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Seasonal, diurnal and tidal variations of dissolved inorganic carbon and pCO₂ in surface waters of a temperate coastal lagoon (Arcachon, SW France)

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Key words: coastal zone; dissolved inorganic carbon; water pCO₂; water-air CO₂ fluxes; tidal, diurnal, seasonal variations; physical, biological, chemical processes.

1 **Abstract**

2 We report on diurnal, tidal and seasonal variations of dissolved inorganic carbon (DIC), water
3 partial pressure of CO₂ (pCO₂) and associated water-air CO₂ fluxes in a tidal creek of a
4 temperate coastal lagoon with 70 % of intertidal flats, during eight tidal/diurnal cycles and two
5 consecutive years covering all seasons. Surface waters of the lagoon were always slightly
6 oversaturated in CO₂ with respect to the atmosphere with an average pCO₂ value of 496 ± 36
7 ppmv. Seasonally, subsurface water pCO₂ values were controlled by both temperature and
8 biological / tidal advection effects that compensated each other and resulted in weak annual
9 variations. High-resolution temporal pCO₂ records reveal that highest fluctuations (192 ppmv)
10 occurred at the tidal/diurnal scale as a result of biological activity, advection from the tidal flat
11 and porewater pumping that all contributed to water pCO₂ and carbonate chemistry variations.
12 Total Alkalinity (TA) versus salinity plots suggest a net production of alkalinity in the lagoon
13 attributed to benthic carbonate dissolution and/or anaerobic degradation of organic matter. We
14 specifically highlighted that for a same salinity range, during flooding, daytime pCO₂ were
15 generally lower than nighttime pCO₂ values because of photosynthesis, whereas during ebbing,
16 daytime pCO₂ were higher than nighttime pCO₂ values because of heating. Waters in the lagoon
17 were a relatively weak CO₂ source to the atmosphere over the year compared to other estuarine
18 and lagoon waters elsewhere, and to sediment-air fluxes measured simultaneously by
19 atmospheric Eddy Covariance (EC) in the Arcachon lagoon. Because of low values and small
20 variations of the air-sea pCO₂ gradient, the variability of fluxes calculated using the piston
21 velocity parameterization was greatly controlled by the wind speed at the diurnal and, to a lesser
22 extent, seasonal time scales. During the emersion, the comparison of these pCO₂ data in the
23 tidal creek with EC fluxes measured 1.8 km away on the tidal flat suggests high heterogeneity
24 in air-sea CO₂ fluxes, both spatially and at short time scales according to the inundation cycle

25 and the wind speed. In addition to tidal pumping when the flat becomes emerged, our data
26 suggest that lateral water movement during the emersion of the flat generates strong spatial
27 heterogeneity in water-air CO₂ flux.

28

29 **1. Introduction**

30 Coastal zones represent key systems in biogeochemical cycle couplings between land, oceans
31 and the atmosphere, processing considerable amounts of matter, energy and nutrients (Borges
32 et al. 2005; Cole et al. 2007; Cai 2011). Despite its relatively modest surface area (7 % of the
33 global ocean surface), the coastal zone accounts for 14-30 % of all oceanic primary production
34 and 90 % of sedimentary organic carbon mineralization (Gattuso et al. 1998; Mantoura et al.
35 1991; Pernetta and Milliman 1995). Carbon fluxes within and between coastal subsystems and
36 their alteration by climate and anthropogenic changes are substantial. It is essential to
37 understand and accurately account for the factors regulating these fluxes and how they affect
38 the ocean and global carbon budgets (Bauer et al. 2013).

39 The coastal ocean consists of different but tightly connected ecosystems including estuaries,
40 tidal wetlands, lagoons and the continental shelf. Global CO₂ emissions from estuaries have
41 been estimated at 0.2-0.4 Pg yr⁻¹ (Cai 2011; Borges and Abril 2011; Borges et al. 2005; Laruelle
42 et al. 2010), with fluxes disproportionately high in comparison with their global ocean area
43 portion (near 0.2 %). To the contrary, other coastal systems such as tidal wetlands and
44 continental shelves fix 0.55 and 0.25 Pg yr⁻¹ of atmospheric CO₂ respectively (Bauer et al.
45 2013). The spatiotemporal heterogeneity and complexity of coastal systems in terms of carbon
46 processes and fluxes make difficult to precisely quantify each sub-system's carbon balance. For
47 instance, carbon fluxes in tidal systems such as estuaries and small deltas are relatively well
48 characterized (Laruelle et al. 2010; Borges and Abril 2011; Chen et al. 2013). To the contrary,
49 other systems such as fjords, lagoons and marine embayments are neglected despite a large

50 relative surface among coastal systems, a generally high net primary production and a strong
51 sensitivity to eutrophication phenomenon (Kjerfve 1985; Caumette et al. 1996; Koné et al.
52 2009; Polsenaere et al. 2012a; Cotovicz et al. 2015). As a consequence, carbon flux
53 measurements are needed for an accurate estimate of global and regional carbon budgets.

54 In coastal ecosystems, dissolved inorganic carbon (DIC) concentrations, and particularly partial
55 pressures of CO₂ (pCO₂) are driven by several thermodynamic and biotic factors, induced by
56 river water inputs, tidal exchanges, mixing of water masses, photosynthesis and
57 aerobic/anaerobic degradation of organic matter, precipitation/dissolution of calcium
58 carbonate, benthic/pelagic couplings and air-water exchanges (Cai 2011; Gazeau et al. 2004;
59 Cotovicz et al. 2015; Ribas-Ribas et al. 2011). Efforts have been made to understand the
60 seasonal and inter-annual processes affecting carbon system dynamics over coastal ecosystems
61 (Yates et al. 2007). Seasonal measurements of the carbonate system parameters (pCO₂, Total
62 Alkalinity, DIC, and/ or pH) have been performed for instance across salinity gradient transects
63 or strategically located stations along lagoon channels or estuaries (Frankignoulle et al. 1996;
64 Cai et al. 1999; Koné et al. 2009; Ribas-Ribas et al. 2011; Burgos et al. 2018; Wang et al. 2018;
65 Vaz et al. 2019). The influence of smaller-scale rhythms (tidal and diurnal variations) on
66 seasonal and annual carbon dynamics and budgets has also been studied through direct diurnal
67 cycles for instance over east monsoon coastal systems and subtropical bays (Dai et al. 2009;
68 Yates et al. 2007), temperate sea, lagoon, estuary, river, channel and creek (Saderne et al. 2013;
69 Bozec et al. 2011; Borges and Frankignoulle 1999; Ribas-Ribas et al. 2013; Burgos et al. 2018),
70 sub-Antarctic island coastal waters (Delille et al. 2009), tropical and equatorial mangrove
71 waters (Maher et al. 2013; Bouillon et al. 2007; Borges et al. 2003) and continental and
72 subtropical saltmarsh and marsh-dominated estuary systems (Wang and Cai, 2004; Wang et al.
73 2016; Wang et al. 2018). For instance, in a tidal creek among the Duplin River salt marsh-
74 estuary coastal ecosystem (Georgia, USA), Wang et al. (2018) measured strong seasonal and

75 tidal/diurnal pCO₂ variations with values ranging from 500 ppmv at high tide to 4000 ppmv at
76 low tide and to 1600 ppmv at high tide to 12,000 ppmv at low tide during coldest and warmest
77 months, respectively. Horizontal advection with river water entrance to the creek at high tide
78 was mentioned to explain lower measured values. To the contrary, for higher values observed
79 at low tide, creek bank pore-water drainage associated to tidal pumping process, as observed
80 over mangroves or wetlands (Maher et al. 2013; Neubauer and Anderson 2003) was highlighted
81 (Wang et al. 2018). Over tidally-influenced shallow systems, all these processes are generally
82 but differently involved and can in turn significantly and specifically influence atmospheric
83 CO₂ fluxes.

