

Efficiency and sustainability of gravel augmentation to restore large regulated rivers: Insights from three experiments on the Rhine River (France/Germany)

Valentin Chardon, Laurent Schmitt, Fanny Arnaud, Hervé Piégay, Anne

Clutier

To cite this version:

Valentin Chardon, Laurent Schmitt, Fanny Arnaud, Hervé Piégay, Anne Clutier. Efficiency and sustainability of gravel augmentation to restore large regulated rivers: Insights from three experiments on the Rhine River (France/Germany). Geomorphology, $2021, 380, 10.1016/j.$ geomorph. 2021.107639 . hal-03634707

HAL Id: hal-03634707 <https://cnrs.hal.science/hal-03634707v1>

Submitted on 13 Apr 2022

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

29

Abstract 30

Over the last two centuries, rivers were strongly regulated, inducing notable sediment starvation and flow reduction. This regulation altered aquatic habitats, biocenosis and ecosystem services. Gravel augmentation (GA) is increasingly promoted to restore bedload transport, rejuvenate bed sediments and diversify aquatic and riverine habitats. However, practical questions remain in terms of efficiency and sustainability. The objective of this study is to provide feedback from three GA experiments conducted along the Rhine River downstream of the Kembs dam (France/Germany). The restored sites were monitored over a time period from 1 to 7 years according to the sites so that the stockpile deposits (SDs) and channel responses were compared to highlight the strengths and limitations of such experiments. The monitoring was based on topo-bathymetry, bedload tracking and grain size surveys. Two-dimensional hydraulic modeling was also performed to link the SD erosional patterns to critical flow conditions. We show that (i) SD erosion is dependent on the grain size of the introduced sediments and the local flow velocities, (ii) sediment dispersion occurred mostly along the thalweg and increased over time, with a progressive decrease in the virtual velocities, (iii) bathymetric simplification and bed grain-size fining were observed when the downstream sediment diffusion began, but (iv) at mid-term (~ 5 years) local channel diversifications occurred due to the fragmentation of the sediment waves with deposition occurring preferentially on riffles, and (v) new sediment starvation conditions 31 32 33 34 35 36 37 38 39 40 41 42 43 44 45 46 47 48

appeared when the sediment waves travelled downstream in the absence of an upstream sediment supply. The results allow us to propose operational recommendations and management scenario to improve both the efficiency and the sustainability of GAs conducted in large regulated rivers. Key-words: Gravel augmentation, Restoration, Large rivers, Geomorphic monitoring, 49 50 51 52 53

Feedbacks, Operational guidelines 54

1. Introduction 55

Numerous large rivers have been strongly affected by regulation (Kondolf et al., 2014). Damming is one of the most common infrastructures affecting both water and sediment fluxes, and it often generates sediment starvation below the dams (Kondolf 1997; Wheaton et al., 2004; Petts and Gurnell 2005; Magilligan and Nislow 2005). These conditions cause severe and irreversible geomorphic alterations, such as bed incision, bed armoring and active channel narrowing, which disturb both aquatic and riverine habitats and impair ecosystem services (Kondolf, 1997; Cooper et al., 2017). Moreover, dams accumulate sediments, reducing the reservoir capacities and their life span (Kondolf et al., 2014). Several management strategies were tested to improve the transparency of dams and partially restore downstream bedload transport. These strategies include permanent remediation solutions, such as sediment bypass and sediment pass-through (Sumi et al., 2004). Mechanical excavation of reservoirs and downstream sediment releases have also been used (Kondolf et al., 2014). Dams are also removed, as in the USA where 1,100 dams have been removed in small to moderate watersheds (10 to 1000 km²) since the 1970s (Folev et al., 2017). Alternative measures consist of mitigating sediment starvation below dams, e.g., by reactivating bank erosion after removing riprap protection (Staentzel et al., 2018a), 56 57 58 59 60 61 62 63 64 65 66 67 68 69 70 71

reactivating old side channels, remobilizing sediments stored in floodplains (Landon, 2008; Thorel et al., 2018), and implementing gravel augmentations (GAs) (Kondolf et al., 2014; Ock et al., 2013). 72 73 74

GA is becoming a common practice because it provides rapid geomorphic improvement by diversifying bedforms, grain size, and flow conditions (Wheaton et al., 2004.a). First, GAs were conducted in the USA during the 1960s in the form of an ambitious program for fish spawning rehabilitation along 13 dammed rivers in California (Kondolf and Matthews, 1991; Pasternack et al., 2004). In some cases, the GAs were combined with the mechanical creation of in-channel geomorphic features, such as riffles (Pasternack et al., 2004; Ock et al., 2015). Between 1968 and 2000, 316,000 m^3 of sediments were introduced for a total cost of \$3,000,000 (Rollet 2007). More recently, GAs were also implemented below 15 Japanese dams (Sumi et al., 2004; Ock et al., 2013). In Europe, the first GAs were realized in 1978 on the Rhine River below the Iffezheim dam (vearly mean volume of 170,000 m^3) to counteract bed incision and maintain a stable water level for navigation (CHR, 2009). GAs were also employed on gravel-bed alpine rivers (Brousse et al., 2018; Heckmann et al., 2017), boulder-bed steep reaches (Downs et al., 2016) and large regulated rivers (Schälchli et al., 2010; Klösch, 2011; Arnaud et al., 2017). Ock et al. (2013) noted several strategies employed to inject sediment into rivers as in-channel bed stockpile, high-flow stockpile, point bar stockpile and high-flow direct injection. 75 76 77 78 79 80 81 82 83 84 85 86 87 88 89 90

The restoration of fluvial macroforms by GA improves river ecological quality and functionalities. For example, Ock et al. (2015) showed on the Trinity River that in-channel bar development following GA enhanced hyporheic fluxes. Zeug et al. (2013) showed on the Lower American River that both salmon and trout populations increased following GA due to 91 92 93 94

an increase in spawning areas. Staentzel et al. (2018b) showed that the taxonomic richness of macroinvertebrates increased after a GA experimental operation on the Rhine. Geomorphic assessment of GA is generally carried out over short-term periods (1 to 3 years) and usually focuses on channel topo-bathymetric changes and sediment mobility [\(Table 1](#page-6-0)). Grain size diversification and fining are expected, but the results are infrequent because measuring underwater grain size is still challenging. On the other hand, field studies and modeling approaches evaluating the critical conditions for eroding the stockpile deposit (SD) are needed to improve such restoration practices. Numerical modeling studies have shown that various geomorphic responses can be obtained depending on the SD geometric configuration (Battisacco et al., 2016; Juez et al., 2016). This was recently validated in the field by Stähly et al. (2019). To optimize the channel response to GA, additional insights into channel grain size, SD location and geometry and local hydraulic conditions are needed. Implementing a GA operation is not risk-free. Gaeuman et al., 2017 presented three different modes of sediment pulses which are dispersion, translation/dispersion and fragmentation. In case of sediment translation, large amounts of sediment may travel quickly and impact navigable reaches downstream of the bypassed reaches. Moreover, bedload transport reactivation may destabilize the armor layer (Koll et al., 2010) and bank protections (Arnaud et al., 2017). The proper hydrological functioning of floodplain excavations along incised rivers for flood retention purposes may also be disturbed by bed elevation changes. The aim of this paper is to present a comparative analysis of three GAs based on short/midterm (from 2009 to 2017) geomorphic monitoring conducted on a bypassed reach of the Rhine River downstream of the Kembs dam (France/Germany). The objectives are to (i) define the best hydraulic conditions to erode, and therefore remobilize rapidly, the SD, (ii) 95 96 97 98 99 100 101 102 103 104 105 106 107 108 109 110 111 112 113 114 115 116 117 118

evaluate channel responses in terms of sediment mobility, morphodynamics and bed grain size as well as the corresponding temporal scales of these evolutions, (iii) perform a risk analysis to avoid any negative impacts on human and management issues such as bed incision, navigation and flood retention, to consolidate some preliminary results of Arnaud et al. (2017), and (iv) provide some operational guidelines and management scenario for enhancing efficiency and sustainability of GA-based restorations of large rivers. 119 120 121 122 123 124

Table 1: Characteristics of some GA monitoring sites. (1) corresponds to a predictive model 125

(SRH-2D) to design SDs. 126

2. Materials and methods 127

2.1 Study sites 128

The Rhine River is the third largest river in Europe, with a drainage basin of 185 000 km² and 129

- a river course of 1250 km long. The Upper Rhine, located between Basel and Bingen-am-130
- Rhein, exhibits a nivo-glacial hydrological regime ([Figure 1](#page-9-0).a; Schmitt et al., 2018). The mean 131
- annual discharge is 1059 m $3/$ s at the Basel gauging station (Wantzen et al., 2019). In this 132
- riffle-pool reach, the Rhine is bypassed by the 50 km-long "Grand Canal d'Alsace" (GCA) 133

between the Kembs diversion dam and Breisach-am-Rhein [\(Figure 1](#page-9-0).b). The minimum flow into this bypassed reach varies between 52 $\text{m}^3\text{.s}^{\text{-}1}$ (winter) and 150 $\text{m}^3\text{.s}^{\text{-}1}$ (spring-summer) in relation to the flow requirements for aquatic and riparian communities (Garnier and Barillier, 2015). Spills occur when the Rhine discharge exceeds 1400 m^3 . s⁻¹ at Basel gauging station, which corresponds to the maximal capacity of the GCA power plants ([Figure 1](#page-9-0)). This discharge was exceeded an average of 69 days per year between 1932 and 2011 (Arnaud et al., 2015a). This Rhine section exhibits poor ecological functionalities due to severe geomorphological alterations induced by successive engineering works: damming and bypassing, as well as channelization and construction of groin fields since the 19th century (Piégay et al., 2011). Channelization and groin fields induced bed incision, bed armoring and floodplain and side channel disconnection, whereas damming increased active channel narrowing and bedform simplification due to a reduction in morphogenic discharges without any upstream sediment supply (Arnaud et al., 2019). 134 135 136 137 138 139 140 141 142 143 144 145 146

