.org/10.1038/s41598-018-20393-2>

934 Grum-Grzhimaylo, A.A., Georgieva, M.L., Bondarenko, S.A., Debets, A.J.M., Bilanenko,
935 E.N., 2016. On the diversity of fungi from soda soils. *Fungal Divers.* 76, 27–74.
936 <https://doi.org/10.1007/s13225-015-0320-2>

937 Gunde-Cimerman, N., Plemenitaš, A., Oren, A., 2018. Strategies of adaptation of
938 microorganisms of the three domains of life to high salt concentrations. *FEMS*
939 *Microbiol. Rev.* 42, 353–375. <https://doi.org/10.1093/femsre/fuy009>

940 Guo, Y., Fujimura, R., Sato, Y., Suda, W., Kim, S., Oshima, K., Hattori, M., Kamijo, T.,
941 Narisawa, K., Ohta, H., 2014. Characterization of Early Microbial Communities on
942 Volcanic Deposits along a Vegetation Gradient on the Island of Miyake, Japan.
943 *Microbes Environ.* 29, 38–49. <https://doi.org/10.1264/jsme2.ME13142>

944 Harantová, L., Mudrák, O., Kohout, P., Elhottová, D., Frouz, J., Baldrian, P., 2017.
945 Development of microbial community during primary succession in areas degraded
946 by mining activities. *Land Degrad. Dev.* 28, 2574–2584.
947 <https://doi.org/10.1002/ldr.2817>

948 Herbst, D.B., 1998. Potential salinity limitations on nitrogen fixation in sediments from
949 Mono Lake, California. *Int. J. Salt Lake Res.* 7, 261–274.
950 <https://doi.org/10.1007/BF02441878>

951 Hill, T.C.J., Walsh, K.A., Harris, J.A., Moffett, B.F., 2003. Using ecological diversity
952 measures with bacterial communities. *FEMS Microbiol. Ecol.* 43, 1–11.
953 <https://doi.org/10.1111/j.1574-6941.2003.tb01040.x>

954 Hingole, S., Pathak, A., 2013. Report on efficient salt stable *Azospirillum* a Lonar Soda Lake
955 isolate. *Sci. Res. Report.* 3, 200–203.

956 Hodgkinson, I.D., Webb, N.R., Coulson, S.J., 2002. Primary community assembly on land –
957 the missing stages: why are the heterotrophic organisms always there first? *J. Ecol.*
958 90, 569–577. <https://doi.org/10.1046/j.1365-2745.2002.00696.x>

959 Ihrmark, K., Bödeker, I.T.M., Cruz-Martinez, K., Friberg, H., Kubartova, A., Schenck, J.,
960 Strid, Y., Stenlid, J., Brandström-Durling, M., Clemmensen, K.E., Lindahl, B.D.,
961 2012. New primers to amplify the fungal ITS2 region – evaluation by 454-sequencing
962 of artificial and natural communities. *FEMS Microbiol. Ecol.* 82, 666–677.
963 <https://doi.org/10.1111/j.1574-6941.2012.01437.x>

964 International Aluminium Institute (IAI), European Aluminium (EA), 2015. Bauxite Residue
965 Management: Best Practice [WWW Document]. URL www.world-aluminium.org

966 Jamieson, J., Prommer, H., Kaksonen, A.H., Sun, J., Siade, A.J., Yusov, A., Bostick, B.,
967 2018. Identifying and Quantifying the Intermediate Processes during Nitrate-
968 Dependent Iron(II) Oxidation. *Environ. Sci. Technol.* 52, 5771–5781.
969 <https://doi.org/10.1021/acs.est.8b01122>

970 Ji, M., Kong, W., Yue, L., Wang, J., Deng, Y., Zhu, L., 2019. Salinity reduces bacterial
971 diversity, but increases network complexity in Tibetan Plateau lakes. *FEMS*
972 *Microbiol. Ecol.* 95. <https://doi.org/10.1093/femsec/fiz190>

973 Jones, B.E.H., Haynes, R.J., Phillips, I.R., 2011. Influence of organic waste and residue mud
974 additions on chemical, physical and microbial properties of bauxite residue sand.
975 *Environ. Sci. Pollut. Res.* 18, 199–211. <https://doi.org/10.1007/s11356-010-0364-5>

976 Jones, R.T., Robeson, M.S., Lauber, C.L., Hamady, M., Knight, R., Fierer, N., 2009. A
977 comprehensive survey of soil acidobacterial diversity using pyrosequencing and clone
978 library analyses. *ISME J.* 3, 442–453. <https://doi.org/10.1038/ismej.2008.127>