84 CO₂ fluxes at the water-air interface can be measured directly using the Eddy Covariance
85 (Zemmelink et al. 2009; Polsenaere et al. 2012a), floating chambers (Frankignoulle et al. 1998)
86 or be calculated from water pCO₂ measurements and a given gas transfer velocity. CO₂ flux
87 computations can be subject to large uncertainties because of the difficulty in accurately
88 assessing the gas transfer velocity (Raymond and Cole 2001, Vachon et al. 2010). This bias is
89 potentially more important when the water-air CO₂ gradient is small. Diurnal and tidal
90 variations in water pCO₂ of dynamic coastal ecosystems can even add uncertainties on air-sea
91 CO₂ flux estimations. For example, in the Guadalquivir estuary, Ribas-Ribas et al. (2013)
92 observed pCO₂ values fluctuating from source to sink over the same tidal cycle.

93 In the present study we report tidal, diurnal and seasonal dynamics of inorganic carbon and
94 associated air-sea CO₂ fluxes in a tidal creek of the temperate coastal Arcachon lagoon (SW
95 France). The lagoon is a typical tidal flat of the French Atlantic coast subjected to both marine
96 and continental influences. A complex channel and tidal creek waters network drains the tidal
97 flat during ebb tides. Tidal pumping through anoxic pore-water seeping from mud sediments to
98 tidal creek waters is an important component of the biogeochemical functioning of the lagoon
99 in terms of nutrient dynamics (Deborde et al., 2008a). In this paper, DIC parameters and

100 associated water-air CO₂ fluxes as estimated based on several *in situ* tidal/diurnal cycles (six
101 24-hour and two 12-hour cycles) carried at different seasons during two years (2008 and 2009)
102 are presented. We describe and explain small-scale (diurnal and tidal) and seasonal pCO₂
103 variability along with relevant environmental controls in studied tidal creek waters. We
104 compute air-sea CO₂ fluxes, and discuss associated variability according to small (diurnal and
105 tidal) and seasonal scales and also methodology from coastal regional/global carbon budgets
106 point of view. The computed water-air CO₂ fluxes estimated during these tidal/diurnal cycles
107 are also compared with those measured in summer 2008 by atmospheric Eddy Covariance (EC)
108 over the tidal flat (Polsenaere et al. 2012a).

109

110 **2. Materials and Methods**

111 **2.1. Study area**

112 The Arcachon bay is a temperate macro-tidal lagoon of 174 km² on the southwestern Atlantic
113 coast of France (44°40' N, 01°10' W) (Fig. 1). Several studies have focused on the system in
114 particular its hydrodynamics through modelling simulations (Plus et al. 2009; Fauvelle et al.
115 2018); *Zostera* seagrass meadow dynamic (Auby and Labourg 1996; Plus et al. 2010; Cognat
116 et al. 2018) and relationships with sediment hydrodynamics or redox status (Ganthy et al. 2013;
117 Delgard et al. 2013; Deborde et al. 2008a,b); nutrient and carbon dynamics and export from the
118 Arcachon catchment through *in situ* adapted sampling and modelling strategies (Auby et al.
119 1994; Rimmelin et al. 1998; Canton et al. 2012; Polsenaere et al. 2012b; Polsenaere and Abril
120 2012); and in terms of benthic and pelagic primary production measurements through various
121 *in situ*, laboratory and modelling approaches (Polsenaere et al. 2012a; Migné et al. 2016; Glé
122 et al. 2008; Plus et al. 2015).

123 This triangle-shaped bay is surrounded by the coastal plain of the *Landes de Gascogne*, and
124 communicates with the Atlantic Ocean through a narrow channel 8 km in length (Fig. 1). With

125 a mean depth of 4.6 m, the shallow lagoon presents semi-diurnal tides with amplitudes varying
126 from 0.8 to 4.6 m (Plus et al. 2009). During a tidal cycle, the flat exchanges approximately 264
127 $\times 10^6 \text{ m}^3$ and $492 \times 10^6 \text{ m}^3$ of water with the ocean, respectively during average neap and spring
128 tides. The flushing time of the lagoon ranges between 12.4 and 17.4 days in winter and summer
129 respectively (Plus et al. 2009). Water temperatures in the lagoon vary from 6 °C in winter to
130 22.5 °C in summer, and water salinity varies from 22 to 35 according to freshwater input
131 variations during the year.

132 The flats are tidally submerged by relatively saline waters, and receive moderate amounts of
133 freshwater with an annual input of $813 \times 10^6 \text{ m}^3$ ($1.8 \times 10^6 \text{ m}^3$ at each tidal cycle), of which 8
134 % is from groundwater, 13 % is from rainfall and 79 % is from rivers and small streams
135 (Rimmelin et al. 1998). In total, the carbon export from the watershed to the Arcachon lagoon
136 was estimated at $15,870 \text{ t C yr}^{-1}$ or $6 \text{ t C km}^{-2} \text{ yr}^{-1}$, mostly in the form of dissolved organic
137 carbon (DOC) (35 %), DIC (24 %) as excess CO_2 and $\text{DIC}_{\text{equilibrium}}$ (i.e. the theoretical DIC
138 concentration at the atmospheric equilibrium), and dissolved CO_2 rapidly lost as degassing to
139 the atmosphere (34 %) (Polsenaere et al. 2012b). Tidal pore-water drainage at low tide is not
140 documented for carbon, but contributes to respectively 55 and 15 % of the dissolved inorganic
141 phosphate (DIP) and nitrate (DIN) input to the lagoon waters (Deborde et al. 2008a).

142 The lagoon surface is composed of 57 km^2 of channels (33 % of the lagoon surface area), with
143 a maximum depth of 25 m at the mouth of the lagoon. These channels drain a large muddy tidal
144 flat of 117 km^2 (67 % of the lagoon). *Zostera noltei* seagrass beds are particularly extensive and
145 colonize up to 60 % of this intertidal area (i.e. 70 km^2) between -1.9 m and $+0.8 \text{ m}$ relative to
146 local Mean Sea Level (Amanieu, 1967). Annual mean net primary production of *Zostera noltei*
147 expressed as carbon fixation was estimated at $362.9 \pm 88.1 \text{ t C km}^{-2} \text{ yr}^{-1}$ (Ribaudou et al. 2017).
148 Polsenaere et al. (2012a) and Migné et al. (2016) gave annual values of the same order of
149 magnitude through atmospheric EC and benthic chamber measurements with 456 and 263 t C

150 $\text{km}^{-2} \text{yr}^{-1}$ respectively. However, seagrass beds in the lagoon declined by 33 % since the end of
151 the 80's whereas the exact causes remain unclear, leading to a C fixation decline from $24846 \pm$
152 6030 t C yr^{-1} (1989) to $16564 \pm 4020 \text{ t C yr}^{-1}$ (2007) (Plus et al. 2009; Plus et al. 2015; Ribaud
153 et al. 2017). The microphytobenthic (MPB) communities also represent a significant proportion
154 of the total benthic primary production, which is estimated at between 104-114, 628 and 788 t
155 $\text{C km}^{-2} \text{yr}^{-1}$ (Auby personal communication; Polsenaere et al. 2012a; Migné et al. 2016).
156 Together, these two categories of benthic primary production represent more than half of the
157 total primary production of the flat. Annual integrated phytoplankton primary production has
158 been estimated at $103 \text{ t C km}^{-2} \text{yr}^{-1}$ which could represent 30 % of the total primary production
159 of the lagoon and places Arcachon within the low to moderate phytoplankton primary
160 production systems (Glé et al. 2008). Microphytobenthic resuspension is also supposed to
161 significantly contribute to planktonic production in Arcachon as highlighted by Savelli et al.
162 (2019) on the Brouage mudflat of the Marennes-Oléron Bay up North along the French Atlantic
163 coast with a 43 % resuspended MPB primary production at this location. The presence of
164 circular tidal pools on the flat also contributes to the biogeochemical functioning of Arcachon
165 bay especially at tidal and diurnal scales (Rigaud et al. 2018) through large spatio-temporal
166 variations in water temperature and irradiance influencing MPB activity and benthic oxygen,
167 nutrients, and reduced compound fluxes.

168 Finally, net CO_2 fluxes between the lagoon and the atmosphere measured at the ecosystem scale
169 by atmospheric EC generally showed small negative (influx) and positive (efflux) values (-13
170 $\mu\text{mol m}^{-2} \text{s}^{-1}$ for influxes and $19 \mu\text{mol m}^{-2} \text{s}^{-1}$ for effluxes). Emersion during the day was almost
171 always associated with a net uptake of atmospheric CO_2 due to an enhanced benthic primary
172 production at low tide. In contrast, during immersion (day and night) and emersion at night,
173 CO_2 fluxes were positive, negative or close to zero, depending on the season and the study site
174 (Polsenaere et al. 2012a).