Since 2010, three GAs have been performed in the first 10 km of the Old Rhine (Figure 1.b) to restore bedload transport, promote bed grain size diversification and increase the heterogeneity of the channel bathymetry to enhance the ecological functionalities of the Old Rhine. The GAs were implemented following the point bar SD method (Ock et al., 2013). In October 2010, a first experimental GA, named Interreg, was achieved within an Interreg program. A SD was implemented within a concave bend and separated from the right bank (Arnaud et al., 2017). Within the framework of the Kembs dam relicensing, Kembs I1 and Kembs I2 GAs ([Figure 1](#page-9-0).c-d-e) were achieved by Électricité de France in March-April 2015 and March-April 2016, respectively (Garnier and Barillier, 2015). The SDs were located in convex locations. The Kembs I1 SD was connected to the left bank, whereas the Kembs I2 SD was separated between 5 and 10 m from the left bank. The sediments were extracted from 147 148 149 150 151 152 153 154 155 156 157

the floodplain close to the SD injection areas. The SD volumes equal to 23 000 m^{3} . 13 000 m^{3} 158

- and 30 000 m^3 (Table 2). These volumes were consistent with the mean annual bedload 159
- transport capacity estimated to be 23,000 m^3 . yr⁻¹ for this part of the reach (El Kadi 160
- Abderrezzak, 2009). The critical discharge for bedload entrainment Q_c was estimated to be 161
- 550 m^3 . s^1 (El Kadi Abderrezzak, 2009). 162

SD	Implementation	KP (km)	Volume (m ³)	Length (m)	Width (m)	Height (m)	Mean D_{50}	Mean D ₉₅	D_{95}/D_{50}
Interreg	October 2010	$182.4 - 183.1$	23,000	620	$11.00 - 15.00$	$2.00 - 3.50$	35.00	100.60	3.28
Kembs $\mathsf{I}1$	March - April 2015	174.10 - 174.525	13,000	425	$10.00 - 40.00$	$2.00 - 3.50$	28.00	181.00	6.40
Kembs 12	March - April 2016	176.20 - 176.66	30,000	460	$15.00 - 40.00$	$2.00 - 4.50$	18.50	105.00	5.95

Table 2: Characteristics of the three GAs performed on the Old Rhine. See Section 2.2.4 for 163

details on grain size measurements. The sorting coefficient was based on (Inman 1952). 164

166

Figure 1: (a) Location of the Old Rhine river in the Rhine Basin, (b) location of the river

network, the hydraulic structures, the gauging stations and the 3 GAs performed along the 167

- studied reach, (c) Kembs I2 SD in progress, (d) Kembs I2 SD completed and (e) residual 168
- Kembs I2 SD after a Q_4 flood in May 2016. White arrows indicate the flow direction. 169

2.2 Data collection 170

- 2.2.1 Monitoring framework 171
- Geomorphic monitoring was applied for each GA at two spatial scales: (i) the SD and (ii) the 172
- downstream channel reach, extending 4 to 12 km long depending on the sites. The 173

monitoring followed a before-after temporal framework that lasted two, three and seven years on Kembs I2, Kembs I1 and Interreg GAs, respectively, corresponding to three, four and seven monitoring periods ([Figure 2](#page-11-0)). The beginnings and ends of each Period *Pⁱ* correspond to the S_{i-1} and S_i Survey dates, respectively. For each GA, P_0 is the period before the implementation of the SD, and $P_{0 bis}$ is the period after the SD implementation and before the first floods. The following P_i are a series of periods following this first flood. Between one and three significant hydrological events occurred for each of these periods, and the flood intensity ranged from Q_1 (1080 m³.s⁻¹) to Q_{15} (2480 m³.s⁻¹). The monitoring methods combined topo-bathymetric surveys, grain size surveys, sediment tracking and twodimensional hydraulic modeling. The Interreg monitoring presented in this paper ($P_{5}\.\dot{\!\iota}$ is in continuity of the work of Arnaud et al. (2017), who characterized periods \overline{P}_0 to $\overline{P}_4.$ 174 175 176 177 178 179 180 181 182 183 184 185

Figure 2: (a) Temporal monitoring framework related to the maximum daily flow in the Old Rhine, obtained by subtracting 1400 m^3 .s⁻¹ from the discharge at the Basel gauging station and (b) spatial monitoring framework. In (a) circles with a star indicate the volumetric samplings conducted on the SDs. S_x indicates the Survey date and P_x the monitoring Period between two surveys. See Table 3 for details on topo-bathymetric surveys.

2.2.2 Topo-bathymetric surveys 193

Topo-bathymetric surveys were conducted to assess accurately: (i) the erosion of SDs after first floods and (ii) the habitat diversification along the reaches located downstream the SDs. For Kembs I1 and Kembs I2, S_0 was surveyed in February 2015 and November 2015, respectively, based on cross-section profiles at 5 to 50 m-spaced intervals using a Tritech PA500 single beam echo-sounder ([Table 3](#page-13-0)). An airborne topo-bathymetric LiDAR survey was performed on Kembs I1 in March 2015 (S_{0bis}), followed by another survey in March 2016 that covered both Kembs I1 (S_1) and Kembs I2 $(S_{0\,bin})$ [\(Figure 2;](#page-11-0) Table 3). Another survey in March 2017 covered Kembs I1 (S_2) , Kembs I2 (S_1) and Interreg (S_5) . All LiDAR surveys were performed at flight elevations close to 400 m using the Optec Titan sensor and during low water turbidity. The spot size was estimated to 0.08-0.12 m in above-water areas and 0.20 m in underwater conditions. The resolution of the LiDAR surveys ranged between 20 and 40 pts/ $m²$ for emerged areas and 5 to 10 pts/ $m²$ in underwater conditions, depending on the water depth. Due to a sensor malfunction in March 2015, no data were recorded in underwater areas deeper than one meter for this survey. We estimated that about 70 % of the submerged bed area was deeper than 1 m**.** For above-water surfaces the range of mean error was equal to -0.032m to 0.044m and the mean root mean square error was equal to 0.042m for the three airborne topo-bathymetric surveys. During the survey of March 2016, five crosssection profiles were performed simultaneously to control the accuracy of bathymetric data (see table 3). The range of mean error was estimated to -0.086m to +0.083m and the root mean square error equal to 0.083m.". 194 195 196 197 198 199 200 201 202 203 204 205 206 207 208 209 210 211 212 213 214

215

216

Table 3 : Characteristics of topo-bathymetric surveys achieved in this study. 218

219

2.2.3 PIT-tags monitoring 220

We equipped particles with Passive Integrated Transponders (PIT-tags) to achieve several purposes: (i) evaluate the erosion of SDs according to flow conditions and particle grain size, and (ii) characterize the sediment mobility (travel distances and virtual velocities) and the mode of sediment pulse (dispersion, translation, fragmentation) along the three downstream reaches. We used 23-mm PIT tags for the Interreg GA, a combination of 23-mm and 32-mm tags for Kembs I1, and 32-mm tags for Kembs I2. The PIT tags were developed by Texas Instrument Enterprise and have a maximum vertical reading range of 0.95 m (Arnaud, et al., 2015). For Interreg, bedload tracers were equally distributed between the three principal grain size classes, whereas the tracer grain size distributions (GSD) correspond to the GSD of a gravel bar near Kembs for Kembs I1 and Kembs I2. The PIT tags were placed during $S_{0\,bis}$ following various injection protocols depending on the GA [\(Table 4\)](#page-14-0). Each 221 222 223 224 225 226 227 228 229 230 231

individual tracer and tracer cluster was geopositioned with a global positioning system 232

Trimble GeoXT 6000 (XY accuracy = 0.1 m). 233

Table 4 : Characteristics of bedload tracer placed for each GA. 234

235

For each bedload tracking survey, the tracers were detected using an antenna connected to a reader system developed by CIPAM Corporation. For emerged and water areas less than 0.6 m deep, a 0.46 m-diameter loop antenna was used, and the tracer IDs were manually recorded with the GPS. In water areas deeper than 0.6 m, a rectangular antenna 1.2 m wide was pulled by boat in the downstream direction, and the same GPS was used in tracking mode (continuously recording each second) to map the tracked areas and determine the position of each detected tracer by synchronizing the on-board GPS and the chronometer. The accuracy of the tracer location was estimated to be 2 m in emerged conditions and water conditions less than 0.6 m while the accuracy was estimated to 5 m in water areas deeper than 0.6 m. 236 237 238 239 240 241 242 243 244 245

2.2.4 Grain size monitoring 246

We performed volumetric samplings to measure the grain size of the injected sediments which composed Kembs I1 (n=2; *S1*) and Kembs I2 (n=2; *S0bis*) SDs. We sampled a minimum mass of sediment according to the grain size of the surface layer, following the equation proposed by Church et al., (1987): 247 248 249 250

$$
mmw = 2,882 * Dmax - 47,6
$$
\n(1)

where mmw corresponds to the minimum mass weight and D_{max} the mean diameter of the 30 coarsest particles of the surface layer. 251 252

