

# The diversity of biotic interactions complements functional and phylogenetic facets of biodiversity

Pierre Gaüzère, Louise O'connor, Christophe Botella, Giovanni Poggiato, Tamara Münkemüller, Laura Pollock, Ulrich Brose, Luigi Maiorano, Michael Harfoot, Wilfried Thuiller

## ▶ To cite this version:

Pierre Gaüzère, Louise O'connor, Christophe Botella, Giovanni Poggiato, Tamara Münkemüller, et al.. The diversity of biotic interactions complements functional and phylogenetic facets of biodiversity. Current Biology - CB, 2022, 32 (9), pp.2093-2100.e3. 10.1016/j.cub.2022.03.009. hal-03750719

# HAL Id: hal-03750719 https://cnrs.hal.science/hal-03750719v1

Submitted on 16 Aug2022

**HAL** is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers. L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.

# 1 Manuscript

2

# The diversity of biotic interactions complements functional and phylogenetic facets of biodiversity

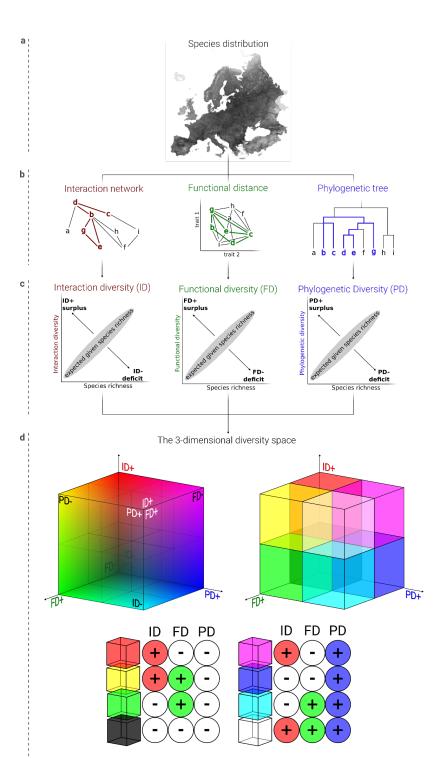
- Pierre Gaüzère<sup>1</sup>, Louise O'Connor<sup>1</sup>, Christophe Botella<sup>1</sup>, Giovanni Poggiato<sup>1</sup>, Tamara Münkemüller<sup>1</sup>, Laura J.
   Pollock<sup>2</sup>, Ulrich Brose<sup>3,4</sup>, Luigi Maiorano<sup>5</sup>, Michael Harfoot<sup>\*</sup>, Wilfried Thuiller<sup>1</sup>
- 8 <sup>1</sup> Univ. Grenoble Alpes, Univ. Savoie Mont Blanc, CNRS, LECA, F-38000 Grenoble
- 9 <sup>2</sup> Biology Department, McGill University, Montréal, Québec, Canada H3A 1B1
- 10 <sup>3</sup> Institute of Biodiversity, Friedrich Schiller University Jena, Jena, Germany
- 11 <sup>4</sup> German Centre for Integrative Biodiversity Research (iDiv) Halle-Jena-Leipzig, Leipzig, Germany
- 12 <sup>5</sup> Department of Biology and Biotechnologies "Charles Darwin", Sapienza University of Rome, Italy
- 13 <sup>x</sup> United Nations Environment Programme World Conservation Monitoring Centre (UNEP-WCMC), Cambridge, UK
- 14
- 15 Keywords

#### 16 biodiversity, vertebrates, macroecology, biogeography, conservation

## 17 Summary

Taxonomic, functional and phylogenetic diversities are important facets of biodiversity. Studying them 18 together has improved our understanding of community dynamics, ecosystem functioning and conservation 19 values <sup>1–3</sup>. In contrast to species, traits, and phylogenies, the diversity of biotic interactions has so far been 20 21 largely ignored as a biodiversity facet in large-scale studies. This neglect represents a crucial shortfall because biotic interactions shape community dynamics, drive important aspects of ecosystem functioning 4-22 23 <sup>7</sup>, provide services to humans, and have intrinsic conservation value <sup>8,9</sup>. Hence the diversity of interactions 24 can provide crucial and unique information with respect to other diversity facets. Here, we leveraged large 25 datasets of trophic interactions, functional traits, phylogenies and spatial distributions of >1000 terrestrial vertebrate species across Europe at a 10km resolution. We computed the diversity of interactions 26 27 (Interaction Diversity, ID) in addition to functional (FD) and phylogenetic diversities (PD). After controlling 28 for species richness, surplus and deficits of ID were neither correlated with FD nor with PD, thus 29 representing unique and complementary information to the commonly studied facets of diversity. A three-30 dimensional mapping allowed for simultaneously visualizing different combinations of ID-FD-PD. 31 Interestingly, the spatial distribution of these diversity combinations closely matched the boundaries 32 between ten European biogeographic regions, and revealed new, interaction-rich areas in the European 33 Boreal region and interaction-poor areas in central Europe. Our study demonstrates that the diversity of 34 interactions adds new and ecologically relevant information to multi-facetted, large-scale diversity studies 35 with implications for understanding eco-evolutionary processes and informing conservation planning.

- 36 37
- 38
- 39



41

Results and discussion

Figure 1. Conceptual workflow for a joint analysis of phylogenetic, functional, and interaction diversity. (a) Occurrences and probability of presence for 1149 terrestrial vertebrate species on 117,000 10×10km cells across Europe are combined with (b) the phylogenetic tree, a set of functional traits, and the trophic interactions of species. (c) We combined species distribution with phylogenetic, functional and trophic species attributes to compute local terrestrial vertebrate diversities using Hills numbers (q = 0, i.e "richness") and statistically corrected the diversity values by the local species richness. Note that the expected relationships (gray ellipses) are not necessarily linear. (d) We projected the diversity values in a 3dimensional space with each axis representing a diversity facet and a color in the Red-Blue-Green space (x = PD / blue, y = FD / green, z = ID / associated with deficits in FD and PD (ID>0, FD<0, PD<0); Yellow identifies surpluses of ID and FD associated with deficits in PD (ID>0, FD>0, PD<0); Green identifies deficits in ID and PD associated with FD surpluses (ID<0, FD>0, PD<0); Black identifies deficits in ID, PD and FD, Pink identifies surpluses of ID and FD associated with ID deficits; white identifies surpluses in ID, PD and FD deficits; Light blue identifies surpluses of PD and FD associated with ID deficits; white identifies surpluses in ID, PD and FD.

43

44 45

46 47

48 49

50

51

52

55 Biodiversity -the diversity of life on Earth- was originally used to refer to species diversity, but it is now used to reflect a multi-faceted concept <sup>3</sup>. Given the evidence that species diversity alone cannot appropriately 56 57 describe community assembly, ecosystem functioning and variation in community composition <sup>10</sup>, several complementary measures of biodiversity have emerged in the last three decades <sup>1,2</sup>. The most important 58 are the diversity of species' evolutionary histories (i.e phylogenetic diversity, PD) and their ecological 59 functions (i.e functional diversity, FD), but while PD and FD are becoming central to many studies <sup>11,12</sup>, the 60 diversity of biotic interactions (i.e interaction diversity, ID) has been poorly considered as a biodiversity 61 facet in large-scale studies (but see refs<sup>8,13</sup>). This is a major gap since biological interactions are tightly 62 63 linked to species coexistence <sup>14</sup>, ecosystem productivity and functioning <sup>4–7</sup>.