175 **2.2. Sampling strategy and laboratory analysis**

176 In total, six 24-hour cycles (April, July, September 2008, and April, June and September 2009)
177 and two 12-hour cycles (November 2008 and January 2009) were carried out in a subtidal creek
178 at the centre of the lagoon (Fig. 1). For each cycle, water height (H), salinity (S), temperature
179 (T), partial pressure of CO₂ in the water (pCO₂), total alkalinity (TA) and dissolved inorganic
180 carbon isotopic ratio ($\delta^{13}\text{C-DIC}$) were measured at the subsurface (0.5 m depth). T and S were
181 measured every minute with an YSI multiparameter probe on board and H was measured every
182 hour by the SHOM (Service Hydrographique et Océanographique de la Marine) at Arcachon
183 (Eyrac pier, 44°39.9001'N 01°09.8130'W). Before each cycle, the salinity (conductivity)
184 sensor was checked and calibrated with 10 and 50 mS cm⁻¹ solutions. Water pCO₂ was
185 measured every minute with a marble-type equilibrator system (Frankignoulle et al. 2001;
186 Polsenaere et al. 2012b). An Infra-Red Gas Analyzer (LI-COR, LI-820) was used to measure
187 the pCO₂ in dry air equilibrated with seawater. The LI-820 was calibrated at the laboratory one
188 day before the field experiment using three gas standards of 0, 500 ± 10 ppm and 2959 ± 59
189 ppm. The equilibrator consists of a Plexiglas cylinder (height: 90 cm, diameter: 16 cm) that is
190 filled with marbles to increase the exchange surface area. Water pumped by a peristaltic pump
191 (Masterflex, 1 L min⁻¹), runs from the top to the bottom of the equilibrator, and air is pumped
192 upwards (1 L min⁻¹). The pCO₂ of air equilibrates with the pCO₂ of water and is then measured
193 by the LI-COR after being dried by a Dierite grain tube. Response time of the equilibrator as
194 determined in the laboratory is shorter than 5 minutes (Cotovicz et al. 2016).

195 Discrete water samples were also taken every hour of the tidal cycle for TA and $\delta^{13}\text{C-DIC}$ in
196 the channel subsurface waters by using a 5L Niskin sampler. Total alkalinity (TA) was
197 measured by titration with HCl 0.1 N on 100 mL filtered samples and was calculated by a Gran
198 function linearisation (Gran 1952) between pH 4.2 and 3. The reproducibility between the
199 measures was better than ±5 μmol kg⁻¹. Accuracy of TA measurements was checked on regular

200 titrations of a secondary standard of 0.2 μm filtered seawater with a well-known TA
201 concentration, previously calibrated by comparison with a standard from Scripps Institution of
202 Oceanography. The $\delta^{13}\text{C}$ -DIC measurements were made following Gillikin and Bouillon
203 (2007). In 100 mL vials that were filled to the top, a headspace was first created with Helium
204 gas to obtain a volume of approximately 20 % of the total volume of the vial. Then, 0.3 mL of
205 warm 85 % phosphoric acid was added to transform the carbonate forms into CO_2 . To ensure
206 gas equilibration, the vials were shaken and placed upside down for 1.5 h. Measurements were
207 performed by coupling an elemental analyzer (EA; Carlo Erba NC2500) to an Isotope Ratio
208 Mass Spectrometer (IRMS; Micromass Isoprime) equipped with a manual gas injection. We
209 injected 3 mL of headspace gas from the vial headspace. $\delta^{13}\text{C}$ -DIC was calibrated against a
210 laboratory standard (45 mg of Na_2CO_3 were dissolved in a sealed vial flushed with He gas, with
211 3 mL of warm 85 % phosphoric acid H_3PO_4). This standard had been calibrated against a
212 certified standard (NBS19, -1.96 ‰) using a dual-inlet IRMS. The isotopic value of the standard
213 Na_2CO_3 was -4.5 ± 0.2 ‰. Finally, the equation of Miyajima et al. (1995) was applied to correct
214 for the isotopic fractionation of the CO_2 between the headspace and the water phase and to
215 calculate the $\delta^{13}\text{C}$ of the total DIC. The repeatability was approximately ± 0.1 ‰ between
216 samples. The stable isotopic composition of DIC ($\delta^{13}\text{C}$ -DIC) varies over a large range in
217 terrestrial and coastal waters since carbon reservoirs that act as a source of DIC (soil,
218 groundwater, bedrocks and atmosphere) have distinct isotopic signatures (Yang et al. 1996).
219 The $\delta^{13}\text{C}$ of atmospheric CO_2 is about -7.5 ‰, whereas carbonate rocks have a $\delta^{13}\text{C}$ of about 0
220 ‰ (Mook et al. 1983). In a system where soil CO_2 is primarily derived from decomposition of
221 plant organic matter, the CO_2 produced has a $\delta^{13}\text{C}$ - CO_2 value close to the initial substrate (i.e.,
222 -30 to -24 ‰ in the case of C3 plants and -16 to -10 ‰ in the case of C4 plants). Aquatic
223 primary production, in contrast, tends to increase $\delta^{13}\text{C}$ -DIC and generate strong diel variations

224 (Parker et al. 2005). Finally, gas exchange along subsurface waters generates an isotopic
225 equilibration with the atmosphere and makes the $\delta^{13}\text{C}$ -DIC less negative.

226

227 **2.3. Calculations and statistical analysis**

228 Dissolved inorganic carbon (DIC) concentrations and species (bicarbonate HCO_3^- and
229 carbonate CO_3^{2-} ions) were calculated from the measured salinity, temperature, pCO_2 and TA
230 using the carbonic acid constants sets from by Mehrbach et al. (1973) as modified by Dickson
231 and Millero (1987), the borate acidity constant from Lee et al. (2010), the K_{HSO_4} constant from
232 Dickson (1990) and the CO_2 solubility coefficient of Weiss (1974). The CO_2 System
233 Calculation program (version 2.1.) developed by Lewis and Wallace (1998) performed all
234 calculations.

235 Seasonal temperature (T_{pCO_2}) versus non-temperature (N_{pCO_2}) effects on measured water
236 pCO_2 during the eight tidal cycles were studied applying equations from Takahashi et al. (2002)
237 as follows:

$$238 \quad T_{\text{pCO}_2} = \text{pCO}_{2\text{mean}} \times \exp[0.0423(T_{\text{obs}} - T_{\text{mean}})] \quad (1)$$

$$239 \quad N_{\text{pCO}_2} = \text{pCO}_{2\text{obs}} \times \exp[0.0423(T_{\text{mean}} - T_{\text{obs}})] \quad (2)$$

240 Where $\text{pCO}_{2\text{obs}}$, T_{obs} , $\text{pCO}_{2\text{mean}}$, T_{mean} are respectively the measured pCO_2 , temperatures at each
241 time step and the annual mean pCO_2 and temperature calculated across the whole measured
242 dataset. T_{pCO_2} indicates the pCO_2 variations around the mean pCO_2 that would be expected
243 only from temperature fluctuations that occurred over the sampling period (2008-2009). N_{pCO_2}
244 represent pCO_2 variations due to biological activity or non-temperature effects (advection, tidal
245 pumping and benthic-pelagic coupling). The constant 0.0423 corresponds to the temperature
246 effect on pCO_2 in isochemical conditions ($q \ln \text{pCO}_2 / qT$), i.e. $+4.23 \text{ \% } ^\circ\text{C}^{-1}$.