For the Interreg SD (S_0) , volumetric sampling was performed in the floodplain sediment excavation area (Dittrich et al., 2010). In addition, surface changes in bed grain size were evaluated after the first morphogenic flood (S_1) on the three emerged residual SDs using the Wolman (1954) technique, to evaluate if a coarser surface layer has been developed, which may reduce the residual SD erodibility by the subsequent floods. Moreover, surface grain size changes were also evaluated along the downstream reaches, in both emerged and underwater (water depth < 0.6 m) areas of the active channel, to verify if the GAs diversified and lowered the grain size distributions. For the Interreg downstream reach, the grain size sampling in S_5 (n=414) followed the same protocol as Arnaud (2012), who characterized S_0 $(n=310)$, S_3 (n=192) and S_4 (n=143): random sampling of 10 particles at each node of quadrats (10 x 10 m). The surface sampled in *S0*, *S3*, *S4* and *S5* is equal to 14, 8, 6 and 14% of the monitoring reach, respectively. These relatively low sampled surfaces are explained by the water level which exceeded 0.6 m in the major part of the monitoring reach (> 80%). For Kembs I1 and I2 downstream reaches, an underwater photosieving protocol was applied along transects spaced at 60 m intervals, with sampling points spaced at 5 - 10 m intervals. This photosieving strategy was adopted because the water depth exceeded 1 m in most of 253 254 255 256 257 258 259 260 261 262 263 264 265 266 267 268

the study areas and reached 4 m locally. For S_0 at Kembs I2, the data collection (n=240) was performed by diving and using a GoPro Hero4+ camera with a constant height equal to 0.52 m. For *S*1, high definition (1080p) subaquatic films were obtained along the same transects, by using an Olympus TG-4 camera placed perpendicularly on an aluminum telescopic bar, all deployed from a boat. A rule was placed at the bottom of the telescopic bar to determine the pixel size. Digital images (n=324) were extracted from the subaquatic films using VirtualDub software v.1.10.4. Chardon et al. (2020) showed the capacity of photosieving in underwater conditions to quantify bed grain size. Fisheye deformation on all photos was corrected by using a correction algorithm with Photoshop software. Photos were then cropped to eliminate the corners impacted by the correction. We digitized the ten largest particles in each photograph and calculated the D_{50} , as this metric is close to the D_{95} of the total apparent particles on each photography. Indeed, Houbrechts et al. (2011) demonstrated the advantage to use the mean diameter of ten largest particles to estimate the D_{95} of the grain size distributions. We assumed that this metric is sensitive to bed grain size fining. 269 270 271 272 273 274 275 276 277 278 279 280 281 282 283

2.2.5 Two-dimensional hydraulic modeling 284

We evaluated the local hydraulic conditions at the SDs by performing two-dimensional hydraulic unsteady flow modeling to compare the intensity of the erosional processes between them. We used the HEC-RAS v.5.0.3 software between the Kembs dam (KP 174.00) and the Rheinweiler gauging station (KP 186.2), which is located 4 km downstream of the Interreg GA ([Figure 1](#page-9-0)). The model contains 146 619 mesh cells with an average cell area of 16 m^2 . The digital elevation model (DEM) used is based on the airborne topo-bathymetric LiDAR of March 2017 because it included a detailed topography of the channel for the entire study reach. The geometries of the three SDs measured in *S0bis* were integrated into the 285 286 287 288 289 290 291 292

DEM. Water levels and instantaneous discharges recorded at the Rheinweiler gauging station during the flood event between May and June 2017 (time step = 15 min) were used as downstream boundary conditions because these discharges ranged between 52 and 2000 $\mathsf{m}^3.\mathsf{s}^{\text{-}1}$, which included the peak discharge of periods P_1 for the three SDs [\(Figure 2\)](#page-11-0). We assumed that no significant discharges were provided by tributaries along the modeling reach . Only a small tributary is present, the Kander River, for which the 100-year flood is estimated to 56.4 m^3 .s⁻¹ (information from the RPF). Thus, information from the Rheinweiler gauging station were also used as upstream boundary conditions. The Manning's coefficients implemented in the model are close to those calculated by Béraud (2012) and equal to 0.035 for the active channel and 0.066 for the vegetation areas on the groin fields. Modeling output provided the maximal flow velocities for the peak discharges that occurred during 293 294 295 296 297 298 299 300 301 302 303

periods P_1 for the three SDs. 304

2.3 Data processing and analysis 305

2.3.1 Morphological changes in the SDs 306

DEMs were produced from point clouds obtained by the airborne topo-bathymetric LiDAR and the terrestrial topo-bathymetric cross-sections datasets conducted along Kembs I1 and Kembs I2 using the 3D Analyst toolbox of ArcMap v.10.3 (cell size = 0.5 m). Cross sections spaced at 25 m intervals were extracted from the DEMs using the ETGeoWizards add-on to calculate the surface eroded on each SD and related to the modelled flow conditions. For the Interreg SD, surface eroded were directly calculated from topo-bathymetric cross-sections $(Arnaud et al., 2017)$. We calculated for each profile the section area at *Si* (in m²): 307 308 309 310 311 312 313

Section area_(S_i) =
$$
\sum_{i=1}^{n} \lambda(x_{i+1} - x_i \lambda) \vee \lambda \left(\frac{Z_{i+1} + Z_i}{2} \right) \lambda
$$
 (2)

314

- points (m). n defines the number of points. The reference used to calculate the section area 316
- was the toe at the left bank of the SD at S_0 . 317

- From the previous equation, we estimated the resulting eroded surface for each cross-318
- section for each SD (in m^2) between two surveys: 319

$$
Froded surface_{S_i \to S_{i+1}} = Section \, area_{[S_{i+1}]} - Section \, area_{[S_i]}
$$
 (3)

We estimated the elevation changes uncertainties between S_{obs} and S_x for Kembs I1 and Kembs I2 SDs by estimating the combined elevation change errors of two paired DEMs from airborne topo-bathymetric LiDAR and according to the equation of Brasington, Langham, and Rumsby (2003): 320 321 322 323

$$
\sigma_{\text{diff}} = \sqrt{(\theta \lambda \dot{\delta} S_i)^2 + (\theta \lambda \dot{\delta} S_{i+1})^2 \dot{\delta} \dot{\delta}}
$$
\n(4)

where $\theta_{_{S_i}}$ and $\theta_{_{S_{i+1}}}$ are the estimated registration errors at S_i and S_{i+1}, respectively. The estimated registration error was estimated to 0.1 m on the Z-axis for all surveys. Following Bennett et al. (2012), a statistical t-test was performed to identify significant elevation changes for Kembs I1 and Kembs I2 SDs, between two paired DEMs from airborne topo-bathymetric LiDAR and for a given, when $|t|>1$, corresponding to 66% confidence interval: 324 325 326 327 328 329

$$
t = \frac{Z_{S_{i+1}} - Z_{S_i}}{\sigma_{\text{diff}}}
$$
\n
$$
\tag{5}
$$

Where $Z_{_{S_i}}$ and $Z_{_{S_{i+1}}}$ are the mean elevation of the DEM cell at S_i and S_{i+1}. From this equation, the elevation changes smaller than $σ_{diff}$ were considered uncertain and not integrated in the analysis of elevation changes (Bennett et al., 2012) but also for sediment budgets. 330 331 332

We calculated the sediment budgets for the Kembs I1 and Kembs I2 SDs from the DEM of Differences (DoD): 333 334

$$
\Delta V_{S_i \to S_{i+1}} = \sum_{i=1}^{n} A * (Z_{S_{i+1}} - Z_{S_i})
$$
\n(6)

where A is the area of the DEM cell, and $Z_{_{S_i}}$ and $Z_{_{S_{i+1}}}$ are the mean elevations of the DEM cell at S_i and S_{i+1} , respectively. 335 336

We estimated the volume errors for Kembs I1 and Kembs I2 SDs following the equation 337

developed by (Lane et al., 2003): 338

$$
\sigma_{v} = d^{2} \cdot \sqrt{n} \cdot \sigma_{diff} \tag{7}
$$

where *d* is the cell size of the DEM (m) and *n* is the number of cells in each area of 339

signification elevation changes identified by the statistical t-test. For Interreg SD, see Arnaud 340

et al. (2017) for the method used to calculate both the sediment budget and the volume 341

errors. 342

2.3.2 Suitable areas to implement SDs 343

We identified suitable areas in terms of best hydraulic conditions to implement future SDs in 344

order to minimize deposit fixation because the success of GA is dependent on SD erosion. 345

This study was performed for frequent peak flood equal to the Q $_2$ (1417 $\mathrm{m}^3.\mathrm{s}^{-1}$). We 346

developed a suitability index ratio (SIR) for each 500 m length of the active channel and 347

calculated as follows: 348

$$
Suitability Index Ratio = \frac{Area\ of\ flow\ velocity \ge 2.5 m.s^{-1}}{Area\ of\ active\ channel}
$$
 (8)

- We used a flow velocity threshold equal to 2.5 m.s⁻¹ because it was found to be the minimal 349
- velocity to erode significantly the SDs (see section 3.1.3). 350
- 2.3.3 Tracer mobility analysis 351
- We evaluated the mean tracer cloud travel distance for each period *Pi* using the tracer cloud centroid metric CP_(Si) (Piégay et al., 2016; Arnaud et al., 2017): 352 353

$$
CP_{(Si)} = \frac{\sum_{i=1}^{nrecovered} X_i}{n_{recovered}}
$$
 (9)

where *Xi*is the longitudinal position (in m) of each tracer linearly referenced along the active channel centerline. For Interreg, the first 12 m of the SD were excluded from the analysis because the tracers do not move. For Kembs I1, tracers detected in the residual SD were not integrated into the calculation of the tracer cloud centroid metric. For each period $P_{\scriptscriptstyle{b}}$, the travel distance of the tracer cloud centroid (in m) was calculated as follows: 354 355 356 357 358

$$
L_{cp} = CP_{(Si+1)} - CP_{(Si)}
$$
\n
$$
\tag{10}
$$

Then, we calculated the virtual velocity (in m.s⁻¹) for each period $P_{\it i}$ by using the travel distance of the tracer cloud centroid: 359 360

$$
V_{pi} = \frac{L_{CP}}{t_{\geq Q_c}}
$$
\n⁽¹¹⁾

where *t ≥Q^c* is the time during which the discharge is greater than or equal to Qc (Church and Hassan, 1992). To calculate this time duration, we used the instantaneous discharge estimated at the Basel gauging station every 10 min and compared it to *Q^c* . 361 362 363