979 Kalwasińska, A., Felföldi, T., Szabó, A., Deja-Sikora, E., Kosobucki, P., Walczak, M., 2017.
980 Microbial communities associated with the anthropogenic, highly alkaline
981 environment of a saline soda lime, Poland. *Antonie Van Leeuwenhoek* 110, 945–962.
982 <https://doi.org/10.1007/s10482-017-0866-y>

983 Kalwasińska, A., Jankiewicz, U., Felföldi, T., Burkowska-But, A., Brzezinska, M.S., 2018.
984 Alkaline and Halophilic Protease Production by *Bacillus luteus* H11 and Its Potential
985 Industrial Applications. *Food Technol. Biotechnol.* 56, 553–561.
986 <https://doi.org/10.17113/ftb.56.04.18.5553>

987 Kassambara, A., 2021. rstatix: Pipe-Friendly Framework for Basic Statistical Tests.

988 Kassambara, A., 2020. ggpubr: “ggplot2” Based Publication Ready Plots.

989 Kassambara, A., Mundt, F., 2020. factoextra: Extract and Visualize the Results of
990 Multivariate Data Analyses.

991 Ke, W., 2021. Appropriate human intervention stimulates the development of microbial
992 communities and soil formation at a long-term weathered bauxite residue disposal
993 area. *J. Hazard. Mater.* 10.

994 Khaitan, S., Dzombak, D.A., Swallow, P., Schmidt, K., Fu, J., Lowry, G.V., 2010. Field
995 Evaluation of Bauxite Residue Neutralization by Carbon Dioxide, Vegetation, and
996 Organic Amendments. *J. Environ. Eng.* 136, 1045–1053.
997 [https://doi.org/10.1061/\(ASCE\)EE.1943-7870.0000230](https://doi.org/10.1061/(ASCE)EE.1943-7870.0000230)

998 Kielak, A.M., Barreto, C.C., Kowalchuk, G.A., van Veen, J.A., Kuramae, E.E., 2016. The
999 Ecology of Acidobacteria: Moving beyond Genes and Genomes. *Front. Microbiol.* 7.
1000 <https://doi.org/10.3389/fmicb.2016.00744>