247 Gas transfer velocities (k_{600}) and air-sea CO_2 fluxes for the eight tidal cycles were computed
248 each hour at high-tide day and night periods. Wind speed values normalized to a 10 meters

249 height (U_{10}) values were computed from wind speed values measured at 9 meters high by the
250 Lège-Cap Ferret Meteo-France station (44°37.900'N 01°14.900'W, 12.5 km far from the 24-H
251 site, Fig. 1) using the Amorocho and DeVries (1980) equation. Gas transfer velocities (k) were
252 estimated according to parameterizations of Wanninkhof et al. (1992), Raymond and Cole
253 (2001) and Abril et al. (2009), noted W92, RC01 and A09, respectively. The W92 equation has
254 been proposed for oceanic water, whereas the RC01 and A09 equations were developed for
255 estuarine waters. For the latter equation, hourly suspended particulate matter (SPM)
256 concentrations measured during every cycle (data not shown) and bottom water velocity
257 currents estimated with the hydrodynamic model from Ifremer (MARS3D, Lazure and Dumas,
258 2008) were used. Air-sea CO_2 fluxes were then calculated from k , water and air pCO_2 values.
259 The gas transfer coefficients normalized to a Schmidt number of 600 (k_{600}) estimated with the
260 three parameterizations were converted to the gas transfer velocity of CO_2 (k) at *in situ*
261 temperature and salinity following the procedure of Jähne et al. (1987). Air-water CO_2 fluxes
262 were estimated using the formulation as followed:

$$263 \quad F_{CO_2} = \alpha k \Delta pCO_2 \quad (3)$$

264 where F_{CO_2} is the vertical CO_2 exchange at the air-water interface, α is the CO_2 solubility
265 coefficient (Weiss 1974), k is the gas transfer velocity, and ΔpCO_2 is the gradient between water
266 pCO_2 measured by the equilibrator and air pCO_2 set to a mean value of 390 ppm according to
267 simultaneous Eddy Covariance (EC) measurements done in the lagoon (44°42.9858'N
268 01°08.6160'W, 1.829 km far from the 24-H site, Fig. 1) in 2008 and 2009 (Polsenaere et al.
269 2012a). For comparison, U_{10} and air-sea CO_2 flux values obtained in July 2008 from
270 simultaneous EC measurements (Polsenaere et al. 2012a) and water carbon parameters (pCO_2)
271 during the July 2008 cycle are also presented.

272 Data post-processing (graphs and statistics) was performed using the GraphPad Prism version
273 6.00 software (La Jolla California USA, www.graphpad.com). The Shapiro-Wilk test was used

274 to test the normality of the data (p-value below 0.05). Due to non-normality of our data, Mann-
275 Whitney and Kruskal-Wallis (p-value below 0.05) were performed to detect significant
276 differences in carbon and associated parameters between tidal/diurnal cycles (seasons) and also
277 across diurnal and tidal rhythms (emersion around low tide during the day (LT/Day), emersion
278 at night (LT/Night), immersion around high tide during the day (HT/Day) and immersion at
279 night (HT/Night)) (emersion and immersion periods last during 4 and 8 hours twice per day in
280 the Arcachon lagoon). The Dunn's post-hoc multiple comparisons' test to Kruskal-Wallis was
281 chosen to detect significant differences among groups.

282

283 **3. Results**

284 **3.1. Water pCO₂ and associated parameter temporal variations**

285 *3.1.1. At the seasonal time scale*

286 In 2008/09 at our 2.5 meter deep (on average, hydrographic zero) subtidal creek in the Arcachon
287 lagoon, water temperatures averaged 17.7 ± 4.2 °C over the years 2008 and 2009 and ranged
288 between 8.9 ± 0.4 °C in January 2009 and 22.2 ± 0.6 °C in July 2008 (Table 1, Fig. 2). Water
289 mean salinity value was 30.5 ± 3.4 with a particularly low mean value measured in January
290 2009 (Table 1, 23.2 ± 1.5 , Fig. 3).

291 Bicarbonates (HCO₃⁻), carbonates (CO₃²⁻) and the sum of dissolved CO₂ and carbonic acid
292 H₂CO₃ concentrations (CO₂^{*}) represented respectively 91 to 95 %, 3 to 8 % and about 1 % of
293 the whole DIC (data not shown). TA values followed DIC patterns with a lowest mean value
294 of 1.646 ± 0.086 mmol kg⁻¹ in January 2009 and a highest mean value 2.255 ± 0.021 mmol kg⁻¹
295 ¹ in September 2008 (Fig. 3, Table 1). δ¹³C-DIC mean values were close to 0 ‰ and ranged
296 between -1.1 ± 1.1 ‰ in April 2008 (minimum values in winter/spring) and -0.2 ± 0.2 ‰ in
297 June 2009 (maximum values in summer/autumn). Particularly low negative values up to -3.7
298 ‰ were measured in April 2008 (Table 1, Fig. 3). Over the whole studied period, global

299 significant and positive TA, DIC and $\delta^{13}\text{C}$ -DIC values versus salinity values were obtained
300 ($\text{TA}=0.05143*\text{S} + 0.5079$, $\text{R}^2: 0.88$; $\text{DIC}=0.03815*\text{S} + 0.7594$, $\text{R}^2: 0.82$; $\delta^{13}\text{C}\text{-DIC}=0.08697*\text{S}$
301 $- 3.284$, $\text{R}^2: 0.15$; $p<0.0001$, Fig. 4a, c, d). At each cycle (season) apparent 0 end-member TA
302 and *in situ* catchment TA values were computed from the Y-intercept of TA versus S cross-
303 correlations (Fig. 4a) and *in situ* TA measurements carried out at the same time over the
304 Arcachon lagoon watershed watercourses (Polsenaere et al. 2012b) calculated here as
305 discharge-weighted TA means, respectively (Fig. 4b). Apparent 0 end-member TA values were
306 generally higher than *in situ* catchment values except during the end of summer and autumn
307 period (September and November 2008, Fig. 4b).

308 Water pCO_2 averaged 496 ± 36 ppmv over the sampling years and were always oversaturated
309 whatever the season and period of the day and tide, with regards to the atmospheric equilibrium
310 of 390 ppmv (Figs. 2, 3, 5, Table 1). Mean values ranged from 461 ± 14 ppmv in July 2008 to
311 530 ± 39 ppmv in September 2009 (minimum and maximum values in summer/autumn) with
312 in general (except between November and July 2008 and between April and September 2009)
313 only significant differences between seasons (Kruskal-Wallis test, $p<0.05$) though lower than
314 those observed at the diurnal/tidal scale (Table 1, Figs. 2, 3, 5).

315 Seasonal variations in *in situ* measured temperature, pCO_2 values and in temperature (TpCO_2 ,
316 tpCO_2) versus non-temperature (NpCO_2) effects on pCO_2 are presented in Figure 2. During
317 2008, pCO_2 values showed weak seasonal variations increasing from April 2008 (474 ± 14
318 ppmv) to September 2008 (515 ± 36 ppmv) before decreasing towards November 2008 (463
319 ppmv). In 2009, pCO_2 increased from January to April (525 ± 14 ppmv) before decreasing
320 towards June (490 ± 27 ppmv) and increasing again to September (530 ± 39 ppmv, Table 1,
321 Fig. 2) as in 2008. In 2008, NpCO_2 values showed larger seasonal variations first decreasing
322 from April to July (557 ± 22 and 381 ± 16 ppmv respectively), and then increasing towards
323 November (583 ± 27 ppmv). To the contrary in 2009, a decrease was first observed from

324 January (698 ± 16 ppmv) to June 2009 (417 ± 24 ppmv) before increasing towards September
325 (463 ± 37 ppmv). In 2008 and 2009 respectively, TpCO_2 followed exactly the opposite patterns.
326 Peak-to-peak seasonal amplitudes in *in situ* temperature and pCO_2 values were 9.9 °C and 52
327 ppmv in 2008 and 12.6 °C and 50 ppmv in 2009 (Fig. 2, Table 1).

328 **3.1.2. At the diurnal/tidal time scales**

329 Significant variations in inorganic carbon and associated parameters occurred at the
330 diurnal/tidal scales (Table 1, Figs. 2, 3, 5). Water heights ranged from 0.4 m (July 2008 and
331 June 2009) at low tide to 4.5 m (July 2008, January and June 2009) at high tide (Fig. 3). The
332 largest salinity variation occurred during winter and spring with for instance 5.2 and 6.7 salinity
333 unit difference between low and high tides in January and April 2009 respectively. During these
334 seasons, at low tide, the lowest salinity values of 20.4 and 24.8 respectively were observed. To
335 the contrary, during summer and autumn seasons, salinity values were higher (minimum values
336 30.2 at low tide in July 2008 and June 2009) and varied by less than 2.4 units throughout the
337 tidal cycle (Table 1, Figs. 3, 5). Water temperature variations at the tidal/diurnal scales were
338 small and ranged from $1.4/1.3$ °C in September 2008/2009 to 3.3 °C in June 2009 (Table 1,
339 Figs. 3, 5).