- In order to study the relation between the sediment mobility and the flow competence, we used the two metrics calculated for each period \overline{P}_i as follows: 364 365
- (i) The excess unit stream power: 366

$$
\omega_{p_{(i)}} = \frac{\rho_w g \left(Q_p - Q_c \right) S}{W} \tag{12}
$$

where ρ_w is the water density (kg.m⁻³), g is the acceleration due to gravity (m.s⁻²), Q_p is the maximum peak discharge (m 3 .s⁻¹), Q_c is the critical discharge (m 3 .s⁻¹), S is the average local bed slope (m.m-1) estimated from topo-bathymetric surveys, and *W* is the mean active channel width (m). 367 368 369 370

(ii) The cumulative excess stream energy (Schneider et al., 2014) which assess the excess stream energy for an entire period P_i : 371 372

$$
\sum \left(\omega_{(t)} - \omega_c \right) \tag{13}
$$

where $ω_{[t]}$ is the stream energy for a discharge value greater than the critical discharge and 373

 $\omega_{\vert_{c\vert}}$ is the stream energy for the critical discharge value. (J.m⁻²). As the time-step 374

measurement at the Basel gauging station is 10 min, we multiplied the excess stream energy 375

by 600 s to calculate all excess stream energy of the period time between two discharges 376

- measurements and cumulated all values for the entire period P_i. 377
- In order to identify partial or full bedload mobility, we calculated the mean dimensionless 378
- travel distance of the tracer cloud centroid for each grain size class as follows: 379

Mean dimensionless travel distance
$$
\dot{\iota}
$$
 $\frac{L_i}{L_{d50}}$ (14)

where L_i is equal to the mean travel distance of the tracer cloud centroid for each grain size class of the recovered tracers, and L_{d50} is the mean travel distance of the median grain size class of the recovered tracers for each survey. 380 381 382

383

2.3.4 Channel geometry diversification analysis 384

The evolution of the channel geometry in response to GA operations was assessed using the cross-section diversity index (CSD) developed by Gostner et al. (2013) from topo-bathymetric cross sections along the downstream channel reaches: 385 386 387

$$
CSD = \frac{\sum_{i=2}^{n} |\Delta Z_i|}{\sum_{i=1}^{n-1} Y_i}
$$
 (15)

with

$$
\left|\Delta Z_{i}\right|=Z_{i-1}-Z_{i}\tag{16}
$$

|*∆Zi*| is the absolute elevation difference between two consecutive points on a cross section and \overline{Y}_i is the distance between them. The CSD calculation for Kembs I1 and Kembs I2 in S $_{\rm o}$ was based on cross sections surveyed by the RPF in 2009 (not indicated in [Figure 2](#page-11-0)) because the LiDAR surveys of 2015 were only extended 200 m downstream from the two SD sites [\(Figure 2.](#page-11-0)b). We also used topo-bathymetric datasets for the period 1985/93-2009 provided by the RPF to compare the CSD of GAs with longer-term changes in the Old Rhine. 388 389 390 391 392 393

3. Results 394

3.1 Monitoring of the stockpile deposits 395

3.1.1 Grain size comparison and evolution 396

The D₅₀ calculated from the volumetric samples in S_{0bis} was globally similar for the Interreg and Kembs I1 SDs, and lower for the Kembs I2 SD ([Table 2](#page-8-0)). Regarding the D_{95} , introduced sediments were coarser for Kembs I1 than for the two other SDs. The range of the S_1/S_{Obs} D₅₀ ratio was 1.36 to 2.67, 0.45 to 1.27 and equal to 2.35 for Kembs I1, Interreg and Kembs I2, respectively. The surface grain size of the upstream and downstream parts of the Kembs I1 residual SD was very close to the coarse surface layer GSD of a historical gravel bar located at KP 178.2 (D84 = 90.5 mm), indicating partial mobility conditions [\(Figure 3.](#page-23-0)b). For Interreg and Kembs I2, no coarse layer was appeared comparatively with the surface GSD of the historical gravel bar (Figure 3.a). 397 398 399 400 401 402 403 404 405

Figure 3: Surface grain size of the three residual SDs in S_1 (a) for Interreg (n=4) and Kembs I2 $(n=1)$, and (b) for Kembs I1 (n=6). Gray lines indicate the surface grain size of a gravel bar 407 408

Figure 4: (a) to (c) DEM of differences from *S0bis* to *S1* for the Kembs I1, Kembs I2 and Interreg SDs, with the location of mobile (red) and stationary tracers (black). Boxplots correspond to the elevation changes along each SD surface from *S0bis* to *S1*. Intervals correspond to the 25th and $75th$ percentiles, and the lower and upper whiskers correspond to 1.5 times the interval quartile range. Black lines correspond to the median values, and the red points correspond to the mean values. (d) to (f) three cross sections extracted from DoD (cross symbol on (a) to (c)) illustrate the morphological changes for each SD from S_0 to S_1 . 424 425 426 427 428 429 430

3.1.3 Relations between flow conditions and morphological changes of the SDs 431

- We explain the significant differences of the eroded volumes of the SDs due to significative 432
- differences on local hydraulic conditions. Statistically significant differences in modeled flow 433

velocities were found between the three SDs during P_1 (ANOVA-test, P-value = 2e-16). Velocities were lower for Kembs I1 (μ =2.06 m.s⁻¹ ± 0.24 m.s⁻¹) compared to those for Interreg $(\mu=2.20 \text{ m.s}^{-1} \pm 0.26 \text{ m.s}^{-1})$ and Kembs I2 $(\mu=2.52 \text{ m.s}^{-1} \pm 0.41 \text{ m.s}^{-1})$. A statistically significant relation at the 95% confidence level exists between the eroded areas and the maximum velocity areas calculated on each cross section (R^2 = 0.63; P-value = 4.9e-10; [Figure 5](#page-26-0).a). Using the maximum velocity as a proxy of the measured eroded areas, the root mean square error (RMSE) and the mean absolute error (MAE) between the predicted eroded areas and the measured eroded areas were equal to 5.37 m^2 and 4.31 m^2 , respectively [\(Figure 5](#page-26-0).b). 434 435 436 437 438 439 440 441

Figure 5: (a) Measured eroded areas according to the maximum velocity calculated for peak floods during P_1 along each SD, and (b) predicted eroded areas by maximal velocities compared with the measured eroded area (m²). The dashed line corresponds to the 1:1 line. The recovery rates of the tracers tracked in S_1 were equal to 42%, 33% and 43% for Kembs 443 444 445 446

- I1, Kembs I2 and Interreg, respectively. For all SDs, the percentage of mobile tracers ranged 447
- from 68% to 87%. Based on logistic regressions, we observed that the tracer movement 448
- probability is linked to the flow velocity (P-value < 0.0001). Partial tracer mobility occurred 449

3.1.4 Assessment of suitable areas for implementing SDs according to flow velocity In order to define the best areas in terms of hydraulic conditions for implementing future GAs on the Old Rhine, we used the empirical relationship found between the measured eroded areas and the maximum flow velocity. The critical flow velocity was fixed to 2.5 m.s⁻¹ as shown in Section 4.1.3 ([Figure 5](#page-26-0).a and [Figure 6](#page-27-0)). For a modeled discharge equal to the Q_2 flood, the mean value of the Suitability Index Ratio (SIR) calculated on elementary reach of 456 457 458 459 460 461

- 53
- by flow velocities lower than 2.5 m.s⁻¹ with a much lower mobility for the coarsest fraction 450

500 m length is equal to 0.59 with minimum and maximum values of 0.11 and 0.93, respectively ([Figure 7](#page-28-0)). The majority of the upstream reaches of Kembs I2 were characterized by values lower than the mean SIR and inversely for the downstream reaches of Kembs I2. Moreover, our analysis showed that the velocity values were systematically lower than 2.5 $ms⁻¹$ along the emerged bars at low flow for this modeled discharge. 462 463 464 465 466

467

Figure 7: Longitudinal distribution of the SIR value for a Q_2 flood along the bypassed reach for each elementary reach of 500 m length. The blue line corresponds to the mean value of the SIR, black arrows locate the three SDs and gray bars are the reaches covered (completely or partly) by the SDs. 468 469 470 471

3.2 Monitoring of the downstream channel reaches 472

3.2.1 Assessment of tracer mobility 473

The tracer recovery rates ranged between 11% and 43% on the three studied reaches over 474

- the monitoring periods (Table 3). The recovery rates globally decreased over time, which 475
- was notably due to the decrease in the sampling pressure related to the tracer cloud 476
- propagation ([Table 5](#page-30-0)). Translation occurred in the upstream part of the Interreg reach during 477
- P_1 and P_2 ([Figure 8](#page-31-0).a-b). During P_3 and P_4 , tracers travelled through a clear translation-478

Table 5 : Statistical parameters for each sediment tracking survey conducted on each GA. The Interreg results between $P_{\rm 1}$ and $P_{\rm 4}$ were provided 492

by Arnaud et al. (2017). 493

Figure 8: Probability density function of the longitudinal position of the tracers recovered for each survey on the (a-f) Interreg, where gray columns locate riffles, (g-i) Kembs I1 and (j-k) Kembs I2 downstream reaches.