- 1001 Kim, Y.M., Park, H., Chandran, K., 2016. Nitrification inhibition by hexavalent chromium
1002 Cr(VI) – Microbial ecology, gene expression and off-gas emissions. *Water Res.* 92,
1003 254–261. <https://doi.org/10.1016/j.watres.2016.01.042>
- 1004 Kiskira, K., Lymperopoulou, T., Tsakanika, L.-A., Pavlopoulos, C., Papadopoulou, K.,
1005 Ochsenkühn, K.-M., Lyberatos, G., Ochsenkühn-Petropoulou, M., 2021. Study of
1006 Microbial Cultures for the Bioleaching of Scandium from Alumina Industry By-
1007 Products. *Metals* 11, 951. <https://doi.org/10.3390/met11060951>
- 1008 Krishna, P., Babu, A.G., Reddy, M.S., 2014. Bacterial diversity of extremely alkaline bauxite
1009 residue site of alumina industrial plant using culturable bacteria and residue 16S
1010 rRNA gene clones 12.
- 1011 Kumaraswamy, R., Sjollem, K., Kuenen, J.G., van Loosdrecht, M., Muyzer, G., 2006.
1012 *Paracoccus ferrooxidans*.
- 1013 Lê, S., Josse, J., Husson, F., 2008. FactoMineR: An R Package for Multivariate Analysis. *J.*
1014 *Stat. Softw.* 25. <https://doi.org/10.18637/jss.v025.i01>
- 1015 Levard, C., Borschneck, D., Grauby, O., Rose, J., Ambrosi, J.P., 2018. Goethite, a tailor-
1016 made host for the critical metal scandium: The $\text{Fe}_{1-x}\text{Sc}_x\text{OOH}$ solid solution.
1017 *Geochem. Perspect. Lett.* 9, 16–20. <https://doi.org/10.7185/geochemlet.1832>
- 1018 Li, K., Liu, R., Zhang, H., Yun, J., 2013. The Diversity and Abundance of Bacteria and
1019 Oxygenic Phototrophs in Saline Biological Desert Crusts in Xinjiang, Northwest
1020 China. *Microb. Ecol.* 66, 40–48. <https://doi.org/10.1007/s00248-012-0164-1>
- 1021 Liang, J., Huang, X., Yan, J., Li, Y., Zhao, Z., Liu, Y., Ye, J., Wei, Y., 2021. A review of the
1022 formation of Cr(VI) via Cr(III) oxidation in soils and groundwater. *Sci. Total*
1023 *Environ.* 774, 145762. <https://doi.org/10.1016/j.scitotenv.2021.145762>
- 1024 Litchfield, C.D., 2011. Potential for industrial products from the halophilic Archaea. *J. Ind.*
1025 *Microbiol. Biotechnol.* 38, 1635–1635. <https://doi.org/10.1007/s10295-011-1021-9>
- 1026 Lozupone, C.A., Hamady, M., Kelley, S.T., Knight, R., 2007. Quantitative and Qualitative β
1027 Diversity Measures Lead to Different Insights into Factors That Structure Microbial
1028 Communities. *Appl. Environ. Microbiol.* 73, 1576–1585.
1029 <https://doi.org/10.1128/AEM.01996-06>
- 1030 Lyu, F., Hu, Y., Wang, L., Sun, W., 2021. Dealkalization processes of bauxite residue: A
1031 comprehensive review. *J. Hazard. Mater.* 403, 123671.
1032 <https://doi.org/10.1016/j.jhazmat.2020.123671>
- 1033 Ma, L., Wang, H., Wu, J., Wang, Y., Zhang, D., Liu, X., 2019. Metatranscriptomics reveals
1034 microbial adaptation and resistance to extreme environment coupling with
1035 bioleaching performance. *Bioresour. Technol.* 280, 9–17.
1036 <https://doi.org/10.1016/j.biortech.2019.01.117>
- 1037 Ma, L., Wang, X., Feng, X., Liang, Y., Xiao, Y., Hao, X., Yin, H., Liu, H., Liu, X., 2017. Co-
1038 culture microorganisms with different initial proportions reveal the mechanism of
1039 chalcopyrite bioleaching coupling with microbial community succession. *Bioresour.*
1040 *Technol.* 223, 121–130. <https://doi.org/10.1016/j.biortech.2016.10.056>
- 1041 Maes, S., Claus, M., Verbeken, K., Wallaert, E., De Smet, R., Vanhaecke, F., Boon, N.,
1042 Hennebel, T., 2016. Platinum recovery from industrial process streams by halophilic
1043 bacteria: Influence of salt species and platinum speciation. *Water Res.* 105, 436–443.
1044 <https://doi.org/10.1016/j.watres.2016.09.023>
- 1045 Maharaja, P., Nanthini, E., Swarnalatha, S., Sekaran, G., 2018. Studies on the Production of
1046 Salt-Tolerant Alkaline Protease Isolated from *Proteus mirabilis* and Its Degradation of
1047 Hyper-Saline Soak Liquor, in: Singh, V.P., Yadav, S., Yadava, R.N. (Eds.),
1048 *Environmental Pollution, Water Science and Technology Library*. Springer,
1049 Singapore, pp. 439–457. https://doi.org/10.1007/978-981-10-5792-2_35

1050 Maleke, M., Valverde, A., Gomez-Arias, A., Cason, E.D., Vermeulen, J.-G., Coetsee-Hugo,
1051 L., Swart, H., van Heerden, E., Castillo, J., 2019. Anaerobic reduction of europium by
1052 a *Clostridium* strain as a strategy for rare earth biorecovery. *Sci. Rep.* 9, 14339.
1053 <https://doi.org/10.1038/s41598-019-50179-z>

1054 Mamo, G., 2020. Challenges and Adaptations of Life in Alkaline Habitats, in: Mamo, G.,
1055 Mattiasson, B. (Eds.), *Alkaliphiles in Biotechnology, Advances in Biochemical*
1056 *Engineering/Biotechnology*. Springer International Publishing, Cham, pp. 85–133.
1057 https://doi.org/10.1007/10_2019_97

1058 Markus Gräfe, Matthew Landers, Ryan Tappero, Peter Austin, Bee Gan, Alton Grabsch,
1059 Craig Klauber, 2011. Combined Application of QEM-SEM and Hard X-ray
1060 Microscopy to Determine Mineralogical Associations and Chemical Speciation of
1061 Trace Metals. *J. Environ. Qual.* <https://doi.org/10.2134/jeq2010.0214>

1062 Maza-Márquez, P., Lee, M.D., Bebout, B.M., 2021. The Abundance and Diversity of Fungi
1063 in a Hypersaline Microbial Mat from Guerrero Negro, Baja California, México. *J.*
1064 *Fungi Basel Switz.* 7, 210. <https://doi.org/10.3390/jof7030210>