340 With regards to inorganic carbon parameters, DIC and TA concentrations showed weak
341 diurnal/tidal variations with significant differences between high and low tides except in
342 September 2008 (ranges of 0.078 mmol kg^{-1} for TA and 0.069 mmol kg^{-1} for DIC) and in June
343 2009 where lower values were measured during low tide periods, concomitant to lower
344 salinities. $\delta^{13}\text{C}$ -DIC values were significantly higher during high tide in July 2008 and June
345 2009 (ranges: -1.7 to 0.2 ‰ and -0.4 to 0.2 ‰, respectively) (Table 1, Fig. 3).

346 Water pCO_2 tidal/diurnal variations ranged between 50 ppmv in winter and spring (51 and 54
347 ppmv in January 2009 and April 2008 respectively) to 150 - 200 ppmv in summer and autumn
348 (148 and 192 ppmv in September 2009 and 2008) (Figs. 2, 3, 5). Whatever the season and hour,

349 water pCO₂ values were always oversaturated compared to the atmospheric value of 390 ppm;
350 at minimum over all tidal/diurnal cycles, pCO₂ value dropped to 405 ppmv in September 2008)
351 (Table 1, Figs. 2, 3, 5). Some significant pCO₂ variations also occurred in the creek between
352 nighttime and daytime and between the emersion and the immersion of the tidal flat (LT_D,
353 LT_N, HT_D and HT_N periods, see Fig. 5 caption and M&M section) (Non-parametric Mann-
354 Whitney test, p<0.05) (Figs. 3 and 5, Table 1). In winter-spring (April 2008, November 2008,
355 January 2009 and April 2009), during daytime pCO₂ were significantly higher and daytime
356 salinity were lower than at night. To the contrary, in summer-autumn (July 2008, September
357 2008/09 and June 2009), daytime pCO₂ values were significantly lower than nighttime pCO₂
358 and salinity were higher at night. During the eight cycles, pCO₂ values at low tide were
359 significantly higher than values at high tide (salinity values at low tide were also below high
360 tide values) (Figs. 2, 3, 5, Table 1).

361 A correlation analysis showed significant negative correlations between water pCO₂ and
362 salinity (or water height) values (i.e. Spearman coefficients pCO₂ versus salinity of -0.69 and -
363 0.92, p<0.01 in July 2008 and September 2009 respectively), except in September and
364 November 2008 (Fig. 3). pCO₂ values were also negatively correlated to δ¹³C-DIC values in
365 July 2008 and over all 2019 cycles (R: -0.63, -0.72 and -0.58 in July 2008, January and June
366 2009 respectively, p<0.01). TA concentrations appeared to be always significantly and
367 positively correlated to salinity (or water height) (R: 0.93 and 0.80, p<0.01 in September 2008
368 and April 2009 for instance) and to DIC concentrations (Figs. 2, 3). δ¹³C-DIC values were
369 significantly and positively correlated to salinity and water height (H) values in July 2008 and
370 over all 2019 cycles (R: 0.45, 0.70 in July 2008 and April 2009 respectively, p<0.01), to TA
371 values in September 2008 and over all cycles in 2019 (R: 0.46, 0.59 in September 2008 and
372 2009 respectively, p<0.01) and to DIC concentrations over all cycles in 2009 (R: 0.61 and 0.76
373 in April and June 2009 respectively) (Fig. 4).

374 **3.2. Gas transfer velocities and atmospheric CO₂ fluxes from the Arcachon lagoon waters**

375 Wind speeds (U_{10}) ranged between 2.41 ± 1.05 and 7.74 ± 1.77 m s⁻¹ in September and
376 November 2008 respectively with significant differences noticed between September 2008 -
377 January 2009 (minimum values) and July - November 2008 - September 2009 (maximum
378 values) (Kruskal-Wallis and Dunn's post-test, $p < 0.01$, Table 2). k_{600} values were in average
379 estimated at 8.54 ± 6.75 , 17.29 ± 15.76 and 17.00 ± 6.23 cm h⁻¹ according to W92, RC01 and
380 A09 equations respectively (see Material and Methods, Table 2). Minimum mean values
381 occurred in September 2008 - January 2009 (1.82 ± 1.25 , 4.4 ± 1.36 and 9.58 ± 3.3 cm h⁻¹ in
382 January 2009 with the three methods respectively, Table 2) whereas the highest values were
383 reached in November 2008 - September 2009 (19.37 ± 8.37 , 33.33 ± 19.51 and 28.38 ± 6.12
384 cm h⁻¹ in November 2008 with the three methods respectively, Table 2). In April 2008 and
385 September 2009 at high tide, k_{600} (and U_{10}) values were also found significantly different
386 between day (higher values) and night (lower values) periods (Mann-Whitney test, $p < 0.05$)
387 whatever used methods (Table 2).

388 Mean water-air CO₂ fluxes were estimated at 0.27 ± 0.22 , 0.56 ± 0.54 and 0.55 ± 0.22 mmol
389 m⁻² h⁻¹ according to W92, RC01 and A09 equations respectively (Table 2). Minimum mean
390 values were observed in January 2009 (0.06 ± 0.04 , 0.14 ± 0.06 and 0.29 ± 0.13 mmol m⁻² h⁻¹
391 with the three methods respectively) whereas the highest values were reached in September
392 2009 (0.62 ± 0.66 , 1.67 ± 2.36 and 0.82 ± 0.47 mmol m⁻² h⁻¹ with the three methods respectively,
393 Table 2). In April 2008 and September 2009, water-air CO₂ flux values were also found
394 significantly different between day (higher values) and night (lower values) periods (Mann-
395 Whitney test, $p < 0.05$) as it was the case for k_{600} and U_{10} values whatever used methods (Table
396 2). Water-air fluxes estimated from A09 equation were significantly higher than values
397 estimated from W92 (and RC01) equations (Kruskal-Wallis and Dunn's post-test, $p < 0.01$)

398 except in November 2008 - April 2009 - September 2009 where no significant difference was
399 observed between methods.

400 In July 2008, an atmospheric EC mast was deployed in the Arcachon lagoon simultaneously
401 and close (at 1.8 km) to the 24 hours cycle location (Fig. 1, Table 2; Polsenaere et al. 2012a).
402 During the whole tidal cycle in July 2008, EC fluxes averaged $0.45 \pm 7.71 \text{ mmol m}^{-2} \text{ h}^{-1}$ and
403 ranged from -17.98 to 43.19 $\text{mmol m}^{-2} \text{ h}^{-1}$. EC water-air CO_2 fluxes during the immersion were
404 on average $-1.39 \pm 2.67 \text{ mmol m}^{-2} \text{ h}^{-1}$ and ranged from -12.38 to 2.10 $\text{mmol m}^{-2} \text{ h}^{-1}$ with
405 significantly lower values during daytime period (Mann-Whitney test, $p < 0.05$). Significant
406 differences were computed between direct water-air CO_2 fluxes measured during the immersion
407 at the EC site and calculated CO_2 fluxes from pCO_2 records at the channel site according to
408 RC01 - A09 equations (Kruskal-Wallis and Dunn's post-test, $p < 0.01$; Table 2).

409

410 **4. Discussion**

411 **4.1. Seasonal compensation of thermal and non-thermal effects on water pCO_2**

412 In the Arcachon lagoon, seasonal variations in water inorganic carbon parameters and
413 especially pCO_2 were relatively weak and on the lower range for waters influenced by an
414 intertidal ecosystem when compared to other coastal ecosystems worldwide (Borges et al. 2005;
415 Table 3). We observed a maximal pCO_2 variation of 196 ppmv between the lowest and the
416 highest values measured during each 12- or 24-hour cycles and the maximum pCO_2 (averaged
417 over each 24-hour cycle) amplitude barely reached 70 ppmv.