The tracer cloud centroids travelled between 101 m and 390 m for moderate floods (Qix \leq Q4) on the three GA reaches. During the highest flood in May 2013, the tracer cloud centroid

Figure 9: (a) Virtual velocity of the tracer cloud centroids for each *Pi* according to the cumulative excess stream energy and (b) virtual velocity of the tracer cloud centroids composed of tracers whose grain size ranges between 0.75 and 1.25 of the channel bed D_{50} as a function of the excess unit stream power for each Pi (modified after Houbrechts et al., 2015). The class 22.6-32 mm of Kembs I1 and Kembs I2 was excluded to make possible comparisons with Interreg. 518 519 520 521 522 523

No relationship between the dimensionless travel distance of the tracer cloud centroid and the grain size classes of recovered tracers was found for Kembs I1 during P_1 , while a relation was observed during P_2 , indicating a partial mobility for this period [\(Figure 10\)](#page-34-0). For Kembs I2, a marked negative relationship is observed between the two variables during P_1 , also indicating selective sediment transport. For Interreg, no relationship was found between the two variables up to P_4 . Thereafter, a marked negative relationship was established during P_5 , again showing selective transport. 524 525 526 527 528 529 530

65

531

Figure 10: Mean dimensionless travel distance of the tracer cloud centroid for each grain size class, (a) for Kembs I1 and Kembs I2 and (b) for Interreg. Tracer clouds composed of less than 10 particles were excluded from the analysis because they were considered not statistically significant. For Kembs I1, only tracers recovered in the channel were included in the analysis. 532 533 534 535

3.2.2 Assessment of spatiotemporal grain size evolution 536

On a gravel bar sampled in the upstream part of the Interreg surveyed reach, the D_{95} decreased from S_0 to S_3 - S_4 [\(Figure 11](#page-35-0).a). The grain size fining reached 36 mm between S_0 and S_3 and 53 mm between S_0 and S_4 . New coarsening appeared in S5 with grain size values close to the S_0 values. For Kembs I2, a statistically significant bed fining (95% confidence interval) was observed (Kruskal-Wallis test, p-value < 0.0001) between S_0 and S_1 , with a decrease in the mean D_{50} of the ten largest particle samples from 110 to 84 mm (Figure [11](#page-35-0).b);. In S_0 , a downstream grain size coarsening (from KP 176.3 to KP 177.5) was observed and was probably related to the downstream increase in the flow velocities [\(Figure 11](#page-35-0).c). In S_1 , a statistically significant downstream bed fining (95% confidence level) was observed (R^2 = 0.26; p-value < 0.016). The bed fining may be related to the partial mobility identified by the 537 538 539 540 541 542 543 544 545 546

bedload tracking surveys on the Kembs I1 surveyed reach during P_2 , and the Kembs I2 547

surveyed reach during P_1 (see Section 3.2.1). 548

Figure 11: (a) D_{95} measured on each macroform and each survey along the Interreg restored reach with filled and empty symbols indicating the underwater and emerged samplings, respectively. (b) Boxplots of the D_{50} of the ten largest particles digitalized on each underwater digital image close to and downstream of the Kembs I2 SD for S_0 and S_1 and (c) D_{50} of the ten largest particles digitalized on each underwater digital image for S_0 and S_1 550 551 552 553 554

close to and downstream of the Kembs I2 SD; the gray line indicates the maximum velocity for the peak flow during P_1 for each grain size transect; the black line and black dashed line correspond to the linear regression lines of S_0 and S_1 , respectively. 3.2.3 Assessment of the spatiotemporal evolution of the channel diversity [Figure 12](#page-37-0) shows the temporal and longitudinal evolution of the cumulative differences in the cross-section diversity index (CSD) between 1985/93-2009 and the successive monitoring surveys along the three restored reaches. The results show a negative downstream trend in the cumulative differences of the post-GA index along all the reaches, which is an inverse tendency by comparison with the 1985/93-2009 dataset. Moreover, the rates of changes were higher after GA compared with those in the historical dataset. The results for the Interreg reach show that the negative trend intensity decreased over time due to the sediment dispersion along the channel bed. Marked positive trends of the index were first observed in S_5 [\(Figure 12.](#page-37-0)a), which seems to be correlated to the fragmentation of the bedload pulse ([Figure 8](#page-31-0).f). 555 556 557 558 559 560 561 562 563 564 565 566 567 568

569

Figure 12: Cumulative differences in the CSD between the initial state and the successive surveys in the downstream direction (a) for the period 1985/93 - 2009 and S_0 to S_5 for the Interreg restored reach and (b) for the period 1985/93 - 2009 and S_0 to S_2 for Kembs I1 and S_0 to S_1 for Kembs 12. 570 571 572 573

4. Discussion 574

4.1 Optimal stockpile deposit design 575

The topo-bathymetric monitoring showed that the erosion of the three SDs included two 576

processes closely interlocked: SD toe erosion and mass failure. This is consistent with 577

previous field and modeling studies on GA operations (Battisacco, 2016; Heckmann et al., 2017). Battisacco et al, (2016) and Vonwiller et al, (2018) showed that toe erosion is the predominant process, followed by the mass failure of the upper part of the SD. In our case, both processes occurred from a flow velocity of 2.5 m.s⁻¹ ([Figure 5.](#page-26-0)a), which caused full mobility of the introduced sediment [\(Figure 6\)](#page-27-0). For flow velocities lower than this critical value, only a limited toe erosion was observed with a selective mobility of gravels, which generated a reduction in the erosion intensity and the establishment of a coarser layer on the residual SD, as observed in the upstream and downstream part of Kembs I1 [\(Figure 3.](#page-23-0)b). This coarser layer can significantly reduce the erosion of the residual SD during future flood events, as observed in studies conducted on sediment starved rivers (Rollet et al. 2013; Houbrechts et al. 2012), and explain the lower eroded volumes during P_2 compared to P_1 , despite higher and longer flood events [\(Figure 2](#page-11-0).a). These results highlight the fact that SDs should be implemented/designed in suitable hydraulic areas to maximize the SD erosion probability for a $Q₁₋₂$ flood, considering that entrainment must be performed for frequent floods to avoid any risk of vegetation encroachment and increase in phytostabilization. Based on the critical flow velocities ([Figure](#page-26-0) [5](#page-26-0) and [Figure 6](#page-27-0)), we determined the suitable areas for GA along the first 15 km of the studied reach [\(Figure 7\)](#page-28-0). From this study, we estimate that the potentially suitable areas correspond to 53% of the study area. This approach is useful to plan further GAs in the Old Rhine, and it should be transferable to other rivers. Our results also show that SD implementation on bars is not suitable for a Q_2 peak flow due to flow velocities lower than the critical value. 578 579 580 581 582 583 584 585 586 587 588 589 590 591 592 593 594 595 596 597 598

4.2 Bedload transport patterns 599

At a short term (1 to 3 years), our results show that the mode of sediment transfer varies among a translation pulse for Interreg and a combination of translation and dispersion for Kembs I1 and Kembs I2 [\(Figure 8\)](#page-31-0). Such processes were also observed in both modeling (Koll et al., 2010; Sklar et al., 2009) and field studies on GA operations (Downs et al., 2016). The mode of sediment pulse depends on the grain size of the introduced sediment compared to the predisturbance in-channel grain size, sediment sorting and sediment volume (Lisle et al., 2001; Sklar et al., 2009; Sims et al., 2017). Dispersion occurs in high bed roughness contexts and when the input volumes exceed the bedload transport capacity (Lisle et al., 2001; Sims et al., 2017). At mid-term \sim 5 years), the mode of sediment pulse for the Interreg GA was non-random and fragmented (Gaeuman et al., 2017), driven by the bed morphology, particularly riffle-pool sequences (Sear, 1996) ([Figure 8\)](#page-31-0). Virtual velocities show the same order of magnitude as those in other studies conducted on riffle-pool rivers for the same exceeding peak flow energy, which indicates that the GA sediment transport dynamics were equivalent to the sediment mobility in natural conditions. A decrease in the virtual velocities occurred during the monitoring period despite the increase in the excess flow energy [\(Figure 9](#page-33-0).a). This decreasing tendency of sediment mobility over time was observed in other field studies (Ferguson et al., 2002; Haschenburger, 2011) and modeling works (Koll et al., 2010). This tendency is explained by a vertical mixing of gravels where tracers are progressively buried, which reduces their mobility (Ferguson et al. 2002). Sediment mobility can also be slowed by the bed roughness (Hassan et al., 1991) and the variability in the bed morphology (Pyrce and Ashmore, 2003; Sear, 1996). The relations between bedload mobility and hydrology are more complex because of numerous determinant parameters, such as peak flow intensity, hydrogram shape (duration and peak 600 601 602 603 604 605 606 607 608 609 610 611 612 613 614 615 616 617 618 619 620 621 622

77

number) and initial conditions, such as low flow duration preceding the flood event and sediment transport processes during the previous morphogenic flood (Humphries et al., 2012). Vázquez-Tarrío et al. (2018) showed on the Rhône River that a coarser armor layer is frequently located just below dams due to the propagation of sediment starvation conditions in the downstream direction. Rollet et al. (2013) observed on the Ain River that sediment tracers travel relatively slowly on sediment-starved reaches due to sediment trapping in the interstices of the armor layer. Thus, the differences in the sediment pulse patterns and the virtual velocities among the three GAs can be explained by both lower bedload transport capacities and higher bed roughness for Kembs I1 and I2 compared to Interreg. Overall, the results reinforce the conclusions of Arnaud et al. (2017) that no important sediment volumes travel rapidly downstream in the Old Rhine. Consequently, there is no risk of navigation perturbations downstream of the Breisach dam ([Figure 1\)](#page-9-0). 623 624 625 626 627 628 629 630 631 632 633 634

635

4.3 Channel morphodynamic responses

In the short term, a significant bed grain size fining was observed in the channel and on some emerged bars ([Figure 11](#page-35-0).a-b-c), while the topo-bathymetric results revealed a simplification of the channel bed geometry ([Figure 12\)](#page-37-0). These results are consistent with a modeling study conducted by Battisacco et al. (2016), which demonstrated that SDs having characteristics similar to ours produce sediment spreading along the restored reach, thus favoring bed grain size fining and a limitation in the bed incision tendency rather than increasing habitat heterogeneity. At mid-term, the grain size fining observed in the Interreg reach was temporary due to the downstream bedload transfer without any new supply. A relationship appeared between the dimensionless distances of the tracers and their grain size when the dispersion was advanced ([Figure 10](#page-34-0)), indicating a partial mobility (Church and 636 637 638 639 640 641 642 643 644 645