64 In its simplest form, ID is the total number of interactions shared by all species of a given assemblage<sup>9</sup>. 65 Interactions considered can be of different types and nature, e.g antagonistic (competition for resources), 66 mutualistic (pollination <sup>15,16</sup>), or trophic (predation <sup>17</sup>). Although the concept of interaction diversity is not novel <sup>8,13</sup> and has its own methodological tools <sup>18,19</sup>, the lack of information available on biotic interactions 67 <sup>20,21</sup> has limited its study across large taxonomical and spatial scales <sup>22-26</sup>. Here, we leveraged unique and 68 69 valuable data combining spatial distributions<sup>27</sup> (Figure 1.a), trophic interactions (Figure 1.b), functional traits<sup>28</sup>, and phylogenies<sup>29</sup> of most terrestrial vertebrate species in Europe <sup>30</sup> at a 10 km resolution. Within 70 71 each 10km cell, we computed interaction diversity (ID, as the number of trophic interactions), functional 72 diversity (FD, as the sum of functional pairwise Gower distances between species in the cell), and 73 phylogenetic diversity (PD, as the sum of the branch lengths of the phylogenetic tree containing all species present in the cell <sup>1</sup>) using Hill numbers <sup>31 18</sup>. We statistically corrected each diversity by the local species 74 richness <sup>32,33</sup> in order to measure and map deficits and surpluses of ID, FD, and PD (Figure 1.c). We also 75 76 investigated the correlation and complementarity between the three facets, and created a 3-dimensional 77 diversity space that reveals different local combinations of ID-FD-PD (Figure 1.d) and their distribution 78 across biogeographical regions in Europe.

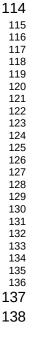
79

#### 80 Surpluses and deficits of diversities

Trophic networks of terrestrial vertebrates found within 10km cells in Europe contained up to 4834 trophic 81 82 interactions with an average of 1958 interactions across cells (Figure 2a). Once corrected for species richness, ID ranged from a deficit of -942 interactions (1667 observed interactions with 202 species involved) 83 to a surplus of +968 interactions (3730 interactions with 210 species involved, see Supplemental 84 85 information - Trophic network examples). Because highly connected assemblages are often considered as the signature of functional and resilient ecosystems <sup>34,35</sup>, areas with high ID are important from a 86 conservation point of view <sup>34,36</sup>. Further, comparing spatial distributions of surplus and deficit IDs with those 87 88 of FD or PD can complement our understanding of community dynamics and underlying processes. 89 Because phylogenetic and trait data contain information about evolutionary history and species niches, the 90 spatial distribution of their diversity (Figure 2b-c) is thought to hold the signature of the eco-evolutionary drivers that shape biodiversity patterns <sup>37–39</sup>. For example, for a given species richness, observed surplus of 91 92 FD (Figure 1d, and Figure 2b, green color) could result from competitive exclusion between species with 93 similar traits, while a deficit of FD might result from environmental filtering constraining the range of locally 94 viable traits or hierarchical competition where a given set of traits is the best adapted locally <sup>40</sup>. PD surplus (Figure 1d and Figure 2c, dark blue color) could result from slow extinction rates of old and distant lineages 95 (i.e., museums of biodiversity <sup>41</sup>), and PD deficit from rapid recent speciation (i.e., cradles of biodiversity). 96 97 ID surplus and deficit brings additional information, as observed ID surplus (Figure 1d and Figure 2a, red color) indicates particularly dense or long trophic networks, such as those emerging from high levels of 98 omnivory and intraguild predation <sup>42</sup>, or from bottom-up control when large amounts of basal resources 99 100 sustain longer trophic chains and the presence of top predators. ID deficits can result from weakened top-101 down control when top predators are absent from local assemblages, for example following human-induced removal 43-45. 102

103 Overall, the different facets of diversity are shaped by eco-evolutionary drivers which are not mutually 104 exclusive <sup>46</sup>. Any combination of ID-FD-PD could potentially exist locally and bring complementary 105 information to the others, although one can expect the facets of diversity to be (partly) correlated when 106 similar drivers influence multiple diversity facets. We showed that FD and PD were clearly and positively

correlated (Figure 3b). This correlation is due to the fact that species tend to retain their ancestral traits 108 through evolution <sup>38,47–49</sup>, and suggests an important effect of evolution and phylogenetic niche conservatism 109 on biodiversity patterns <sup>47,50,51</sup>. While one could expect ID to be related to FD (because of the link between 110 trait similarity and competition for resources) or PD (because biotic interactions can drive the 111 (co)evolutionary history of the species <sup>46,52</sup>), this was not what we observed (Figure 3c-d). Instead, ID 112 represented unique and complementary information to the commonly studied facets of diversity. 113



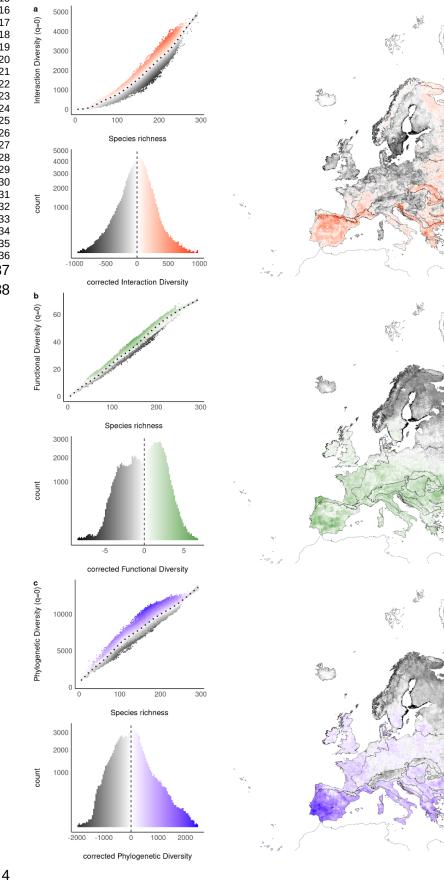
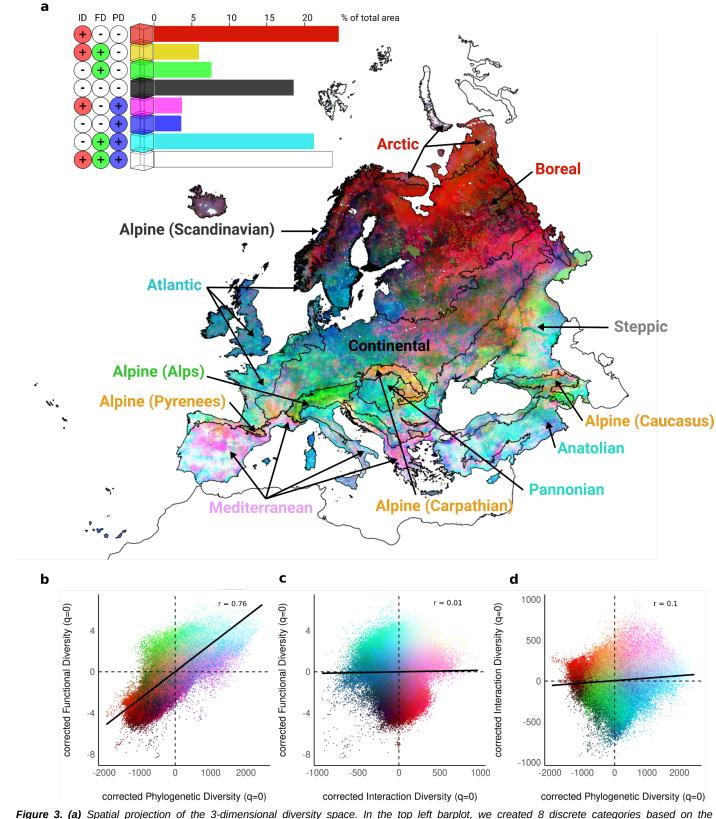


Figure 2. Patterns of Interaction diversity ID (a, in red), Functional diversity FD (b, in green), and Phylogenetic diversity PD (c, in blue). Top left: Relationship between each diversity facet and the species richness. Dotted lines show relationships as fitted by Generalized Additive Models. Bottom left: Distribution of deficits and surpluses of diversities, where model residuals correspond to "corrected diversity" values with deficits (dark shades) and surpluses (red for ID, green for FD, blue for PD). Right: spatial distribution of corrected values for each biodiversity facet, color corresponds to distributions on the left.