418 At the seasonal scale over coastal bays, thermal but also non-thermal effects can significantly
419 influence water pCO_2 variations. In the studied lagoon, we observed that, seasonally, biological
420 effects that consisted in more heterotrophy in winter and more autotrophy in summer, was offset
421 by the thermal effect induced by the variation of water temperature (from 8.9 to 22.2 °C)
422 resulting in relatively constant pCO_2 values throughout the year. Furthermore, other non-

423 thermal effects were induced by mixing of freshwater in the lagoon, particularly in January
424 2009 when salinity decreased down to 20.4 (Table 1, Figs. 3, 4). River waters discharging in
425 the lagoon are yearly oversaturated in CO₂ (Polsenaere et al. 2012b) and their positive influence
426 on pCO₂ values was supposed to be maximal in January 2009 (Table 1, Fig. 3). Overall, thermal
427 and non-thermal components of pCO₂ strongly varied seasonally, almost compensating each
428 other. For instance, the six months Δ TpCO₂ offset (260 ppmv) observed from July 2008 to
429 January 2009 (TpCO₂ from 601 to 341 ppmv) concomitantly to the 13.3 °C water temperature
430 decrease (from 22.2 to 8.9 °C) barely compensated non-thermal effects on water pCO₂ that
431 occurred during that six-month period (Δ NpCO₂ = 317 ppmv, from 381 to 698 ppmv) (Fig. 2).
432 In spring and summer, more autotrophic metabolism by all benthic and planktonic primary
433 producers of the lagoon is favored and less remineralization occurs resulting in the lowest
434 NpCO₂ values, while the effect of autotrophy on pCO₂ is offset by water heating. Phytoplankton
435 blooms occur in the lagoon especially from the spring season that shows the highest primary
436 production rates (between 231.4 and 496.6 mg C m⁻³ d⁻¹ in 2003, Glé et al. 2008). Influxes of
437 CO₂ in spring and early autumn were measured on the flat by the EC station (Polsenaere et al.
438 2012a). *Zostera* seagrass meadows also contribute significantly to the total primary production
439 of the Bay (Plus et al. 2015; Polsenaere et al. 2012a; Ribaudou et al. 2017) and then may
440 influence pCO₂ dynamics as observed elsewhere in the eastern shore of Virginia by Berg et al.
441 (2019). In summer in the lagoon, higher *Zostera* seagrass net primary production associated
442 with lower heterotrophic respiration result in more autotrophic sediment metabolism; together
443 with phytoplanktonic and immersed *Zostera* production and lower freshwater inputs to the
444 lagoon explain the NpCO₂ drawdown from January to June 2009 (281, from 698 to 417 ppmv).
445 However, in summer, *Zostera* primary production probably occurs using mostly CO₂ directly
446 from the atmosphere during the emersion and even more if low tide is around midday
447 (Polsenaere et al. 2012a). Thus, in the studied channel the direct influence of *Zostera*

448 metabolism on water pCO₂ is probably modest. To the contrary from spring to summer-autumn
449 2009 periods, NpCO₂ values increased (46, from 417 to 463 ppmv) whereas primary production
450 decreased. In autumn and winter contrarily to summer season, heterotrophy is favored because
451 the seagrass biomass is recycled and decomposed in the sediments and the waters (Auby and
452 Labourg 1996). The *Zostera* meadow decline feeds the lagoon sediments and waters with
453 organic matter, favoring heterotrophy which together with a lower planktonic primary
454 production and also more freshwater inputs to the lagoon, contribute to the highest NpCO₂
455 values measured in November 2008 and January 2009 (Fig. 2). At the same time, this
456 combination of all factors on water pCO₂ is compensated by water cooling and in consequence,
457 a pCO₂ value seasonally constant at about 500 ppmv was measured in the lagoon. Other studies
458 showed a more pronounced effect of either thermal or non-thermal variations on water pCO₂ at
459 the seasonal scale as for instance Ribas-Ribas et al. (2011) that reported in the Bay of Cadiz
460 during winter a dominant thermal control in comparison with non-thermal effects, or to the
461 contrary, Bozec et al. (2011) who showed in the Bay of Brest, a stronger influence of non-
462 thermal (biology) processes on seasonal pCO₂ dynamics (Table 3).

463

464 **4.2. Seasonal carbonate chemistry and mixing patterns of freshwater and seawater DIC**

465 Besides biological activity and its light-temperature controlled seasonal evolution, horizontal
466 advection and exchanges between terrestrial-lagoon-oceanic waters producing different mixing
467 patterns, strongly drives observed carbonate chemistry at the seasonal scale in the lagoon.
468 Hydrodynamic forcing importance was shown in other tidal systems such as a freshwater or
469 salty marshes where intense exchanges between marsh and coastal/estuarine waters occurred
470 during hydrological floods (Neubauer and Anderson 2003, and references in Table 3). Intense
471 exchanges of oceanic and continental waters occurred within the Arcachon lagoon as well
472 through channels or over the tidal flat around high tide (Plus et al. 2009; Polsenaere et al. 2012a,

473 b). A significant influence of freshwater discharge in the lagoon occurred in winter of the year
474 2009 with the lowest salinities (20.5-24.5) associated to the lowest TA concentrations (1.493-
475 1.733 mmol kg⁻¹) measured through our tidal cycle sampling (Fig. 3e). These low values
476 especially at low tide reflected the freshwater inputs to the lagoon from rivers draining the
477 podzolized acidic catchment (Polsenaere et al. 2012b). Furthermore, in the lagoon, water
478 seasonal mean TA values were always above corresponding DIC concentrations (maximum
479 hourly measured DIC values were below minimum TA values too) except in January and April
480 2009. It highlights the seawater influence at the sampled channel station in the middle position
481 inside the lagoon where CO₃²⁻ provides the first line of buffering capacity as observed by Wang
482 et al. (2016) in the Sage Lot Pond salt marsh system. The significant positive correlations
483 between δ¹³C-DIC values and salinity values, TA values and DIC concentrations over all 2019
484 cycles (Fig. 4a, c, d) supports the importance of alkalinity (CO₃²⁻) among other DIC species
485 and thus an overall seasonal DIC variation control through tidal forcing.

486 Comparing apparent 0 end-member TA and *in situ* catchment TA values at each cycle (season),
487 we highlighted that alkalinity production occurred within the lagoon especially in April, July
488 2008 and September 2009 when 0 end-member TA values were at least two times higher than
489 *in situ* catchment weighted mean TA values (Fig. 4b). The Arcachon flat constitutes an
490 important stock of carbonates of about 120 Mt of several shellfish species represented at 95 %
491 by *Crassostrea gigas* (Polsenaere et al. 2012a). From the end of the reproduction season
492 (November) to the spat removing in spring, early development stages of bivalves are
493 particularly sensitive to dissolution-induced mortalities as shown by Barros et al. (2013)
494 through larval viability laboratory experiments on the Pacific oyster *Crassostrea gigas*. Thus,
495 CaCO₃ dissolution could occur in wet mud sediments in presence of such shellfishes patchily
496 distributed on the tidal flat. To the contrary, in late summer (September) and autumn
497 (November) 2008, *in situ* catchment TA values were slightly higher than 0 end-member TA

498 values supporting the possibility of calcification processes in the lagoon (Fig. 4b). Seagrass
499 systems are recognized to be important site of high calcium carbonate (CaCO_3) cycling, both
500 in terms of production and dissolution, such as *Posidonia oceanica* meadows where CaCO_3
501 precipitation can overwhelm sediment CaCO_3 dissolution particularly during net autotrophic
502 periods (Barron et al. 2006). In any cases, net TA production in the lagoon remains more
503 important in comparison with calcification at the annual scale.