Hassan, 1992). This may reveal the formation of a new armor layer due to new sediment starvation conditions explained by the lack of upstream sediment supply (Piégay et al., 2016). All results show that the positive geomorphic effects of the three GAs are transitory, with a duration of approximately 3-4 flood events depending on the remobilized sediment volumes, the flow velocities and the bedload transport processes (Sklar et al., 2009). An increase in the channel diversity downstream of the Interreg SD occurred due to a preferential sediment deposition on riffles ([Figure 12](#page-37-0)) and explain notably by lower flow velocities and shear stresses than in pools during flood events (Sear, 1996). This result indicates that a diversified downstream channel morphology tends to enhance the positive morphological effects of GAs and should be taken into account, in addition to local flow conditions around the SDs, for designing GA operations. The sediment wave fragmentation along the Interreg reach has been highlighted because 646 647 648 649 650 651 652 653 654 655 656 657

the monitoring was conducted over a long period (7 years) and an important channel length (maximum of 12 km). Thus, monitoring must be performed on extended lengths (several kilometers) and over mid to long temporal scales to adequately evaluate the effects of GAs. We can hypothesize that a fragmentation of the sediment wave of both of the Kembs GAs will also occur more or less rapidly depending on flow conditions. 658 659 660 661 662

Our results showed that even if GA operations are repeated over time in the Old Rhine, it is unlikely that a sustainable diversification of aquatic habitats may be obtained due to the narrow and homogeneous channel geometry of the study reach. Channel widening by removal of bank protection, partial removal of groin fields and/or excavation of floodplain areas connected to the channel should be tested downstream of GAs so as to promote a reduction and a diversification of flow conditions, inducing the deposition of a part of the introduced sediments and potentially creating new fluvial forms, such as side channels and 663 664 665 666 667 668 669

bars colonizable by pioneer vegetation (Arnaud et al., 2017; Staentzel et al., 2018b).

Sediment excavated from channel widening should be stored temporarily and introduced 671

upstream with a rhythm depending on the hydrological regime. The success of this mode of 672

sediment management has been demonstrated along the Mur River by Klösch (2011), which 673

highlighted that channel widening locally decreased the bedload transport capacity of the 674

river, favoring bedload deposition and, consequently, local bank erosion as a positive 675

retroaction loop. However, the self-sustaining of geomorphic processes is strongly 676

dependent on the upstream bedload supply (Klösch, 2011). 677

In terms of risks, our results show that the evolution of channel topography at the mid-term 678

is relatively low ([Figure 12](#page-37-0)). Our results confirm the absence of risk of uncontrolled bed 679

degradation, which was also noticed by Arnaud et al. (2017) over a shorter monitoring 680

period for the Interreg GA. Furthermore, the hydrological functioning of the excavated areas 681

in terms of flow retention may not be impacted either. In addition, no significant bank 682

erosion was observed during the monitoring period. 683

4.4 Recommendations and management scenario to improve the efficiency and 684

sustainability of GAs 685

Following this study, several recommendations and management scenario can be addressed for improving the efficiency of GA operations and the sustainability of their effects at the scale of the SDs as well as the scale of the downstream reach: 686 687 688

1. perform hydraulic modeling to identify the most suitable areas for SD implementation; 689 690

2. on this basis, prioritize areas with high flow velocities during frequent floods to implement SDs to limit the risk of SD fixation accentuated by the development of the vegetation; 691 692 693

3. adapt the grain size of the GAs to the local flow competence to minimize the risk of the formation of a coarser layer at the top of the residual SD, which would decrease the SD erodibility; 694 695 696

4. implement SDs upstream of reaches showing the highest diversity of macroforms and flow conditions to favor a fragmentation of the sediment transport and thus 697 698

maximize potential habitat diversification, notably by bed grain size fining. 699

5. Test channel widening downstream of GAs to lower flow conditions, and to induce sediment deposition (and a general slowdown of the sediment waves), the 700 701

establishment of pioneer vegetation and the diversification of aquatic and riverine 702

habitats (Arnaud et al. 2017), as bars and secondary channels (Figure 13.a). Sediment 703

deposition may also induce local bar/bank erosion, provoking local bedload supply 704

into the channel in a positive feedback loop. Widening areas should be performed 705

ideally close to riffles to reinforce the sediment wave fragmentation by gravel 706

deposition, extending spawning areas for salmonid fish species. Riffle aggradation 707

and bar formation may also foster hyporheic water exfiltration areas, which are 708

summer thermal refuges favoring the adaptation of aquatic ecosystems to climate 709

change (Eschbach et al., 2017). Furthermore, habitat diversification may also limit the 710

upstream iteratively ([Figure 13](#page-44-0).a.2). This management scenario would increase both

relative development of invasive species (Staentzel et al., 2018a) and raise the 711

recreative value of the Old Rhine (Arnaud et al., 2017). The resulting excavated 712

sediments should be stored temporarily close to the channel and introduced 713

714

the extent and the durability of the GA benefits, in comparison with the management 715

717

Figure 13: (a) Restoration scenarios based on 1-GA only, 2-combination of GA with downstream channel widening, (b) benefits of both scenarios in terms of spatial extent and duration. 718 719 720

5. Conclusions 721

This study provides original feedback based on multi-compartment monitoring of three *in situ* GAs conducted along the bypassed section of the Rhine River downstream of Kembs, notably in terms of SD erodibility, sediment transport dynamics and channel responses. The erosion of SDs is dependent on local flow velocities and also on the grain size of the introduced sediment according to the local transport capacity. The Interreg GA was the best GA design compared to the two others. More generally, 53% of the 15 km upstream of the Old Rhine are suitable to implement GAs for flood flows or higher Q_2 . Along the reaches affected by the GAs in the short term, bathymetric simplification was observed due to the transfer of sediments mostly along the thalweg and a transitory grain size fining. At mid-722 723 724 725 726 727 728 729 730

term (~ 5 years), local habitat diversification was observed due to fragmentation of the sediment wave with preferential bedload deposition on riffles, which is a positive effect for fish spawning habitats. Our results also show that GAs induce no risk for human activities. This study highlights that to improve restoration efficiency and sustainability, GAs must be implemented in upstream reaches characterized by the highest hydromorphological diversity to increase positive effects. However, new sediment starvation conditions appeared > 5 years due to the lack of upstream sediment supply, showing that GAs must be repeated over time. To increase the efficiency of GA operations, channel widening should be performed downstream of the GAs to improve the sustainability of the GA effects. Sediment excavated from channel widenings should be stored temporarily and introduced upstream iteratively. **References** Arnaud F. 2012. Historical and Experimental Geomorphology for the Process-based Restoration of a Modified Fluvial System: the Old Rhine Between Kembs and Breisach (France, Germany), PhD Dissertation. University of Lyon (in French with English abstract). Arnaud, F., Piégay, H., Béal, D., Collery, P., Vaudor, L., Rollet, A.-J., 2017. Monitoring gravel augmentation in a large regulated river and implications for process-based restoration: Monitoring gravel augmentation in a large regulated river. Earth Surf. Process. Landforms 42, 2147–2166. https://doi.org/10.1002/esp.4161 Arnaud, F., Piégay, H., Schmitt, L., Rollet, A.J., Ferrier, V., Béal, D., 2015a. Historical geomorphic analysis (1932 – 2011) of a by-passed river reach in process-based restoration perspectives : The Old Rhine downstream of the Kembs diversion dam (France , Germany). Geomorphology 236, 163–177. 731 732 733 734 735 736 737 738 739 740 741 742 743 744 745 746 747 748 749 750 751 752 753 754

89

- https://doi.org/10.1016/j.geomorph.2015.02.009 755
- Arnaud, F., Piégay, H., Vaudor, L., Bultingaire, L., Fantino, G., 2015b. Technical specifications 756
- of low-frequency radio identification bedload tracking from field experiments: 757
- Differences in antennas, tags and operators. Geomorphology 238, 37–46. 758
- https://doi.org/10.1016/j.geomorph.2015.02.029 759
- Arnaud, F., Schmitt, L., Johnstone, K., Rollet, A., Piégay, H., 2019. Geomorphology 760
- Engineering impacts on the Upper Rhine channel and floodplain over two centuries. 761

Geomorphology 330, 13–27. https://doi.org/10.1016/j.geomorph.2019.01.004 762

- Battisacco, E., Franca, M.J., Schleiss, A.J., 2016. Sediment replenishment: Influence of the 763
- geometrical configuration on the morphological evolution of channel-bed. Water 764
- Resour. Res. 52. https://doi.org/10.1002/2016WR019157 765
- Bennett, G.L., Molnar, P., Eisenbeiss, H., Mcardell, B.W., 2012. Erosional power in the Swiss 766
- Alps: Characterization of slope failure in the Illgraben. Earth Surf. Process. Landforms 767
- 37, 1627–1640. https://doi.org/10.1002/esp.3263 768
- Béraud C. 2012. Numerical Modelling of Sediment Feeding Impacts in Engineered Rivers. The 769
- Case of the Old Rhine Between Kembs and Breisach, PhD Dissertation. University of 770
- Lyon (in French with English abstract). 771
- Brasington, J., Langham, J., Rumsby, B., 2003. Methodological sensitivity of morphometric 772
- estimates of coarse fluvial sediment transport. Geomorphology 53, 299–316. 773
- https://doi.org/10.1016/S0169-555X(02)00320-3 774
- Brousse, G., Arnaud-Fassetta, G., Liébault, F., Bertrand, M., Melun, G., Loire, R., Malavoi, J.R., 775
- Fantino, G., Borgniet, L., 2019. Channel response to sediment replenishment in a large 776