**140** Distribution of diversity combinations

141 To investigate the congruence between the interaction, functional and phylogenetic facets of biodiversity, 142 we created a 3-dimensional space where each dimension represents one diversity facet. We further attributed a color channel for each diversity facet (red = ID, green = FD, blue = PD) to visualize all possible 143 144 combinations of biodiversity facets (Figure 1d). Each combination of three color channels (Red, Blue, 145 Green) resulted in a particular color in the RGB color space that corresponds to a given combination of 146 three diversity facets, and allowed us to identify a continuum of ID-FD-PD combinations (Figure 1d). We 147 also interpreted particular types of combinations by discretizing colors based on the combinations of 148 surplus and deficits of each diversity facet (Figure 1d).

This joint analysis of diversity facets highlighted various local combinations of ID-FD-PD, with all kinds of 149 150 combinations being observed in different proportions (Figure 3a). The most commonly observed combinations were ID surpluses with FD and PD deficits (covering 21.8% of the total study area); surpluses 151 in ID, FD, and PD (white, 21.6%); surpluses of FD and PD with deficits in ID (light blue 19.6%); and deficits 152 in ID, FD, and PD (black, 17.3%), which is consistent with the positive correlation observed between FD 153 154 and PD (Figure 3b). The spatial structure of diversity combinations aligned well with many boundaries of European biogeographical regions (Figure 3a), a striking spatial congruency considering that the 155 identification and delimitation of bioregions are based on geographic distribution of vegetation types <sup>53</sup>. 156 Beyond species distribution, biodiversity facets such as phylogenetic diversity already have been shown to 157 match some ecological regions across the globe <sup>54</sup>. ID strongly varies between different regions (e.g. 158 between the Mediterranean region and the Alps, or between the Continental region and the Carpathian 159 160 mountains) and thus further refines boundaries between them. These results suggest that species 161 interactions (along with species co-occurrences and phylogeny) could have a strong structuring effect on 162 (bio)regional species pools. Such a question, however, would require a deeper analysis based on the turnover of interactions within and between regions as regional diversity is connected to local diversity by 163 the turnover in composition between locations. Interestingly, the mapping of diversity combinations also 164 revealed the specificity of several sub-regions within their biogeographical region, e.g the Balkan peninsula 165 166 sub-region in the Mediterranean region, or the Carpathian mountains in the Alpine region. These results further highlight that biotic interaction diversity adds new and independent information and that a dense 167 168 network of trophic interactions can occur in areas of poor functional and phylogenetic diversity.



**Figure 3. (a)** Spatial projection of the 3-dimensional diversity space. In the top left barplot, we created 8 discrete categories based on the combinations of deficits (-) and surpluses (+) of each diversity and reported the number of cells falling in each category. In the map, points are colored by their location in the Red-Green-Blue 3-dimensional color space, with each diversity facet corresponding to a distinct channel: Red channel = Interaction Diversity, Green Channel = Functional Diversity, Blue channel = Phylogenetic Diversity. black shows lowest ID-FD-PD values, white shows highest ID-FD-PD, and so on for each combination. Black lines show the boundaries of the European biogeographical regions. (b) Pair plot of corrected FD (y-axis) VS corrected PD (x-axis), (c) corrected FD (y-axis) VS corrected ID (x-axis), (d) corrected ID (y-axis) VS corrected PD (x-axis). In top right, r is the value of Pearson's product-moment correlation between y and x. Points colors correspond to colors in the map. 177



Southern Europe showed strong diversity surpluses in all diversity facets (white / light color shades in Figure 3a), which confirms the Mediterranean bioregion as a multifaceted biodiversity hotspot <sup>55,56</sup>. This result shows that, for a given number of species, local assemblages of Mediterranean terrestrial vertebrate species were particularly rich in terms of ecological strategies, contained long evolutionary history, and had 183 particularly dense trophic networks. In the Mediterranean basin, the warm climate and the geographical 184 proximity with Africa and Asia explains the high diversity of amphibians and reptiles, as well as the 185 presence of unique evolutionary lineages, leading to high functional and phylogenetic diversities compared 186 to the rest of Europe (Supplemental information - Groups and species distributions across Europe). In 187 addition to these high levels of functional and phylogenetic diversities, the Mediterranean region showed 188 surpluses in interaction diversity, in particular in the subregion of the Balkan peninsula. The densely 189 connected trophic networks observed in the Mediterranean region resulted from (i) numerous top predators in this region (Supplemental information - Maps of relevant network properties) previously identified as 190 191 birds, felids, and snakes predating upon small reptiles and rodents <sup>22</sup>; and (ii) to a lesser extent from a high 192 degree of omnivory in the Iberian peninsula<sup>23</sup>.

Conversely, the northernmost areas tended to show low levels of diversities (black areas in North of 194 195 Scandinavia and Iceland, Figure 3a). The Boreal and Arctic bioregions showed deficits in functional and phylogenetic diversities, but tended to sustain surpluses in interaction diversity (red areas, Figure 3a). In 196 197 these regions, FD deficits were likely to be driven by the cold climate constraining the range of functional 198 traits that can be found in these regions, and similarly for PD via trait conservatism. In particular, the fact that cold temperature limits the presence of ectotherms (amphibians and reptiles) in high latitudes reduces 199 200 functional and phylogenetic diversities, in line with the expected effect of environmental filtering on these diversity facets <sup>37,38</sup>. The consideration of ID brings additional and complementary information since FD-PD 201 deficits are associated with ID surpluses in Northern Europe. The presence of ID surpluses in the Boreal 202 203 and Arctic bioregions likely resulted from a high degree of omnivory (Supplemental information - Maps of relevant network properties), which is known to increase trophic network connectance <sup>26,57</sup>. Species that live 204 under high latitudes tend to be trophic generalists<sup>23</sup> because the higher seasonality in high latitudes 205 206 promotes the evolution of larger niche breadth, in accordance with the latitude-niche breadth hypothesis 57,58 207

208

193

209 Within the Alpine bioregion, different mountain ranges displayed contrasting diversity combinations. The 210 marked differentiation between the Alps and the Carpathian mountains subregions is a striking example 211 supporting the consideration of interaction diversity in biodiversity studies and conservation biogeography. These two mountain ranges located in Central Europe are part of the same Alpine bioregion, which partly 212 213 explains their similarity in terms of functional surpluses and phylogenetic deficits (Figure 2b-c). Based on 214 functional and phylogenetic diversities alone, these two mountain ranges would be considered as similarly 215 diverse - but they are markedly different in terms of interaction diversity. The Carpathians displayed a clear 216 ID surplus in (Figure 3, yellow), while the Alps were deficitary (Figure 3, green). The proximate cause of 217 such difference was the rarity of top predators in the Alps compared to the Carpathians (see Supplemental information - Maps of relevant network properties). Human influence likely explains this discrepancy 218 219 because many apex-predators (bears, wolves, lynx) that are often trophic generalists are still present in the Carpathians, while they were exterminated in the Alps <sup>59</sup>. 220