504

505 **4.3. Processes controlling tidal and diurnal pCO_2 and DIC variations**

506 *4.3.1. Large tidal DIC variations through current advection*

507 Over intertidal systems as Arcachon, horizontal advection can strongly control water pCO_2 and
508 carbonate chemistry variations at the tidal scale too. Significant negative correlations between
509 pCO_2 and salinity or water height occurred most of the time in each season (except in Sept. and
510 Nov. 08) and for each tidal/diurnal periods. The fact that higher pCO_2 values associated to lower
511 salinity and depth values at low tide (opposite pattern during high tide) were systematically
512 measured irrespective of day or night status supports the significant tidal rhythm control on
513 water pCO_2 in the lagoon. For example, during the June 2009 cycle, pCO_2 and TA values were
514 significantly inversely correlated and respectively negatively and positively correlated to water
515 height, salinity and $\delta^{13}\text{C}$ -DIC values. Significantly higher TA concentrations associated to
516 higher $\delta^{13}\text{C}$ -DIC values (typical oceanic TA and $\delta^{13}\text{C}$ -DIC values) were also measured during
517 high tide periods contrarily to pCO_2 values (Table 1, Fig. 3). These results clearly reflected tidal
518 forcing on carbonate chemistry through ocean exchange with more buffered waters on one hand
519 and riverine inputs with more acidic waters on the other hand. In the Arcachon lagoon,
520 maximum pCO_2 values measured at low tide were well below those measured over other coastal
521 systems as tidal flats nearby and small marsh-estuary or mangrove tidal creeks further (Table
522 3). A 24 hour-cycle performed in May 2006 (not shown) in a shallower creek close to the outlet

523 of a small river North East in the back of the Arcachon lagoon showed the highest pCO₂ values
524 (up to 1300 ppmv, salinity values close to 28) at mid-ebb when the tidal flat was still immersed
525 compared to the low tide period, when the flat was emerged and the channel connected to the
526 river, with associated decreased and lower pCO₂ and salinity values (close to 800 ppmv and
527 22-23, respectively, data not shown). Thus, these measurements suggest another associated
528 process (i.e. tidal pumping, see next 4.3.2. discussion part) along with freshwater mixing that
529 occur together in the Arcachon flat.

530

531 *4.3.2. Tidal porewater pumping and associated water alkalinity production and pCO₂* 532 *variations*

533 During tidal cycles conducted at the channel station, TA (or DIC) concentrations showed
534 overall weak tidal variations (Fig. 3). Porewater pumping over tidal coastal vegetated
535 ecosystems (seagrass meadows, salt marshes and mangroves) has been recognized as an
536 important process in water DIC production (Neubauer and Anderson 2003; Wang et al. 2016;
537 Borges et al. 2003; Maher et al. 2013). Indeed, high alkalinity values can be reached through
538 aerobic respiration coupled to carbonate dissolution but also through organic matter anaerobic
539 remineralization products, i.e. N₂ from net denitrification and reduced sulfur (pyrite burial)
540 from net sulfate reduction (Hu and Cai 2011). In Arcachon, tidal pumping has been highlighted
541 as a significant process controlling benthic phosphorus and iron cycles over seagrasses where
542 below the rhizosphere, Fe(II) can be exposed to a reduced environment and precipitates as FeS
543 and pyrite (Deborde et al. 2008b). De Wit (2008) showed the importance of sulfate-reducing
544 bacteria in *Zostera* seagrass sediments at a tidal flat station very close to our 24-hour cycle
545 station. Moreover, Heijs et al. (1999) showed at the same location that a substantial population
546 of aerobic sulfide-oxidizing bacteria was also present, buffering free sulfide through chemical
547 processes with iron leading to high FeS and pyrite concentrations in sediments. In particular,

548 sulfide oxidation (depending on whether sulfide will be oxidized to sulfate or to sulfur) was
549 estimated at 4694×10^3 or $1174 \times 10^3 \text{ mol cm}^{-3} \text{ day}^{-1}$, whereas previously measured sulfate
550 reduction rates at the same station ranged from 150 to $1300 \text{ nmol cm}^{-3} \text{ day}^{-1}$. These findings
551 along with the quantitative apparent 0 end-member and *in situ* catchment TA value (Fig. 4b)
552 suggest a substantial TA production through tidal porewater pumping along the year. Its
553 influence on TA (or DIC) production in the lagoon could not be clearly seen from our chosen
554 site, which have high salinities and TA values, hiding the terrestrial influence and did not cover
555 the 0 end-member. In any cases, in Arcachon, alkalinity production through tidal pumping is
556 not so strong in comparison with other mentioned mangrove or saltmarsh systems. Maher et al.
557 (2013) measured in a small tidal mangrove creek (Australia) during summer and winter a clear
558 tidal trend for DIC as in the present study (highest and lowest values reached during low and
559 high tides, respectively). However, a $0.736 \text{ mmol kg}^{-1}$ DIC range (between 2.064 and 2.8 mmol
560 kg^{-1}) was measured there when in the Arcachon channel we only measured a tidal DIC variation
561 of $0.216 \text{ mmol kg}^{-1}$ (April 2009) at maximum ($0.116 \text{ mmol kg}^{-1}$ in average over the whole study
562 period). In the Sage Lot Pond tidal marsh, Wang et al. (2016) measured in Summer tidal TA
563 variation ranges close to 0.4 mmol kg^{-1} respectively when at Arcachon a variation of 0.261
564 mmol kg^{-1} at maximum ($0.133 \text{ mmol kg}^{-1}$ in average) was seen. Similarly, according to tidal
565 pCO_2 variations, smaller variations were measured in comparison to other salt marsh-estuary
566 or mangrove creek systems, where stronger tidal pCO_2 variations were observed according to
567 porewater mixing (Table 3).

568

569 ***4.3.3 Significant diurnal pCO_2 patterns linked to biological activity***

570 In subtidal coastal ecosystems, pelagic and benthic biological activities can generate large
571 diurnal water pCO_2 variations. Indeed, in comparison with Arcachon, similar and even higher
572 diurnal pCO_2 ranges were measured in macrophyte meadows (Baltic Sea, Florida and Virginia)

573 or in bay channels (Spain) (Table 3). However, in these subtidal ecosystems, tidal pumping was
574 supposed to be minor; in the Arcachon lagoon, pCO₂ variations occur at the tidal scale through
575 porewater pumping and water masses mixing, these latter being superposed on those occurring
576 at the diurnal scale and linked to solar radiation and light availability.

577 During all sampled cycles, no significant pCO₂ versus Chl-*a* concentration relationships were
578 computed and Chl-*a* concentrations never reached more than 1.9 µg L⁻¹ in average (data not
579 shown). Our successive tidal/diurnal cycle samplings in the central lagoon channel were
580 apparently not able to fully catch these blooms especially in spring and summer. However,
581 interesting diurnal pCO₂ patterns related to biological activity could be detected when
582 comparing nighttime with daytime values for a same salinity range. For instance, in September
583 2008, at low tide in the channel for a salinity range values close to 33, lower water pCO₂ and
584 higher temperature values were measured at daytime (400-525 ppmv, > 20 °C) than at nighttime
585 (550-600 ppmv, close to 19 °C) (Fig. 5e, f). The same pattern was observed in June 2009 as
586 well (Fig. 5m, n). Phytoplankton possibly with resuspended microphytobenthos communities
587 in the channel then appeared to be active at that moment. Savelli et al. (2019) showed precisely
588 that almost half (43 %) of MPB primary production can be resuspended annually over intertidal
589 lagoons and the highest occurs in spring tide at the flood beginning due to high current
590 velocities and low water heights. To the contrary, in April 2009 during flooding, water pCO₂
591 (500-560 ppmv) increased (while S increased and T decreased) whereas during ebbing, water
592 pCO₂ (538-494 ppmv) decreased and was lower (while S decreased and T increased) (Fig. 5k,
593 l). At that particular moment in the channel, organic matter produced by primary production
594 could feed community respiration through incoming coastal waters during flooding. The
595 significant negative correlation computed between pCO₂ and δ¹³C-DIC values is consistent
596 with an organic matter mineralization process that occurred in spring in the lagoon.

597 To assess the biological control on measured pCO₂ patterns in the studied channel, we
598 statistically compared, when possible, night and day pCO₂ values averaged over the same
599 salinity range (i.e. with no significant salinity value difference between day and night periods)
600 according to tide phases for each cycle (Table 4). During flooding, daytime pCO₂ were
601 significantly lower than those nighttime pCO₂ values and to the contrary, during ebbing phase,
602 daytime pCO₂ were significantly higher night time pCO₂ values. At the same time, significant
603 variations in water temperature between day and night times for both flooding and ebbing
604 periods occurred. Water temperature gradients exist and could influence diurnal pCO₂
605 variations. However, they cannot alone explain these variations at our sampling station since
606 differences between the theoretical pCO₂ variations due to diel temperature differences and
607 measured changes in water pCO₂ for the same salinity range are large (Table 4).