gravel-bed river: The case of the Saint-Sauveur dam in the Buëch River (Southern Alps, France), in: River Research and Applications. 1-14. <https://doi.org/10.1002/rra.3527> Chardon, V., Schmitt, L., Piégay, H., Lague, D., 2020. Use of terrestrial photosieving and airborne topographic LiDAR to assess bed grain size in large rivers: a study on the Rhine River. Earth Surf. Process. Landforms. https://doi.org/10.1002/esp.4882 CHR. 2009. Erosion, Transport and Deposition of Sediment – Case Study Rhine, Report II-20. CHR/KHR: Lelystad. Church, M., Mclean, D., Wolcott J.F, 1987. River Bed Gravels: Sampling and Analysis. Church, M., Hassan, M.A., 1992. Size and Distrance of Travel Unsconstrained Clasts on a Streambed. Water Resour. Res. 28, 299–303. https://doi.org/10.1029/91WR02523 Cooper, A.R., Infante, D.M., Daniel, W.M., Wehrly, K.E., Wang, L., Brenden, T.O., 2017. Assessment of dam effects on streams and fish assemblages of the conterminous USA. Sci. Total Environ. 586, 879–889. https://doi.org/10.1016/j.scitotenv.2017.02.067 Douglas L. Inman, 1952. Measures for Describing the Size Distribution of Sediments. SEPM J. Sediment. Res. Vol. 22, 125–145. https://doi.org/10.1306/D42694DB-2B26-11D7- 8648000102C1865D Downs, P.W., Keele, V.E., Gilvear, D.J., 2016. Dispersal of augmented gravel in a steep , boulder-bedded reach : early implications for restoring salmonid habitat. Proceedings 11th International Symposium on Ecohydraulics (ISE 2016). Dittrich A., Koll K., Kunz C., 2010. Historische Entwicklung des südlichen Oberrheins unter morphologischen Gesichtspunkten. Bericht Nr. 985. Leichtweiss-Institut für Wasserbau, Technische Universität Braunschweig, 38 p El kadi Abderrezzak K. 2009. Estimate of Bedload Transport Capacity in the Old Rhine 777 778 779 780 781 782 783 784 785 786 787 788 789 790 791 792 793 794 795 796 797 798 799

94

- Between Kembs and Breisach, Report H-P73–2009- 00402-FR. EDF R&D: Chatou (in 800
- French with English abstract). 801
- Eschbach, D., Piasny, G., Schmitt, L., Pfister, L., Grussenmeyer, P., Koehl, M., Skupinski, G., 802
- Serradj, A., 2017. Thermal-infrared remote sensing of surface water–groundwater 803
- exchanges in a restored anastomosing channel (Upper Rhine River, France). Hydrol. 804
- Process. 31, 1113–1124. https://doi.org/10.1002/hyp.11100 805
- Ferguson, R.I., Bloomer, D.J., Hoey, T.B., Werritty, A., 2002. Mobility of river tracer pebbles 806
- over different timescales. Water Resour. Res. 38, 3–8. 807
- https://doi.org/10.1029/2001wr000254 808
- Foley, M.M., Magilligan, F.J., Torgersen, C.E., Major, J.J., Anderson, C.W., Connolly, P.J., 809
- Wieferich, D., Shafroth, P.B., Evans, J.E., Infante, D., Craig, L.S., 2017. Dam removal-810
- listening in Landscape context and the biophysical response of rivers to dam removal in 811
- the United States.PLOS ONE. https://doi.org/10.1371/journal.pone.0180107 812
- Gaeuman D. 2013. Monitoring the Effectiveness of Gravel Augmentations for Salmonid 813
- Habitat Improvement Downstream from Dams, Completion Report Project 0295. US 814
- Bureau of Reclamation Science and Technology: Washington, DC. 815
- Gaeuman D. 2014. High-flow gravel injection for constructing designed in-channel features. 816
- River Research and Applications 30, 685–706. https://doi.org/10.1002/rra.2662 817
- Gaeuman, D., Stewart, R., Schmandt, B., Pryor, C., 2017. Geomorphic response to gravel 818
- augmentation and high flow dam release in the Trinity River, California 2540, 2523-819
- 2540. https://doi.org/10.1002/esp.4191 820
- Garnier, A., Barillier, A., 2015. The Kembs project: environmental integration of a large 821
- existing hydropower scheme. La Houille Blanche 21–28. 822
- https://doi.org/10.1051/lhb/20150041 823

- Gintz, D., Hassan, M.A., Schmidt, K.H., 1996. Frequency and magnitude of bedload transport 824
- in a mountain river. Earth Surf. Process. Landforms 21, 433–445. 825
- https://doi.org/10.1002/(SICI)1096-9837(199605)21:5<433::AID-ESP580>3.0.CO;2-P 826
- Gostner, W., Alp, M., Schleiss, A.J., Robinson, C.T., 2013. The hydro-morphological index of 827
- diversity: a tool for describing habitat heterogeneity in river engineering projects. 828
- Hydrobiologia 712, 43–60. https://doi.org/10.1007/s10750-012-1288-5 829
- Haschenburger, J.K., 2011. The rate of fluvial gravel dispersion. Geophys. Res. Lett. 38 830
- https://doi.org/10.1029/2011GL049928 831
- Haschenburger, J.K., Church, M., 1998. Bed material transport estimated from the virtual 832
- velocity of sediment. Earth Surf. Process. Landforms. 23, 791-808. 833
- https://doi.org/10.1002/(SICI)1096-9837(199809)23:9<791::AID-ESP888>3.0.CO;2-X 834
- Hassan, M.A., Church, M., Ashworth, P.J., 1992. Virtual rate and mean distance of travel of 835
- individual clasts in gravel-bed channels. Earth Surf. Process. Landforms 17, 617–627. 836
- https://ui.adsabs.harvard.edu/link_gateway/1992ESPL...17..617H/doi:10.1002/ 837
- esp.3290170607 838
- Hassan, M.A., Church, M., Schick, A.P., 1991. Distance of movement of coarse particles in 839
- gravel bed streams 27, 503-511. https://doi.org/10.1029/90WR02762 840
- Hassan, M.A., Voepel, H., Schumer, R., Parker, G., Fraccarollo, L., 2013. Displacement 841
- characteristics of coarse fluvial bed sediment. JOURNAL OF GEOPHYSICAL 842
- RESEARCH: EARTH SURFACE, Vol.118, 155–165. 843
- https://doi.org/10.1029/2012JF002374 844
- Heckmann, T., Haas, F., Abel, J., Rimböck, A., Becht, M., 2017. Geomorphology Feeding the 845
- hungry river : Fluvial morphodynamics and the entrainment of arti fi cially inserted 846
- sediment at the dammed river Isar , Eastern 291, 128–142. 847
- https://doi.org/10.1016/j.geomorph.2017.01.025 848
- Houbrechts, G., Levecq, Y., Vanderheyden, V., Petit, F., 2011. Long-term bedload mobility in 849
- gravel-bed rivers using iron slag as a tracer. Geomorphology 126, 233-244. 850
- https://doi.org/10.1016/j.geomorph.2010.11.006 851
- Houbrechts, G., Levecq, Y., Peeters, A., Hallot, E., Campenhout, J. Van, Denis, A., Petit, F., 852
- 2015. Geomorphology Evaluation of long-term bedload virtual velocity in gravel-bed 853
- rivers. Geomorphology 251, 6-19. https://doi.org/10.1016/j.geomorph.2015.05.012 854
- Houbrechts, G., Van Campenhout, J., Levecq, Y., Hallot, E., Peeters, A., Petit, F., 2012. 855
- Comparison of methods for quantifying active layer dynamics and bedload discharge in 856
- armoured gravel-bed rivers. Earth Surf. Process. Landforms 37, 1501–1517. 857
- https://doi.org/10.1002/esp.3258 858
- Humphries, R., Venditti, J.G., Sklar, L.S., Wooster, J.K., 2012. Experimental evidence for the 859
- effect of hydrographs on sediment pulse dynamics in gravel-bedded rivers. Water 860
- Resour. Res. 48 https://doi.org/10.1029/2011WR010419 861
- Juez, C., Battisacco, E., Schleiss, A.J., Franca, M.J., 2016. Advances in Water Resources 862
- Assessment of the performance of numerical modeling in reproducing a replenishment 863
- of sediments in a water-worked channel. Adv. Water Resour. 92, 10–22. 864
- https://doi.org/10.1016/j.advwatres.2016.03.010 865
- Klösch, M., Hornich, R., Baumann, N., Puchner, G., Habersack, H., 2011. Mitigating channel 866
- incision via sediment input and self-initiated riverbank erosion at the Mur River, Austria. 867
- Geophys. Monogr. Ser. https://doi.org/10.1029/2010GM000977 868
- Koll K, Koll K, Dittrich A. 2010. Sediment transport over static armour layers and its impact on 869
- bed stability. In Proceedings, 5th International Conference on Fluvial Hydraulics (River 870
- Flow 2010), Dittrich A, Koll K, Aberle J, Geisenhainer P (eds). Bundesanstalt für 871
- Wasserbau: Braunschweig; 929–936. 872
- Mathias Kondolf, G., Graham, W. V, Matthews, G., 1991. UC Berkeley. 873
- Kondolf, G.M., 1997. Hungry water: Effects of dams and gravel mining on river channels. 874
- Environ. Manage. 21, 533-551 <https://doi.org/10.1007/s002679900048> 875
- Kondolf, G.M., Gao, Y., Annandale, G.W., Morris, G.L., Jiang, E., Zhang, J., Cao, Y., Carling, P., 876
- Fu, K., Guo, Q., Hotchkiss, R., Peteuil, C., Sumi, T., Wang, H.-W., Wang, Z., Wei, Z., Wu, 877
- B., Wu, C., Yang, C.T., 2014. Sustainable sediment management in reservoirs and 878
- regulated rivers: Experiences from five continents. Earth's Futur. 2, 256–280. 879
- https://doi.org/10.1002/2013EF000184 880
- Lamarre, H., Roy, A.G., 2008. The role of morphology on the displacement of particles in a 881
- step–pool river system. Geomorphology 99, 270–279. 882
- https://doi.org/10.1016/j.geomorph.2007.11.005 883
- Landon, N., 2008. Du constat d'enfoncement du lit fluvial aux actions de recharge 884
- sédimentaire: quellessolutions pour une gestion raisonnée de nos cours d'eau? Outils 885
- de gestion de l'eau en territoire demontagne, PNR du Queyras, ONEMA et MEDR, 886
- pp.28-39, 2007. halshs-00279895 887
- Lane, S.N., Westaway, R.M., Hicks, D.M., 2003. Estimation of erosion and deposition 888
- volumes in a large, gravel-bed, braided river using synoptic remote sensing. Earth Surf. 889
- Process. Landforms 28, 249–271.<https://doi.org/10.1002/esp.483> 890
- Lenzi, M.A., 2004. Displacement and transport of marked pebbles , cobbles and boulders 891
- during floods in a steep mountain stream, Hydrol. Process. 18, 1899–1914 (2004). 892
- https://doi.org/10.1002/hyp.1456 893