- 221
- 222223 Potential drivers of diversity facets

While environmental filtering is likely to drive the decrease of FD and PD observed in high latitudes, ID 224 225 might be more influenced by human activities than climate. As such, local deficits of trophic interactions appeared as a marker of high human impact across Europe. This is in line with the negative correlation 226 227 between connectance and human influence previously reported for the same study system <sup>23</sup>, and suggests 228 that the diversity of interactions is influenced by different drivers than functional and phylogenetic diversity. It is, however, noteworthy that other studies reported higher connectance in more human impacted systems 229 230 <sup>9.46</sup>. Indeed, the human-induced relative increase of generalist intermediate predators could counterbalance the decrease in ID due to the loss of a few top predators. The human influence on large-scale diversity has 231 been considered and studied in terms of phylogeny and traits <sup>60</sup>. However, its consequences on large-scale 232 patterns of interaction diversity have been largely overlooked, although they are probably stronger. Indeed, 233 234 human activities have been (and still are) particularly detrimental to large-bodied species 60-63. While this observation is generally viewed as a trait-induced consequence (humans are more detrimental to larger
 animals), it might also be a trophic-induced consequence (humans are more detrimental to apex and
 generalist predators) <sup>45,64</sup>.

238

#### 239 The importance of interaction diversity

A clear understanding of the impact of human activities on ID has yet to emerge. More generally, ID is likely 240 241 to be highly context and taxa dependent, and the understanding of its multi-scale drivers represents a 242 research agenda for the years to come. Among others, the Eltonian shortfall is one big challenge that 243 currently limits the description of ID in many parts of the world where information on biotic interactions is lacking <sup>21</sup>. Here, we overcame this challenge for trophic interactions by inferring local interactions from 244 245 species distributions and their known potential trophic interactions from the literature and expert knowledge (as commonly done, see for example refs <sup>19,25</sup>). While this approach overestimates interactions at a given 246 247 time, "realized" and "potential" number of interactions are very likely to converge in the long term. On the contrary, a field sampling approach would underestimate the realized ID. This underestimation can be quite 248 severe and a massive sampling effort is required to detect most interactions <sup>65</sup>. Combining both approaches 249 (inferring interactions from a metanetwork and species distribution, vs. observing interactions), and 250 comparing their accuracy across a range of temporal and spatial scales will provide valuable insights in 251 252 community ecology and biogeography <sup>66</sup>.

253

Although ID patterns appear robust to data depletion and spatial contexts (see Supplemental information -254 255 Robustness of diversity patterns), whether the patterns described in this study can be extrapolated to other 256 biomes remains an open question. For example, our conclusions from European terrestrial vertebrates 257 might not hold true for tropical rainforests which shelter many trophic specialist species with narrow 258 ecological niches (but comprehensive data on traits and interactions are lacking). Nonetheless, we argue that interaction diversity is a particularly valuable facet for biogeography and conservation planning. 259 Although this view has been empirically challenged <sup>36</sup>, more densely connected trophic networks are 260 generally considered as desirable from a conservation point of view <sup>9</sup>. Areas with surpluses of interactions 261 represent interaction networks that are expected to be more robust to cascading species extinctions <sup>34</sup>, and 262 consequently more resilient to perturbations. Coupled with its apparent sensitivity to human activities <sup>23</sup>, 263 264 interaction diversity might be viewed as a marker of both ecosystem degradation and resistance to future 265 degradation. We argue that a general consideration of interaction diversity as an important and meaningful diversity facet alongside the functional and phylogenetic diversities should be a priority for macroecology 266 and conservation biogeography. 267

## 269 STAR Methods

#### 270 Study area and data

Study area. The study area, hereafter referred to as "Europe", included the entire European subcontinent (with Macaronesia and Iceland) plus Anatolia to include a complete picture of the North Mediterranean coast (Figure 1a). The study area was divided into 117,000 cells on a 10×10 km equal-size area grid (ETRS89). Within the study area, we considered ten biogeographical regions defined by the European Environment Agency <sup>67</sup>: Alpine, Anatolian, Arctic, Atlantic, Boreal, Black Sea, Continental, Macaronesia, Mediterranean, and Steppic. These bioregions are large scale ecological units based on an interpretation of geobotanical data <sup>68</sup>, and represent areas with homogeneous ecological context.

278

279 Species distributions. We extracted the distributions for all terrestrial vertebrates naturally occurring within the study area from Maiorano et al. (2013). Species distributions for 509 bird, 288 mammal, 250 reptile and 280 281 104 amphibian species were mapped by combining the IUCN extent of occurrence for each species with 282 their habitat requirements. A species was considered potentially present in a 10×10 km cell if the grid cell 283 met the three following criteria: i) is within the species extent of occurrence, ii) contains at least one 300x300m area of primary habitat for the species, i.e. habitat where the species can persist (defined by 284 285 experts and published literature) and iii) meets species requirements in terms of elevation and distance from water. A full description of species distribution data and definition of primary habitat can be found in 286 Maiorano et al. 2013<sup>27</sup>. In addition, we used the percentage of primary habitat of the species in each cell as 287 a proxy for the probability to find the species in a random locality within this cell. For example, we 288 considered that if the primary habitat of a species covered 80% of the cell, the probability to find the 289 290 species in a random locality of the cell was 0.8. As such, it represents a proxy for the probability of 291 presence of the species within the cell and was used as a weight in the entropy-based diversity measures 292 (i.e when g=1) provided as supplementary analyses.

293

294 Functional traits. We gathered biological trait data from Thuiller et al. 2015<sup>28</sup>, excluding traits describing diet (and thus trophic interactions) and traits for larvae and juveniles. Our analysis was based on four life-295 history and ecological traits common to mammals, amphibians, birds and reptiles. The only quantitative trait 296 297 was body mass [grams, log-transformed]. The three other traits were multichoice nominal variables coded 298 by binary values. Feeding behavior was coded by four binary columns: opportunistic feeder, active hunter, 299 browser, grazer. Nesting location was coded by eleven binary columns: tree/hole/fissure in bark, ground, 300 rocks. building/artificial. underground water, cave/fissures/burrows, lodge. temporary water, brooks/springs/small rivers, puddles/ponds/pools/small lakes, brackish waters. Activity time was coded by 301 four binary columns: nocturnal, crepuscular, diurnal, arrhythmic. These traits were selected because they 302 represent informative niche dimensions linked to the use and acquisition of resources in space and time, 303 and are related to ecosystem functioning <sup>69,70</sup>. A thorough description of traits and the list of publications 304 305 where the data were gathered is available in supplementary material from <sup>28</sup> available at https://royalsocietypublishing.org/doi/suppl/10.1098/rstb.2014.0005. 306

We computed the pairwise dissimilarities (distances) of this trait matrix using a mixed variable coefficient of distance (using function dist.ktab in ade4) that generalizes Gower's general coefficient of distance to allow the treatment of various statistical types of variables when calculating distances <sup>71</sup>. Euclidean distance was used for body mass, and Jaccard index was used for the four other multichoice nominal variables (S3 coefficient of in Gower and Legendre 1986 <sup>72</sup>).

312

Phylogenetic tree. We used the 100 phylogenetic trees for European terrestrial vertebrates assembled and published by Roquet at al. 2014 <sup>29</sup>. We chose these phylogenetic trees as they are the only species-level phylogenies encompassing all european vertebrates, and have already been valuably used to depict phylogenetic diversity in this context vertebrates in the past <sup>28,32,73,74</sup>.