1065 McDonald, D., Price, M.N., Goodrich, J., Nawrocki, E.P., DeSantis, T.Z., Probst, A.,
1066 Andersen, G.L., Knight, R., Hugenholtz, P., 2012. An improved Greengenes
1067 taxonomy with explicit ranks for ecological and evolutionary analyses of bacteria and
1068 archaea. *ISME J.* 6, 610–618. <https://doi.org/10.1038/ismej.2011.139>

1069 Meng, C., Kuster, B., Culhane, A.C., Gholami, A.M., 2014. A multivariate approach to the
1070 integration of multi-omics datasets. *BMC Bioinformatics* 15, 162.
1071 <https://doi.org/10.1186/1471-2105-15-162>

1072 Milačič, R., Zuliani, T., Ščančar, J., 2012. Environmental impact of toxic elements in red
1073 mud studied by fractionation and speciation procedures. *Sci. Total Environ.* 426, 359–
1074 365. <https://doi.org/10.1016/j.scitotenv.2012.03.080>

1075 Muller, L.A.H., Ballhausen, M.-B., Andrade-Linares, D.R., Pinek, L., Golubeva, P., Rillig,
1076 M.C., 2021. Fungus–bacterium associations are widespread in fungal cultures isolated
1077 from a semi-arid natural grassland in Germany. *FEMS Microbiol. Ecol.* 97, fiab059.
1078 <https://doi.org/10.1093/femsec/fiab059>

1079 Muyzer, G., de Waal, E.C., Uitterlinden, A.G., 1993. Profiling of complex microbial
1080 populations by denaturing gradient gel electrophoresis analysis of polymerase chain
1081 reaction-amplified genes coding for 16S rRNA. *Appl. Environ. Microbiol.* 59, 695–
1082 700.

1083 Navarro-Noya, Y.E., Luna-Guido, M., Dendooven, L., 2016. Cultivable Nitrogen Fixing
1084 Bacteria from Extremely Alkaline-Saline Soils. *Adv. Microbiol.* 6, 412–423.
1085 <https://doi.org/10.4236/aim.2016.66041>

1086 Ohlsson, J.I., Osvatic, J.T., Becraft, E.D., Swingle, W.D., 2019. Microbial Community in
1087 Hyperalkaline Steel Slag-Fill Emulates Serpentinizing Springs. *Diversity* 11, 103.
1088 <https://doi.org/10.3390/d11070103>

1089 Oksanen, J., Blanchet, F.G., Friendly, M., Kindt, R., Legendre, P., McGlenn, D., Minchin,
1090 P.R., O’Hara, R.B., Simpson, G.L., Solymos, P., Stevens, M.H.H., Szoecs, E.,
1091 Wagner, H., 2020. *vegan: Community Ecology Package*.

1092 Oliveira, A., Pampulha, M.E., 2006. Effects of long-term heavy metal contamination on soil
1093 microbial characteristics. *J. Biosci. Bioeng.* 102, 157–161.
1094 <https://doi.org/10.1263/jbb.102.157>

1095 Ollivier, J., Töwe, S., Bannert, A., Hai, B., Kastl, E.-M., Meyer, A., Su, M.X., Kleineidam,
1096 K., Schloter, M., 2011. Nitrogen turnover in soil and global change. *FEMS Microbiol.*
1097 *Ecol.* 78, 3–16. <https://doi.org/10.1111/j.1574-6941.2011.01165.x>

1098 Ortiz-Álvarez, R., Fierer, N., de los Ríos, A., Casamayor, E.O., Barberán, A., 2018.
1099 Consistent changes in the taxonomic structure and functional attributes of bacterial

1100 communities during primary succession. *ISME J.* 12, 1658–1667.
1101 <https://doi.org/10.1038/s41396-018-0076-2>

1102 Panda, S., Costa, R.B., Shah, S.S., Mishra, S., Bevilaqua, D., Akcil, A., 2021.
1103 Biotechnological trends and market impact on the recovery of rare earth elements
1104 from bauxite residue (red mud) – A review. *Resour. Conserv. Recycl.* 171, 105645.
1105 <https://doi.org/10.1016/j.resconrec.2021.105645>

1106 Perez-Garcia, O., Lear, G., Singhal, N., 2016. Metabolic Network Modeling of Microbial
1107 Interactions in Natural and Engineered Environmental Systems. *Front. Microbiol.* 7.
1108 <https://doi.org/10.3389/fmicb.2016.00673>