- Liébault, F., Bellot, H., Chapuis, M., Klotz, S., Deschâtres, M., 2012. Bedload tracing in a high-894
- sediment-load mountain stream. Earth Surf. Process. Landforms 37, 385–399. 895
- https://doi.org/10.1002/esp.2245 896
- Magilligan, F.J., Nislow, K.H., 2005. Changes in hydrologic regime by dams. Geomorphology 897
- 71, 61-78. https://doi.org/10.1016/j.geomorph.2004.08.017 898
- Ock, G., Gaeuman, D., McSloy, J., Kondolf, G.M., 2015. Ecological functions of restored 899
- gravel bars, the Trinity River, California. Ecol. Eng. 83, 49–60. 900
- https://doi.org/10.1016/j.ecoleng.2015.06.005 901
- Ock, G., Sumi, T., Takemon, Y., 2013. Sediment replenishment to downstream reaches below 902
- dams: implementation perspectives. Hydrol. Res. Lett. 7, 54–59. 903
- https://doi.org/10.3178/hrl.7.54 904
- Pasternack, G.B., Wang, C.L., Merz, J.E., 2004. Application of a 2D hydrodynamic model to 905
- design of reach-scale spawning gravel replenishment on the Mokelumne River, 906
- California. River Res. Appl. 20, 205–225. https://doi.org/10.1002/rra.748 907
- Petts, G.E., Gurnell, A.M., 2005. Dams and geomorphology: Research progress and future 908
- directions. Geomorphology 71, 27–47. 909
- https://doi.org/10.1016/j.geomorph.2004.02.015 910
- Piégay, H., Aelbrecht, D., Béal, D., Alonso, C., Armburster, J., Arnaud, F., Béraud, C., Billard, 911
- C., Bouchard, J.P., Piégay, H., Aelbrecht, D., Béal, D., Alonso, C., Armburster, J., 2011. 912
- Restauration morpho-dynamique et redynamisation de la section court-circuitée du 913
- Rhin en aval du barrage de Kembs (projet INTERREG / EDF). 914
- Piégay, H., Arnaud, F., Cassel, M., Depret, T., Alber, A., Michel, K., Rollet, A.-J., Vaudor, L., 915
- 2016. Suivi par RFID de la mobilité des galets: retour sur 10 ans d'expérience en grandes 916
- rivières. BSGLg 67. 917
- Pyrce, R.S., Ashmore, P.E., 2003. The relation between particle path length distributions and 918
- channel morphology in gravel-bed streams : a synthesis. Geomorphology 56, 167–187. 919

https://doi.org/10.1016/S0169-555X(03)00077-1 920

- Rempel, L.L., Church, M., 2009. Physical and ecological response to disturbance by gravel 921
- mining in a large alluvial river. Can. J. Fish. Aquat. Sci. 66, 52–71. 922
- https://doi.org/10.1139/F08-184 923
- Rollet, A.j., 2007. Study and management of the sedimentary dynamics of a river section 924
- downstream of a dam: the case of the lower Ain Valley. PhD Dissertation. University of 925
- Lyon (in French with English abstract). 926
- Rollet, A.J., Dufour, S., Bornette, G., Persat, H., 2013. Assessment of consequences of 927
- sediment deficit on a gravel river bed downstream of dams in restoration perspectives : 928
- application of a multicriteria , hierarchical and spatially explicit diagnosis. River Res. 929
- Appl 30, 939-953. https://doi.org/10.1002/rra 930
- Schälchli, U., Breitenstein, M., Kirchhofer, A., 2010. Kiesschüttungen zur Reaktivierung des 931
- Geschiebehaushalts der Aare-die kieslaichenden Fische freut's. Wasser Energ. Luft 102, 932
- 209–213. 933
- Schmitt, L., Morris, D., Kondolf, G.M., 2018. Managing floods in large river basins in Europe: 934
- The Rhine river, in: Managing Flood Risk: Innovative Approaches from Big Floodplain 935
- Rivers and Urban Streams. https://doi.org/10.1007/978-3-319-71673-2_4 936
- Schneider, J.M., Turowski, J.M., Rickenmann, D., Hegglin, R., Arrigo, S., Mao, L., Kirchner, 937
- J.W., 1Mountain, 2014. Journal of Geophysical Research : Earth Surface. J. Geophys. 938
- Res. Earth Surf. 119, 2315–2334. https://doi.org/10.1002/2014JF003147.Received 939
- Sear, D.A., 1996. Sediment transport processes in pool-riffle sequences. Earth Surf. Process. 940

Landforms 21, 241-262. https://doi.org/10.1002/(SICI)1096-941

- 9837(199603)21:3<241::AID-ESP623>3.0.CO;2-1 942
- Sims, A., Sims, A.J., Rutherfurd, I.D., 2017. Management responses to pulses of bedload 943
- sediment in rivers Geomorphology Management responses to pulses of bedload 944
- sediment in rivers. Geomorphology 294, 70–86. 945
- https://doi.org/10.1016/j.geomorph.2017.04.010 946
- Sklar, L.S., Fadde, J., Venditti, J.G., Nelson, P., Wydzga, M.A., Cui, Y., Dietrich, W.E., 2009. 947
- Translation and dispersion of sediment pulses in flume experiments simulating gravel 948
- augmentation below dams 45, 1–14. https://doi.org/10.1029/2008WR007346 949
- Staentzel, C., Arnaud, F., Combroux, I., Schmitt, L., Trémolières, M., Grac, C., Piégay, H., 950
- Barillier, A., Chardon, V., Beisel, J.-N., 2018a. How do instream flow increase and gravel 951
- augmentation impact biological communities in large rivers: A case study on the Upper 952
- Rhine River. River Res. Appl. 34, 153-164 https://doi.org/10.1002/rra.3237 953
- Staentzel, Cybill, Combroux, I., Barillier, A., Schmitt, L., Chardon, V., Garnier, A., Beisel, J.N., 954
- 2018b. Réponses des communautés biologiques à des actions de restauration de grands 955
- fleuves, La Houille Blanche 1–9. https://doi.org/10.1051/lhb/2018024 956
- Stähly, S., Franca, M.J., Robinson, C.T., Schleiss, A.J., 2019. Sediment replenishment 957
- combined with an artificial flood improves river habitats downstream of a dam. Sci. 958
- Report 9, 5176. https://doi.org/10.1038/s41598-019-41575-6 959
- Sumi, Tetsuya Masahisa Okano, Y.T., 2004. Proceedings of the Ninth International 960
- Symposium on River Sedimentation October 18 21, 2004, Yichang, China 1036–1043. 961
- Vázquez-Tarrío, D., Tal, M., Camenen, B., Piégay, H., 2018. Effects of continuous 962
- embankments and successive run-of-the-river dams on bedload transport capacities 963
- along the Rhône River, France. Sci. Total Environ. 658, 1375-1389. 964

- https://doi.org/10.1016/j.scitotenv.2018.12.109
- Vonwiller, L., Vetsch, D.F., Boes, R.M., 2018. Modeling streambank and artificial gravel
- deposit erosion for sediment replenishment. Water 10, 1–23.
- https://doi.org/10.3390/w10040508
- Wheaton, J.M., Pasternack, G.B., Merz, J.E., 2004. Spawning habitat rehabilitation-I.
- Conceptual approach and methods. Int. J. River Basin Manag. 2, 3–20.
- Wolman, M.G., 1954. A method of sampling coarse river-bed material. Eos, Trans. Am.
- Geophys. Union. https://doi.org/10.1029/TR035i006p00951
- Zeug, S.C., Sellheim, K., Watry, C., Rook, B., Hannon, J., Zimmerman, J., Cox, D., Merz, J.,
- 2013. Gravel augmentation increases spawning utilization by anadromous salmonids : a
- case study from california , 718, 707–718. <https://doi.org/10.1002/rra>
-
-
-
-
-
-
-
-
-
-
-
-
-

Acknowledgements 989

This research was supported by the European Regional Development Fund within the project 'Redynamization of the Old Rhine – 2009–2012' (FEDER INTERREG IV Upper Rhine; coord. Région Alsace/CNRS UMR 5600 EVS) and by the company Electricité de France (EDF) within the research collaborations 'Sediment reintroduction into the Old Rhine from controlled bank erosion – 2009–2012' (EDF 5500-AAH-5910047888), 'Geomorphic monitoring of gravel augmentation experiment into the Old Rhine – years 2013-2014' (EDF 5500-4300937910), 'Redynamization of the Old Rhine – years 2014-2018' (EDF 5910132058) and by the Région Grand Est (328/14/C1 2016-2017). This work was performed within the framework of the EUR H2O'Lyon (ANR-17-EURE-0018) of the University of Lyon within the program "Investissements d'Avenir" operated by the French National Research Agency (ANR). The authors also warmly thank colleagues who provided fieldwork assistance: J. Houssier, J. Serouilou and G. Skupinski. The authors also want to thank two anonymous reviewers who greatly contributed to improve to this paper by providing helpful reviews of an earlier version of this manuscript. 990 991 992 993 994 995 996 997 998 999 1000 1001 1002 1003 1004