- 317 318
- 9

Trophic networks. We used data on species trophic interactions from the metaweb of European terrestrial vertebrates, (Tetra-EU 1.0, <sup>30</sup>). This metaweb is based on expert knowledge, published information and field guides. Potential trophic links between a predator and a prey were identified from published accounts of their observation, morphological similarities between potential prey and literature-referenced prey or -in the absence of this information- the diet of the predator's sister species. The metaweb of European terrestrial vertebrates contained 1,164 species and a total of 50,408 potential trophic interactions. The full dataset and methods description can be found in ref <sup>30</sup>.

326

327 In order to maximize the species coverage for each diversity, we allowed for different sets of species to be 328 used to compute ID, FD and PD. For ID we retained 1149 species for which we had information on their 329 European distribution range and trophic interactions; for FD we retained 1009 species for which we had information on their European distribution range and functional traits; for PD we retained 993 species for 330 331 which we had information on their European distribution range and phylogeny. This varying set of species should have low impact on the assessment of diversities as ID, FD and PD were corrected by their 332 corresponding taxonomic richness to compute surpluses and deficits. In order to investigate the potential 333 bias resulting from the variation of species coverage across space and diversities, we computed diversities 334 based on the same set of 884 species for which we had all shared information. The resulting diversity 335 336 patterns were similar when considering the 884 species or varying set of species (see Supplemental 337 information - Diversities based on the same set of 884 species).

338

339 Diversity measures and the 3-dimensional diversity space

Within each 10x10km cell, we used Hill numbers <sup>31</sup> to compute FD, PD, and ID. In this framework, diversity 340 341 values are converted into effective numbers of species, the Hill numbers. When considering taxonomic 342 diversity, the effective number of species is the number of equally abundant species necessary to produce 343 the observed value of diversity (an analogue to the concept of effective population size in genetics). This 344 approach has then been generalized to incorporate species phylogenetic relatedness and species functional distances. We used the framework from Chao et al. 2014<sup>31</sup> implemented in the R package hillR 345 for phylogenetic and functional diversity and in the package econetwork <sup>18</sup> for interaction diversity. We 346 347 computed each diversity as a Hill number analogous to a measure of richness by setting q=0 (ignoring 348 abundance). The ID richness was the sum of trophic links formed by the species present in the cell, the FD 349 richness was the sum of functional pairwise gower distances between species in the cell, and the PD 350 richness was the mean sum of the branch lengths of the phylogenetic tree connecting all species present in 351 the cell <sup>1</sup> across the 100 trees.

We focused our study on richness-based results (q=0) as they are the easiest to interpret, but we also 352 analyzed and showed results based on Shannon entropy in the Supplemental information - Results based 353 354 on Shannon entropy. To compute the results as a Shannon entropy, we set q=1 and used the % of species' 355 primary habitat within the cell as the probability to find the species in the cell. More precisely, when g=1 the 356 ID entropy is the Shannon entropy over the interaction weights (product of the two species abundances), 357 the FD is the Shannon entropy of effective number of species-pairs with unit-distance between species, 358 and the PD is the mean Shannon entropy of the effective total branch length across the 100 trees. More details on the calculations of FD and PD can be found in Chao et al. 2014<sup>31</sup>, and in Ohlmann et al. 2019<sup>18</sup> 359 360 for ID.

361

We corrected FD, PD, ID richness and Shannon entropy for the number of species in the cell (i.e taxonomic 362 richness) based on the set of species used to compute each diversity. We fitted a thin plate spline 363 regression, a particular Generalized Additive Model (GAM), to predict each diversity measure from species 364 richness. The residuals of each model (one for each diversity facet and order g) were retained as the 365 species richness corrected value of the diversity, with positive residuals considered as surplus and negative 366 residuals considered as deficits given the species richness <sup>32,33</sup>. In other words, a deficit (or surplus) 367 indicates a lower (or higher, respectively) diversity value than expected given the local species richness 368 369 (Figure 1).

371 To investigate the congruence between the interaction, functional and phylogenetic facets of biodiversity, 372 we created a 3-dimensional space where each dimension represents one diversity facet. In order to 373 visualize all possible combinations of biodiversity facets, we attributed a color channel for each diversity 374 facet (red = ID, green = FD, blue = PD) where the residual values for each diversity were rescaled to 0-255 375 value in the corresponding color channel (Figure 1). Hence, each combination of three color channels (Red, 376 Blue, Green) results in a particular color in the RGB color space that corresponds to a given combination of 377 three diversity facets, and allows us to identify a continuum of ID-FD-PD combinations depicted in figure 1.d. We can also interpret particular types of combinations by discretizing colors based on the combinations 378 379 of surplus and deficits of each diversity. As shown in Figure 1d and Figure 3a, Red identifies surpluses of 380 ID and FD associated with deficits in FD and PD (ID>0, FD<0, PD<0); Yellow identifies surpluses of ID and 381 FD associated with deficits in PD (ID>0, FD>0, PD<0); Green identifies deficits in ID and PD associated with FD surpluses (ID<0, FD>0, PD<0); Black identifies deficits in ID, PD and FD, Pink identifies surpluses 382 of ID and PD associated with FD deficits; Dark blue identifies surpluses of PD associated with ID and FD 383 deficits; Light blue identifies surpluses of PD and FD associated with ID deficits; white identifies surpluses 384 in ID, PD and FD. 385

- 386 SUPPLEMENTAL INFORMATION
- 387 Supplemental Information includes six appendix and can be found at:

## 388 AUTHOR CONTRIBUTIONS

P.G and W.T. conceived the study, with early advice from L.O.C, C.B, G.P. and T.M. P.G. performed all analyses and wrote the first version of the manuscript with inputs from W.T, L.O.C, C.B, G.P. and T.M. All authors contributed substantially to the interpretation of the results and to the writing of the manuscripts and its revisions.

# 393 ACKNOWLEDGEMENTS

This study has received funding from the ERA-Net BiodivERsA - Belmont Forum, with the national funder Agence Nationale de la Recherche (ANR-18-EBI4-0009), part of the 2018 Joint call BiodivERsA-Belmont Forum call (project 'FutureWeb'). WT was also supported by the Agence Nationale pour la Recherche through the FORBIC (ANR-18-MPGA-0004) and EcoNet (ANR-18-CE02-0010) projects. We warmly thank Florian Maderspacher for handling the paper, Brittany Pugh, Richard Field, Kamran Safi and one anonymous referee for their valuable reviews and comments. Interaction data depletions