1109 Poudel, R., Jumpponen, A., Schlatter, D.C., Paulitz, T.C., Gardener, B.B.M., Kinkel, L.L.,
1110 Garrett, K.A., 2016. Microbiome Networks: A Systems Framework for Identifying
1111 Candidate Microbial Assemblages for Disease Management. *Phytopathology®* 106,
1112 1083–1096. <https://doi.org/10.1094/PHYTO-02-16-0058-FI>

1113 Presentato, A., Piacenza, E., Turner, R.J., Zannoni, D., Cappelletti, M., 2020. Processing of
1114 Metals and Metalloids by Actinobacteria: Cell Resistance Mechanisms and Synthesis
1115 of Metal(loid)-Based Nanostructures. *Microorganisms* 8, 2027.
1116 <https://doi.org/10.3390/microorganisms8122027>

1117 Qiu, L., Zhang, Q., Zhu, H., Reich, P.B., Banerjee, S., van der Heijden, M.G.A., Sadowsky,
1118 M.J., Ishii, S., Jia, X., Shao, M., Liu, B., Jiao, H., Li, H., Wei, X., 2021. Erosion
1119 reduces soil microbial diversity, network complexity and multifunctionality. *ISME J.*
1120 15, 2474–2489. <https://doi.org/10.1038/s41396-021-00913-1>

1121 Qu, Y., Li, H., Wang, X., Tian, W., Shi, B., Yao, M., Zhang, Y., 2019. Bioleaching of Major,
1122 Rare Earth, and Radioactive Elements from Red Mud by using Indigenous
1123 Chemoheterotrophic Bacterium *Acetobacter* sp. *Minerals* 9, 67.
1124 <https://doi.org/10.3390/min9020067>

1125 R Core Team, 2020. R: A language and environment for statistical computing. R Foundation
1126 for Statistical Computing, Vienna, Austria.

1127 Ren, J., Chen, J., Han, L., Wang, M., Yang, B., Du, P., Li, F., 2018. Spatial distribution of
1128 heavy metals, salinity and alkalinity in soils around bauxite residue disposal area. *Sci.*
1129 *Total Environ.* 628–629, 1200–1208. <https://doi.org/10.1016/j.scitotenv.2018.02.149>

1130 Rivera, R.M., Ounoughene, G., Malfliet, A., Vind, J., Panias, D., Vassiliadou, V.,
1131 Binnemans, K., Van Gerven, T., 2019. A Study of the Occurrence of Selected Rare-
1132 Earth Elements in Neutralized–Leached Bauxite Residue and Comparison with
1133 Untreated Bauxite Residue. *J. Sustain. Metall.* 5, 57–68.
1134 <https://doi.org/10.1007/s40831-018-0206-0>

1135 Roy, A., Dutta, A., Pal, S., Gupta, A., Sarkar, J., Chatterjee, A., Saha, A., Sarkar, P., Sar, P.,
1136 Kazy, S.K., 2018. Biostimulation and bioaugmentation of native microbial
1137 community accelerated bioremediation of oil refinery sludge. *Bioresour. Technol.*
1138 253, 22–32. <https://doi.org/10.1016/j.biortech.2018.01.004>

1139 Sajjad, W., Zheng, G., Ma, X., Xu, W., Ali, B., Rafiq, M., Zada, S., Irfan, M., Zeman, J.,
1140 2020. Dissolution of Cu and Zn-bearing ore by indigenous iron-oxidizing bacterial
1141 consortia supplemented with dried bamboo sawdust and variations in bacterial
1142 structural dynamics: A new concept in bioleaching. *Sci. Total Environ.* 709, 136136.
1143 <https://doi.org/10.1016/j.scitotenv.2019.136136>

1144 Salano, O.A., Makonde, H.M., Kasili, R.W., Wangai, L.N., Nawiri, M.P., Boga, H.I., 2017.
1145 Diversity and distribution of fungal communities within the hot springs of soda lakes
1146 in the Kenyan rift valley. *Afr. J. Microbiol. Res.* 11, 764–775.
1147 <https://doi.org/10.5897/AJMR2017.8510>

- 1148 Santini, T., Kerr, J.L., Warren, L.A., 2015a. Microbially-driven strategies for bioremediation
1149 of bauxite residue. *J. Hazard. Mater.* 293, 131–157.
1150 <https://doi.org/10.1016/j.jhazmat.2015.03.024>
- 1151 Santini, T., Warren, L.A., Kendra, K.E., 2015b. Microbial Diversity in Engineered
1152 Haloalkaline Environments Shaped by Shared Geochemical Drivers Observed in
1153 Natural Analogues. *Appl. Environ. Microbiol.* 81, 5026–5036.
1154 <https://doi.org/10.1128/AEM.01238-15>
- 1155 Santini, T.C., Malcolm, L.I., Tyson, G.W., Warren, L.A., 2016. pH and Organic Carbon Dose
1156 Rates Control Microbially Driven Bioremediation Efficacy in Alkaline Bauxite
1157 Residue. *Environ. Sci. Technol.* 50, 11164–11173.
1158 <https://doi.org/10.1021/acs.est.6b01973>
- 1159 Santini, T.C., Wang, J.C., Warren, K.L., Pickering, G., Raudsepp, M.J., 2021. Simple
1160 Organic Carbon Sources and High Diversity Inocula Enhance Microbial
1161 Bionutralization of Alkaline Bauxite Residues. *Environ. Sci. Technol.* 55, 3929–
1162 3939. <https://doi.org/10.1021/acs.est.0c02534>
- 1163 Santini, T.C., Warren, K., Raudsepp, M., Carter, N., Hamley, D., McCosker, C.,
1164 Couperthwaite, S., Southam, G., Tyson, G.W., Warren, L.A., 2019. Accelerating
1165 Bauxite Residue Remediation with Microbial Biotechnology, in: Chesonis, C. (Ed.),
1166 Light Metals 2019, The Minerals, Metals & Materials Series. Springer International
1167 Publishing, Cham, pp. 69–77. https://doi.org/10.1007/978-3-030-05864-7_10
- 1168 Schmalenberger, A., 2013. Bacterial Communities Established in Bauxite Residues with
1169 Different Restoration Histories. *Env. Sci Technol* 10.
- 1170 Schmidt, S.K., Nemergut, D.R., Darcy, J.L., Lynch, R., 2014. Do bacterial and fungal
1171 communities assemble differently during primary succession? *Mol. Ecol.* 23, 254–
1172 258. <https://doi.org/10.1111/mec.12589>
- 1173 Schmidt, S.K., Reed, S.C., Nemergut, D.R., Stuart Grandy, A., Cleveland, C.C., Weintraub,
1174 M.N., Hill, A.W., Costello, E.K., Meyer, A.F., Neff, J.C., Martin, A.M., 2008. The
1175 earliest stages of ecosystem succession in high-elevation (5000 metres above sea
1176 level), recently deglaciated soils. *Proc. R. Soc. B Biol. Sci.* 275, 2793–2802.
1177 <https://doi.org/10.1098/rspb.2008.0808>
- 1178 Segata, N., Izard, J., Waldron, L., Gevers, D., Miropolsky, L., Garrett, W.S., Huttenhower,
1179 C., 2011. Metagenomic biomarker discovery and explanation. *Genome Biol.* 12, R60.
1180 <https://doi.org/10.1186/gb-2011-12-6-r60>
- 1181 Shameer, S., 2016. Haloalkaliphilic *Bacillus* species from solar salterns: an ideal prokaryote
1182 for bioprospecting studies. *Ann. Microbiol.* 66, 1315–1327.
1183 <https://doi.org/10.1007/s13213-016-1221-7>
- 1184 Shivlata, L., Tulasi, S., 2015. Thermophilic and alkaliphilic Actinobacteria: biology and
1185 potential applications. *Front. Microbiol.* 6. <https://doi.org/10.3389/fmicb.2015.01014>
- 1186 Srichandan, H., Mohapatra, R.K., Parhi, P.K., Mishra, S., 2019. Bioleaching approach for
1187 extraction of metal values from secondary solid wastes: A critical review.
1188 *Hydrometallurgy* 189, 105122. <https://doi.org/10.1016/j.hydromet.2019.105122>
- 1189 Straub, K.L., Benz, M., Schink, B., Widdel, F., 1996. Anaerobic, nitrate-dependent microbial
1190 oxidation of ferrous iron. *Appl. Environ. Microbiol.* 62, 1458–1460.
1191 <https://doi.org/10.1128/aem.62.4.1458-1460.1996>
- 1192 Sun, W., Xiao, E., Häggblom, M., Krumins, V., Dong, Y., Sun, X., Li, F., Wang, Q., Li, B.,
1193 Yan, B., 2018. Bacterial Survival Strategies in an Alkaline Tailing Site and the
1194 Physiological Mechanisms of Dominant Phylotypes As Revealed by Metagenomic
1195 Analyses. *Environ. Sci. Technol.* 52, 13370–13380.
1196 <https://doi.org/10.1021/acs.est.8b03853>