- 400
- 401
- 402

- 403 References
- 404 1. Faith, D.P. (1992). Conservation evaluation and phylogenetic diversity. Biol. Conserv. *61*, 1–10.
- 405 2. Tilman, D. (2001). Functional diversity. Encyclopedia of biodiversity 3, 109–120.
- Pollock, L.J., O'Connor, L.M.J., Mokany, K., Rosauer, D.F., Talluto, M.V., and Thuiller, W. (2020).
   Protecting Biodiversity (in All Its Complexity): New Models and Methods. Trends Ecol. Evol. 35, 1119–
   1128.
- 409 4. Grime, J.P. (1997). Biodiversity and Ecosystem Function: The Debate Deepens. Science *2*77, 1260– 410 1261.
- 5. Tilman, D., Isbell, F., and Cowles, J.M. (2014). Biodiversity and Ecosystem Functioning. Annu. Rev.
  Ecol. Evol. Syst. 45, 471–493.
- Brose, U., Blanchard, J.L., Eklöf, A., Galiana, N., Hartvig, M., R. Hirt, M., Kalinkat, G., Nordström,
   M.C., O'Gorman, E.J., Rall, B.C., et al. (2017). Predicting the consequences of species loss using sizestructured biodiversity approaches. Biol. Rev. Camb. Philos. Soc. *92*, 684–697.
- 416 7. Schneider, F.D., Brose, U., Rall, B.C., and Guill, C. (2016). Animal diversity and ecosystem functioning
  417 in dynamic food webs. Nat. Commun. 7, 12718.
- Thompson, J.N. (1997). Conserving Interaction Biodiversity. In The Ecological Basis of Conservation: Heterogeneity, Ecosystems, and Biodiversity, S. T. A. Pickett, R. S. Ostfeld, M. Shachak, and G. E. Likens, eds. (Springer US), pp. 285–293.
- 421 9. Tylianakis, J.M., Laliberté, E., Nielsen, A., and Bascompte, J. (2010). Conservation of species
  422 interaction networks. Biol. Conserv. *143*, 2270–2279.
- Hooper, D.U., Solan, M., Symstad, A., Diaz, S., Gessner, M.O., Buchmann, N., Degrange, V., Grime,
  P., Hulot, F., Mermillod-Blondin, F., et al. (2002). Species diversity, functional diversity and ecosystem
  functioning. Biodiversity and ecosystem functioning: synthesis and perspectives, 195–208.
- 11. De Palma, A., Kuhlmann, M., Bugter, R., Ferrier, S., Hoskins, A.J., Potts, S.G., Roberts, S.P.M.,
  Schweiger, O., and Purvis, A. (2017). Dimensions of biodiversity loss: Spatial mismatch in land-use
  impacts on species, functional and phylogenetic diversity of European bees. Divers. Distrib. *23*, 1435–
  1446.
- Schweiger, A.K., Cavender-Bares, J., Townsend, P.A., Hobbie, S.E., Madritch, M.D., Wang, R.,
  Tilman, D., and Gamon, J.A. (2018). Plant spectral diversity integrates functional and phylogenetic components of biodiversity and predicts ecosystem function. Nat Ecol Evol *2*, 976–982.
- 13. Dyer, L.A., Walla, T.R., Greeney, H.F., Stireman, J.O., III, and Hazen, R.F. (2010). Diversity of
  Interactions: A Metric for Studies of Biodiversity: Interaction Diversity. Biotropica *42*, 281–289.
- 435 14. Jacquet, C., Moritz, C., Morissette, L., Legagneux, P., Massol, F., Archambault, P., and Gravel, D.
  436 (2016). No complexity–stability relationship in empirical ecosystems. Nat. Commun. 7, 1–8.
- Burkle, L.A., and Alarcón, R. (2011). The future of plant-pollinator diversity: understanding interaction
   networks across time, space, and global change. Am. J. Bot. *98*, 528–538.
- 439 16. Burkle, L.A., Myers, J.A., and Belote, R.T. (2016). The beta-diversity of species interactions:
   440 Untangling the drivers of geographic variation in plant-pollinator diversity and function across scales.
- 441 Am. J. Bot. *103*, 118–128.
- Sandom, C., Dalby, L., Fløjgaard, C., Kissling, W.D., Lenoir, J., Sandel, B., Trøjelsgaard, K., Ejrnaes,
  R., and Svenning, J.-C. (2013). Mammal predator and prey species richness are strongly linked at
  macroscales. Ecology *94*, 1112–1122.

- 445 18. Ohlmann, M., Miele, V., Dray, S., Chalmandrier, L., O'Connor, L., and Thuiller, W. (2019). Diversity
   446 indices for ecological networks: a unifying framework using Hill numbers. Ecol. Lett. 22.
- Poisot, T., Canard, E., Mouillot, D., Mouquet, N., and Gravel, D. (2012). The dissimilarity of species
   interaction networks. Ecol. Lett. *15*, 1353–1361.
- 449 20. Hortal, J., de Bello, F., Diniz-Filho, J.A.F., Lewinsohn, T.M., Lobo, J.M., and Ladle, R.J. (2015). Seven
  450 Shortfalls that Beset Large-Scale Knowledge of Biodiversity. Annu. Rev. Ecol. Evol. Syst. *46*, 523–549.
- Poisot, T., Bergeron, G., Cazelles, K., Dallas, T., Gravel, D., MacDonald, A., Mercier, B., Violet, C.,
  Vissault, S., and Chapman, D. (2021). Global knowledge gaps in species interaction networks data. J.
  Biogeogr. 48, 1552–1563.
- 22. O'Connor, L.M.J., Pollock, L.J., Braga, J., Ficetola, G.F., Maiorano, L., Martinez-Almoyna, C.,
  Montemaggiori, A., Ohlmann, M., and Thuiller, W. (2020). Unveiling the food webs of tetrapods across
  Europe through the prism of the Eltonian niche. J. Biogeogr. *47*, 181–192.
- 457 23. Braga, J., Pollock, L.J., Barros, C., Galiana, N., Montoya, J.M., Gravel, D., Maiorano, L.,
  458 Montemaggiori, A., Ficetola, G.F., Dray, S., et al. (2019). Spatial analyses of multi-trophic terrestrial
  459 vertebrate assemblages in Europe. Glob. Ecol. Biogeogr. *28*, 1636–1648.
- 460 24. Cumming, G.S., Bodin, Ö., Ernstson, H., and Elmqvist, T. (2010). Network analysis in conservation
   461 biogeography: Challenges and opportunities. Diversity and Distributions *16*, 414–425.
- 462 25. Gravel, D., Baiser, B., Dunne, J.A., Kopelke, J.P., Martinez, N.D., Nyman, T., Poisot, T., Stouffer, D.B.,
  463 Tylianakis, J.M., Wood, S.A., et al. (2019). Bringing Elton and Grinnell together: a quantitative
  464 framework to represent the biogeography of ecological interaction networks. Ecography *42*, 401–415.
- 465 26. Baiser, B., Gravel, D., Cirtwill, A.R., Dunne, J.A., Fahimipour, A.K., Gilarranz, L.J., Grochow, J.A., Li,
  466 D., Martinez, N.D., McGrew, A., et al. (2019). Ecogeographical rules and the macroecology of food
  467 webs. Glob. Ecol. Biogeogr. 28, 1204–1218.
- 468 27. Maiorano, L., Amori, G., Capula, M., Falcucci, A., Masi, M., Montemaggiori, A., Pottier, J., Psomas, A.,
  469 Rondinini, C., Russo, D., et al. (2013). Threats from climate change to terrestrial vertebrate hotspots in
  470 Europe. PLoS One *8*, e74989.
- Thuiller, W., Maiorano, L., Mazel, F., Guilhaumon, F., Ficetola, G.F., Lavergne, S., Renaud, J., Roquet,
  C., and Mouillot, D. (2015). Conserving the functional and phylogenetic trees of life of European
  tetrapods. Philos. Trans. R. Soc. Lond. B Biol. Sci. *370*, 20140005.
- Roquet, C., Lavergne, S., and Thuiller, W. (2014). One tree to link them all: a phylogenetic dataset for
   the European tetrapoda. PLoS Curr. 6.
- 476 30. Maiorano, L., Montemaggiori, A., Ficetola, G.F., O'Connor, L., and Thuiller, W. (2020). TETRA-EU 1.0:
  477 A species-level trophic metaweb of European tetrapods. Glob. Ecol. Biogeogr. 29, 1452–1457.
- 478 31. Chao, A., Chiu, C.-H., and Jost, L. (2014). Unifying Species Diversity, Phylogenetic Diversity,
  479 Functional Diversity, and Related Similarity and Differentiation Measures Through Hill Numbers. Annu.
  480 Rev. Ecol. Evol. Syst. 45, 297–324.
- 32. Zupan, L., Cabeza, M., Maiorano, L., Roquet, C., Devictor, V., Lavergne, S., Mouillot, D., Mouquet, N.,
  Renaud, J., and Thuiller, W. (2014). Spatial mismatch of phylogenetic diversity across three vertebrate
  groups and protected areas in Europe. Diversity and Distributions *20*, 674–685.
- 33. Safi, K., Cianciaruso, M.V., Loyola, R.D., Brito, D., Armour-Marshall, K., and Diniz-Filho, J.A.F. (2011).
  Understanding global patterns of mammalian functional and phylogenetic diversity. Philos. Trans. R.
  Soc. Lond. B Biol. Sci. *366*, 2536–2544.
- 487 34. Gilbert, A.J. (2009). Connectance indicates the robustness of food webs when subjected to species
  488 loss. Ecol. Indic. *9*, 72–80.
- 489 35. Thébault, E., and Fontaine, C. (2010). Stability of ecological communities and the architecture of