- 1197 Tian, T., Liu, Z., Zhu, F., Hartley, W., Ye, Y., Xue, S., 2020. Improvement of aggregate-
1198 associated organic carbon and its stability in bauxite residue by substrate amendment
1199 addition. *Land Degrad. Dev.* 31, 2405–2416. <https://doi.org/10.1002/ldr.3609>
- 1200 Ujaczki, É., Zimmermann, Y., Gasser, C., Molnár, M., Feigl, V., Lenz, M., 2017. Red mud as
1201 secondary source for critical raw materials - purification of rare earth elements by
1202 liquid/liquid extraction: Red mud as secondary source for critical raw materials. *J.*
1203 *Chem. Technol. Biotechnol.* 92, 2683–2690. <https://doi.org/10.1002/jctb.5289>
- 1204 Vaidya, S., Dev, K., Sourirajan, A., 2018. Distinct Osmoadaptation Strategies in the Strict
1205 Halophilic and Halotolerant Bacteria Isolated from Lunsu Salt Water Body of North
1206 West Himalayas. *Curr. Microbiol.* 75, 888–895. <https://doi.org/10.1007/s00284-018-1462-8>
- 1207
- 1208 Vind, J., Malfliet, A., Blanpain, B., Tsakiridis, P., Tkaczyk, A., Vassiliadou, V., Panias, D.,
1209 2018. Rare Earth Element Phases in Bauxite Residue. *Minerals* 8, 77.
1210 <https://doi.org/10.3390/min8020077>
- 1211 Wagg, C., Schläeppli, K., Banerjee, S., Kuramae, E.E., van der Heijden, M.G.A., 2019.
1212 Fungal-bacterial diversity and microbiome complexity predict ecosystem functioning.
1213 *Nat. Commun.* 10, 4841. <https://doi.org/10.1038/s41467-019-12798-y>
- 1214 Wang, L., Yin, S., Wu, A., Chen, W., 2020. Synergetic bioleaching of copper sulfides using
1215 mixed microorganisms and its community structure succession. *J. Clean. Prod.* 245,
1216 118689. <https://doi.org/10.1016/j.jclepro.2019.118689>
- 1217 Wang, M., Chen, S., Chen, L., Wang, D., 2019. Responses of soil microbial communities and
1218 their network interactions to saline-alkaline stress in Cd-contaminated soils. *Environ.*
1219 *Pollut.* 252, 1609–1621. <https://doi.org/10.1016/j.envpol.2019.06.082>
- 1220 Wee, S.K., Burns, J.L., DiChristina, T.J., 2014. Identification of a molecular signature unique
1221 to metal-reducing Gammaproteobacteria. *FEMS Microbiol. Lett.* 350, 90–99.
1222 <https://doi.org/10.1111/1574-6968.12304>
- 1223 Wickham, H., 2009. *ggplot2: Elegant Graphics for Data Analysis*, 2197-5736. Springer-
1224 Verlag, New York, NY.
- 1225 Wickham, H., François, R., Henry, L., Müller, K., 2021. *dplyr: A Grammar of Data*
1226 *Manipulation*.
- 1227 Williams, R.J., Howe, A., Hofmockel, K.S., 2014. Demonstrating microbial co-occurrence
1228 pattern analyses within and between ecosystems. *Front. Microbiol.* 5, 358.
1229 <https://doi.org/10.3389/fmicb.2014.00358>
- 1230 Wong, J.W.C., Ho, G., 1994. Sewage sludge as organic ameliorant for revegetation of fine
1231 bauxite refining residue. *Resour. Conserv. Recycl., Environmental biotechnology in*
1232 *waste treatment and recycling* 11, 297–309. [https://doi.org/10.1016/0921-3449\(94\)90097-3](https://doi.org/10.1016/0921-3449(94)90097-3)
- 1233
- 1234 World Aluminium, 2020. World Aluminium [WWW Document]. URL [www.world-](http://www.world-aluminium.org)
1235 [aluminium.org](http://www.world-aluminium.org) (accessed 6.8.21).
- 1236 Wu, H., Chen, L., Zhu, F., Hartley, W., Zhang, Y., Xue, S., 2020. The dynamic development
1237 of bacterial community following long-term weathering of bauxite residue. *J.*
1238 *Environ. Sci.* 90, 321–330. <https://doi.org/10.1016/j.jes.2019.12.001>
- 1239 Wu, H., Liao, J., Zhu, F., Millar, G., Courtney, R., Xue, S., 2019. Isolation of an acid
1240 producing *Bacillus* sp. EEEL02: Potential for bauxite residue neutralization. *J. Cent.*
1241 *South Univ.* 26, 343–352. <https://doi.org/10.1007/s11771-019-4006-x>
- 1242 Wu, H., Tang, T., Zhu, F., Wei, X., Hartley, W., Xue, S., 2021. Long term natural restoration
1243 creates soil-like microbial communities in bauxite residue: A 50-year field study.
1244 *Land Degrad. Dev.* 32, 1606–1617. <https://doi.org/10.1002/ldr.3728>