- 490 mutualistic and trophic networks. Science 329, 853–856.
- 491 36. Heleno, R., Devoto, M., and Pocock, M. (2012). Connectance of species interaction networks and conservation value: Is it any good to be well connected? Ecol. Indic. *14*, 7–10.
- 493 37. McGill, B.J., Enquist, B.J., Weiher, E., and Westoby, M. (2006). Rebuilding community ecology from
  494 functional traits. Trends Ecol. Evol. *21*, 178–185.
- 38. Webb, C.O., Ackerly, D.D., McPeek, M.A., and Donoghue, M.J. (2002). Phylogenies and Community
  Ecology. Annu. Rev. Ecol. Syst. *33*, 475–505.
- 497 39. Münkemüller, T., Gallien, L., Pollock, L.J., Barros, C., Carboni, M., Chalmandrier, L., Mazel, F.,
  498 Mokany, K., Roquet, C., Smyčka, J., et al. (2020). Dos and don'ts when inferring assembly rules from
  499 diversity patterns. Glob. Ecol. Biogeogr. *164*, S165.
- 40. Mayfield, M.M., and Levine, J.M. (2010). Opposing effects of competitive exclusion on the phylogenetic structure of communities. Ecol. Lett. *13*, 1085–1093.
- 502 41. Stebbins, G.L. (1974). Flowering Plants: Evolution Above the Species Level (Belknap Press of Harvard University Press).
- 42. Duffy, J.E., Cardinale, B.J., France, K.E., McIntyre, P.B., Thébault, E., and Loreau, M. (2007). The functional role of biodiversity in ecosystems: incorporating trophic complexity. Ecol. Lett. *10*, 522–538.
- 43. Estes, J. a., Terborgh, J., Brashares, J.S., Power, M.E., Berger, J., Bond, W.J., Carpenter, S.R.,
  Essington, T.E., Holt, R.D., Jackson, J.B.C., et al. (2011). Trophic downgrading of planet Earth.
  Science 333, 301–306.
- 44. de Visser, S.N., Freymann, B.P., and Olff, H. (2011). The Serengeti food web: empirical quantification and analysis of topological changes under increasing human impact. J. Anim. Ecol. *80*, 484–494.
- 45. Pauly, D., Christensen, V., V., Dalsgaard, J., Froese, R., and Torres, F., Jr (1998). Fishing down marine food webs. Science *279*, 860–863.
- 46. Tylianakis, J.M., and Morris, R.J. (2017). Ecological Networks Across Environmental Gradients. Annu.
  Rev. Ecol. Evol. Syst. 48, 25–48.
- 47. Wiens, J.J., and Graham, C.H. (2005). Niche Conservatism: Integrating Evolution, Ecology, and Conservation Biology. Annu. Rev. Ecol. Evol. Syst. *36*, 519–539.
- 48. Peterson, A.T., Soberon, J., and Sanchez-Cordero, V., V. (1999). Conservatism of ecological niches in evolutionary time. Science *285*, 1265–1267.
- 49. Lavergne, S., Mouquet, N., Thuiller, W., and Ronce, O. (2010). Biodiversity and Climate Change:
  Integrating Evolutionary and Ecological Responses of Species and Communities. Annu. Rev. Ecol.
  Evol. Syst. *41*, 321–350.
- 50. Wiens, J.J., Ackerly, D.D., Allen, A.P., Anacker, B.L., Buckley, L.B., Cornell, H.V., Damschen, E.I.,
  Jonathan Davies, T., Grytnes, J.-A., Harrison, S.P., et al. (2010). Niche conservatism as an emerging
  principle in ecology and conservation biology. Ecol. Lett. *13*, 1310–1324.
- 525 51. Münkemüller, T., Boucher, F., Thuiller, W., and Lavergne, S. (2015). Common conceptual and 526 methodological pitfalls in the analysis of phylogenetic niche conservatism. Funct. Ecol. *29*, 627–639.
- 527 52. Williams, R.J., and Martinez, N.D. (2000). Simple rules yield complex food webs. Nature *404*, 180– 183.
- 529 53. Noirfalise, A. (1987). Map of the Natural Vegetation: Of the Member Countries of the European
  530 Economic Community and the Council of Europe.
- 531 54. Holt, B.G., Lessard, J.-P., Borregaard, M.K., Fritz, S.A., Araújo, M.B., Dimitrov, D., Fabre, P.-H., 532 Graham, C.H., Graves, G.R., Jønsson, K.A., et al. (2013). An update of Wallace's zoogeographic
- 14

- regions of the world. Science 339, 74–78.
- 534 55. Myers, N., Fonseca, G. a. B., Mittermeier, R. a., Fonseca, G. a. B., and Kent, J. (2000). Biodiversity 535 hotspots for conservation priorities. Nature *403*, 853–858.
- 56. Mittermeier, R.A., Turner, W.R., Larsen, F.W., Brooks, T.M., and Gascon, C. (2011). Global
  Biodiversity Conservation: The Critical Role of Hotspots. In Biodiversity Hotspots: Distribution and
  Protection of Conservation Priority Areas, F. E. Zachos and J. C. Habel, eds. (Springer Berlin
  Heidelberg), pp. 3–22.
- 540 57. Cirtwill, A.R., Stouffer, D.B., and Romanuk, T.N. (2015). Latitudinal gradients in biotic niche breadth 541 vary across ecosystem types. Proc. Biol. Sci. *282*.
- 542 58. Vázquez, D.P., and Stevens, R.D. (2004). The latitudinal gradient in niche breadth: concepts and 543 evidence. Am. Nat. *164*, E1–19.
- 54. Kaczensky, P., Chapron, G., Von Arx, M., Huber, D., Andrén, H., and Linnell, J. (2013). Status,
  545 Management and Distribution of Large Carnivores Bear, Lynx, Wolf & Wolverine in Europe (Verlag
  546 nicht ermittelbar).
- 60. Fritz, S.A., Bininda-Emonds, O.R.P., and Purvis, A. (2009). Geographical variation in predictors of mammalian extinction risk: big is bad, but only in the tropics. Ecol. Lett. *12*, 538–549.
- 61. Gaston, K.J., and Blackburn, T.M. (1995). Birds, body size and the threat of extinction. Philos. Trans.
  R. Soc. Lond. B Biol. Sci. 347, 205–212.
- 62. Cardillo, M., Mace, G.M., Jones, K.E., Bielby, J., Bininda-Emonds, O.R.P., Sechrest, W., Orme, C.D.L.,
  and Purvis, A. (2005). Multiple causes of high extinction risk in large mammal species. Science *309*,
  1239–1241.
- 63. Böhm, M., Williams, R., Bramhall, H.R., McMillan, K.M., Davidson, A.D., Garcia, A., Bland, L.M.,
  Bielby, J., and Collen, B. (2016). Correlates of extinction risk in squamate reptiles: the relative
  importance of biology, geography, threat and range size. Glob. Ecol. Biogeogr. *25*, 391–405.
- 557 64. Brose, U. (2010). Improving nature conservancy strategies by ecological network theory. Basic Appl.
   558 Ecol. 11, 1–5.
- 65. Chacoff, N.P., Vázquez, D.P., Lomáscolo, S.B., Stevani, E.L., Dorado, J., and Padrón, B. (2012).
  Evaluating sampling completeness in a desert plant-pollinator network. J. Anim. Ecol. *81*, 190–200.
- 66. Fortuna, M.A., Nagavci, A., Barbour, M.A., and Bascompte, J. (2020). Partner Fidelity and Asymmetric Specialization in Ecological Networks. Am. Nat. *196*, 382–389.
- 563 67. European Environmental Agency, EEA (2016). Biogeographical regions.
   564 https://www.eea.europa.eu/data-and-maps/data/biogeographical-regions-europe-3.
- 68. Roekaerts, M. (2002). The Biogeographical Regions Map of Europe. Basic principles of its creation and overview of its development (European Environmental Agency, EEA).
- 567 69. Luck, G.W., Lavorel, S., McIntyre, S., and Lumb, K. (2012). Improving the application of vertebrate
   568 trait-based frameworks to the study of ecosystem services. Journal of Animal Ecology *81*, 1065–1076.
- 569 70. Sekercioglu, C.H. (2006). Increasing awareness of avian ecological function. Trends Ecol. Evol. 21,
   570 464–471.
- 571 71. Pavoine, S., Vallet, J., Dufour, A.-B., Gachet, S., and Daniel, H. (2009). On the challenge of treating
  572 various types of variables: application for improving the measurement of functional diversity. Oikos
  573 118, 391–402.
- 574 72. Gower, J.C., and Legendre, P. (1986). Metric and Euclidean properties of dissimilarity coefficients. J.
   575 Classification *3*, 5–48.

- France France
- 579 74. Saladin, B., Thuiller, W., Graham, C.H., Lavergne, S., Maiorano, L., Salamin, N., and Zimmermann,
  580 N.E. (2019). Environment and evolutionary history shape phylogenetic turnover in European tetrapods.
  581 Nat. Commun. *10*, 249.
- 582 75. Cadotte, M.W., Cardinale, B.J., and Oakley, T.H. (2008). Evolutionary history and the effect of 583 biodiversity on plant productivity. Proc. Natl. Acad. Sci. U. S. A. *105*, 17012–17017.
- 584 76. Cadotte, M.W., Jonathan Davies, T., Regetz, J., Kembel, S.W., Cleland, E., and Oakley, T.H. (2010).
  585 Phylogenetic diversity metrics for ecological communities: integrating species richness, abundance
  586 and evolutionary history. Ecol. Lett. *13*, 96–105.
- 587 77. Davies, T.J., and Buckley, L.B. (2011). Phylogenetic diversity as a window into the evolutionary and
  588 biogeographic histories of present-day richness gradients for mammals. Philos. Trans. R. Soc. Lond. B
  589 Biol. Sci. 366, 2414–2425.
- 590 78. Pollock, L.J., Thuiller, W., and Jetz, W. (2017). Large conservation gains possible for global
  biodiversity facets. Nature *546*, 141–144.
- 592 79. Cadotte, M.W., Carscadden, K., and Mirotchnick, N. (2011). Beyond species: functional diversity and
   593 the maintenance of ecological processes and services. J. Appl. Ecol. 48, 1079–1087.
- 594 80. Díaz, S., and Cabido, M. (2001). Vive la différence: plant functional diversity matters to ecosystem 595 processes. Trends Ecol. Evol. *16*, 646–655.
- Second State Stat
- 82. Brun, P., Zimmermann, N.E., Graham, C.H., Lavergne, S., Pellissier, L., Münkemüller, T., and Thuiller,
  W. (2019). The productivity-biodiversity relationship varies across diversity dimensions. Nat. Commun.
  10, 5691.
- 83. de Bello, F., Lavorel, S., Hallett, L.M., Valencia, E., Garnier, E., Roscher, C., Conti, L., Galland, T.,
  Goberna, M., Májeková, M., et al. (2021). Functional trait effects on ecosystem stability: assembling
  the jigsaw puzzle. Trends Ecol. Evol. *36*, 822–836.
- 84. Loreau, M. (2014). Reconciling utilitarian and non-utilitarian approaches to biodiversity conservation.
  Ethics Sci. Environ. Polit. *14*, 27–32.
- 85. Mazel, F., Wüest, R.O., Gueguen, M., Renaud, J., Ficetola, G.F., Lavergne, S., and Thuiller, W.
  (2017). The Geography of Ecological Niche Evolution in Mammals. Curr. Biol. 27, 1369–1374.
- 86. Chesson, P., and Kuang, J.J. (2008). The interaction between predation and competition. Nature 456,
  235–238.
- 87. Chase, J.M., Abrams, P.A., Grover, J.P., Diehl, S., Chesson, P., Holt, R.D., Richards, S.A., Nisbet,
  R.M., and Case, T.J. (2002). The interaction between predation and competition: a review and
  synthesis. Ecol. Lett. *5*, 302–315.
- 613 88. Stouffer, D.B., Sales-Pardo, M., Sirer, M.I., and Bascompte, J. (2012). Evolutionary conservation of 614 species' roles in food webs. Science *335*, 1489–1492.
- 615 89. Poisot, T., Mouquet, N., and Gravel, D. (2013). Trophic complementarity drives the biodiversity-616 ecosystem functioning relationship in food webs. Ecol. Lett. *16*, 853–861.
- 617 90. Vane-Wright, R.I., Humphries, C.J., and Williams, P.H. (1991). What to protect?—Systematics and the 618 agony of choice. Biol. Conserv. *55*, 235–254.
- 619 91. Isaac, N.J.B., Turvey, S.T., Collen, B., Waterman, C., and Baillie, J.E.M. (2007). Mammals on the 16

- 620 EDGE: conservation priorities based on threat and phylogeny. PLoS One 2, e296.
- 621 92. Rodrigues, A.S.L., and Gaston, K.J. (2002). Maximising phylogenetic diversity in the selection of 622 networks of conservation areas. Biol. Conserv. *105*, 103–111.
- 93. Devictor, V., Mouillot, D., Meynard, C., Jiguet, F., Thuiller, W., and Mouquet, N. (2010). Spatial
  mismatch and congruence between taxonomic, phylogenetic and functional diversity: the need for
  integrative conservation strategies in a changing world. Ecol. Lett. *13*, 1030–1040.
- 94. Peterson, A.T., Soberón, J., Pearson, R.G., Anderson, R.P., Martínez-Meyer, E., Nakamura, M., and
   Bastos Araujo, M. (2011). Ecological niches and geographic distributions (Princeton University Press).
- 95. Díaz, S., Settele, J., Brondízio, E., Ngo, H., Guèze, M., Agard, J., Arneth, A., Balvanera, P., Brauman,
- K., Butchart, S., et al. (2020). Summary for policymakers of the global assessment report on
   biodiversity and ecosystem services of the Intergovernmental Science-Policy Platform on Biodiversity
- 631 and Ecosystem Services (IPBES).