- 1245 X. Kong, Guo, Y., Xue, S., Hartley, W., Wu, C., Ye, Y., Cheng, Q., 2017. Natural evolution
1246 of alkaline characteristics in bauxite residue. *J. Clean. Prod.* 143, 224–230.
1247 <https://doi.org/10.1016/j.jclepro.2016.12.125>
- 1248 Xiao, X., Wang, M., Zhu, H., Guo, Z., Han, X., Zeng, P., 2017. Response of soil microbial
1249 activities and microbial community structure to vanadium stress. *Ecotoxicol. Environ.*
1250 *Saf.* 142, 200–206. <https://doi.org/10.1016/j.ecoenv.2017.03.047>
- 1251 Xue, S., Zhu, F., Kong, X., Wu, C., Huang, L., Huang, N., Hartley, W., 2016. A review of the
1252 characterization and revegetation of bauxite residues (Red mud). *Environ. Sci. Pollut.*
1253 *Res.* 23, 1120–1132. <https://doi.org/10.1007/s11356-015-4558-8>
- 1254 Yuan, Q., Wang, P., Wang, C., Chen, J., Wang, X., Liu, S., 2021. Indicator species and co-
1255 occurrence pattern of sediment bacterial community in relation to alkaline copper
1256 mine drainage contamination. *Ecol. Indic.* 120, 106884.
1257 <https://doi.org/10.1016/j.ecolind.2020.106884>
- 1258 Zech, W., Schad, P., Hintermaier-Erhard, G., 2014. Trockene Subtropen und Tropen, in:
1259 Zech, W., Schad, P., Hintermaier-Erhard, G. (Eds.), *Böden der Welt: Ein Bildatlas.*
1260 Springer, Berlin, Heidelberg, pp. 68–85. [https://doi.org/10.1007/978-3-642-36575-](https://doi.org/10.1007/978-3-642-36575-1_7)
1261 [1_7](https://doi.org/10.1007/978-3-642-36575-1_7)
- 1262 Zeng, J., Lou, K., Zhang, C.-J., Wang, J.-T., Hu, H.-W., Shen, J.-P., Zhang, L.-M., Han, L.-
1263 L., Zhang, T., Lin, Q., Chalk, P.M., He, J.-Z., 2016. Primary Succession of Nitrogen
1264 Cycling Microbial Communities Along the Deglaciated Forelands of Tianshan
1265 Mountain, China. *Front. Microbiol.* 7, 1353.
1266 <https://doi.org/10.3389/fmicb.2016.01353>
- 1267 Zhang, D., Chen, H., Nie, Z., Xia, J., Li, E., Fan, X., Zheng, L., 2020. Extraction of Al and
1268 rare earths (Ce, Gd, Sc, Y) from red mud by aerobic and anaerobic bi-stage
1269 bioleaching. *Chem. Eng. J.* 401, 125914. <https://doi.org/10.1016/j.cej.2020.125914>
- 1270 Zhang, K., Shi, Y., Cui, X., Yue, P., Li, K., Liu, X., Tripathi, B.M., Chu, H., 2019. Salinity Is
1271 a Key Determinant for Soil Microbial Communities in a Desert Ecosystem. *mSystems*
1272 4, e00225-18. <https://doi.org/10.1128/mSystems.00225-18>
- 1273 Zhang, Y., Qin, W., Wang, J., Zhen, S., Yang, C., Zhang, J., Nai, S., Qiu, G., 2008.
1274 Bioleaching of chalcopyrite by pure and mixed culture. *Trans. Nonferrous Met. Soc.*
1275 *China* 18, 1491–1496. [https://doi.org/10.1016/S1003-6326\(09\)60031-5](https://doi.org/10.1016/S1003-6326(09)60031-5)
- 1276 Zhu, F., Xue, S., Hartley, W., Huang, L., Wu, C., Li, X., 2016. Novel predictors of soil
1277 genesis following natural weathering processes of bauxite residues. *Environ. Sci.*
1278 *Pollut. Res.* 23, 2856–2863. <https://doi.org/10.1007/s11356-015-5537-9>
1279