608 Overall, these reproducible patterns strongly support the control of pCO₂ by biological activity
609 (primary production of phytoplankton, MPB and seagrasses). Koné et al. (2009) reported pCO₂-
610 oversaturated waters in some lagoons of Ivory Coast (West Africa) behaving as macrotidal
611 estuaries due to net ecosystem heterotrophy and riverine CO₂ rich water inputs. To the contrary
612 pCO₂-undersaturated waters were found in the other lagoons showing permanent hyaline
613 stratification leading to higher phytoplankton biomass. In other systems such as Tampa and
614 Florida bays, pCO₂ diurnal variability was mostly influenced by biological processes such as
615 photosynthesis/respiration of benthic communities and precipitation/dissolution of calcium
616 carbonate respectively (Table 3). Interestingly and contrary to other studied salt marsh and
617 mangrove systems (Table 3) where tidal pumping is predominant, water pCO₂ and DIC
618 variations in the Arcachon lagoon seemed to result from the subtle combination of thermal,
619 water column primary production, and tidal pumping effects along with thermodynamic mixing
620 between seawater and freshwater.

621

622 **4.4. High heterogeneity in water-air CO₂ fluxes over the Arcachon lagoon**

623 ***4.4.1. Turbulence forcing control on water-air CO₂ flux temporal variability***

624 The subsurface waters of the Arcachon lagoon remained always oversaturated and close to the
625 atmospheric equilibrium at both seasonal and diurnal scales. Annual mean air-sea CO₂ fluxes
626 were estimated at 0.27 ± 0.22 , 0.56 ± 0.54 and 0.55 ± 0.22 mmol m⁻² h⁻¹ depending on the gas
627 transfer velocity-wind parameterizations. Water-air CO₂ flux values remained in the lower
628 range of annual values reported for macrotidal estuaries (0.58-8.42 mmol m⁻² h⁻¹) (Borges et al.
629 2005). Similar intertidal systems showed different values with more variations at both diurnal
630 and seasonal scales (Table 3). In other subtidal and marine systems, fluctuations between CO₂
631 sink and source could also be measured over the year depending on seasons (Table 3).

632 In the Arcachon lagoon, wind speed influence on calculated air-sea CO₂ fluxes was significant
633 and could be more important in determining the intensity of the flux, than the pCO₂ air-sea
634 gradient itself. Despite significant pCO₂ variations generally noticed at seasonal and
635 tidal/diurnal scales, only few significant air-sea CO₂ flux differences were observed (i.e.
636 between November and September 2008 and between September 2008 and 2009),
637 concomitantly to significant differences in U₁₀ and K₆₀₀ values (Table 2). Thus, air-sea CO₂
638 flux variability in the Arcachon lagoon at all temporal scales appears to be controlled by
639 turbulence (K₆₀₀, physical forcing) rather than air-water pCO₂ gradients. This finding is the
640 opposite of what it is generally observed over dynamic coastal bays at temperate and tropical
641 latitudes characterized by higher diurnal/tidal water pCO₂ gradients (Table 3).

642

643 ***4.4.2. High heterogeneity in water-air CO₂ fluxes according to space and methodology***

644 CO₂ fluxes were estimated at the water-air interface based on k₆₀₀ parameterizations from
645 Wanninkhof et al. (1992), Raymond and Cole (2001) and Abril et al. (2009). These
646 parameterizations were obtained over oceanic and estuarine systems with different approaches,

647 i.e. the bomb ^{14}C inventory in the ocean, non-intrusive tracer data and chamber measurements
648 respectively. The Abril et al. (2009) relationship gave significantly higher flux values compared
649 to the two others, potentially due to bottom current contributions from 2 to 20 cm s^{-1} according
650 to seasons and associated hydrodynamic modelling estimations. In July of the year 2008,
651 significant differences were computed between simultaneous direct EC fluxes and estimated
652 fluxes from gas transfer parameterizations (Table 2). No k_{600} calculations or flux data
653 comparisons were attempted from these simultaneous EC flux and water pCO_2 measurements
654 due to strong spatial heterogeneity in water bodies (separated from 1.8 km) highlighted in the
655 present study and due to the strong variability in gas transfer velocity and CO_2 flux calculations
656 according to methodologies too (Polsenaere et al. 2012a; Raymond and Cole 2001; Vachon et
657 al. 2010).

658 Indeed, variability in water pCO_2 and air-sea CO_2 fluxes depends on site locations inside a same
659 coastal ecosystem at the different temporal scales. The spring 2006 cycle carried out in the back
660 of the Arcachon lagoon (4.3.1. discussion part) showed higher water pCO_2 tidal variations (680-
661 1330 ppmv) and resulted in a higher CO_2 source to the atmosphere at this more terrestrial
662 influenced location. These observations suggest that large differences in water pCO_2 and sea-
663 air CO_2 fluxes could occur at smaller spatial scale due to specific processes and characteristics
664 that exist at the sampled area. In the lagoon, benthic primary producers such as *Zostera noltei*
665 seagrass meadows and resuspended microphytobenthic communities could influence in a
666 different way carbon dynamics and air-sea fluxes at the channel station sampled here, compared
667 to shallow waters above the tidal flat as revealed by EC measurements (Polsenaere et al.
668 2012a). Indeed, a CO_2 influx was detected during the emersion by the EC, but we never
669 observed CO_2 undersaturated waters in the channel during the July 2008 cycle. Due to the
670 strong influence of tidal advection on water pCO_2 in the intertidal lagoon, water masses sampled
671 at the studied station were not the same as those caught in the EC footprint during in July 2008

672 (Tables 1 and 2, Polsenaere et al. 2012a). This could also explain the significant differences in
673 air-sea CO₂ fluxes and associated gas transfer velocities we got from both methodologies, and
674 methodological and logistical complexity to get concomitant water-air EC flux and equilibrator
675 water pCO₂ values from the same water mass in the lagoon. Wang et al. (2018) over the Duplin
676 River salt marsh-estuary (Georgia), though different from Arcachon due to the most of the time
677 emersion canopy (ecosystem-atmosphere flux there versus water-atmosphere flux based on
678 water pCO₂ here) specifically suggested EC measurements of salt marsh net ecosystem
679 exchange could underestimate net ecosystem production for not accounting for lateral DIC
680 exchanges with tidal inundation.

681

682 **Conclusions**

683 Waters of the Arcachon lagoon represents a permanent weak CO₂ supersaturation characterized
684 by a seasonal compensation of thermal and biological effects and carbonate chemistry and
685 mixing patterns of freshwater and seawater DIC. At smaller time scales if large tidal DIC
686 variations through current advection were observed, weak tidal porewater pumping and
687 associated water alkalinity production and pCO₂ variations were noticed compared to other
688 tidal creeks surrounded by mangrove or saltmarshes. Accurate water monitoring such as that
689 carried out in this study permits high resolution analysis of the carbon signal; small but
690 significant pCO₂ variations were seen at the diurnal scale (between 5 and 24 ppmv) linked to
691 the light cycle apparently induced by planktonic and benthic productivity in the channel along
692 to day versus night temperature gradients at flooding and ebbing periods. The temporal
693 variability and the high heterogeneity in computed water-air CO₂ fluxes according to space
694 (water masses) and methodology (EC, ...) would require more integrated carbon fluxes and
695 processes over tidal ecosystems as here for the Arcachon lagoon.

696

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702

703 **Competing Interests**

704 The authors declare no competing interests.

705

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995 **Figure/Table captions**

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997 **Table 1** Inorganic carbon parameters measured during eight 24 hour-cycles to six 24h and two
998 12h cycles in the Arcachon flat (44°42.400'N 01°07.500'W) (number of values, mean ±
999 standard deviation in bold and range between brackets). T and S water temperature and salinity,
1000 δ¹³C-DIC dissolved inorganic carbon isotopic ratio, TA total alkalinity, pCO₂ partial pressure
1001 of CO₂, and DIC dissolved inorganic carbon. DIC were estimated from measured salinity,
1002 temperature, TA and pCO₂ values using the CO₂ system calculation program (Lewis and
1003 Wallace 1998) parameterized with the carbonic acid constants sets proposed by Mehrbach et
1004 al. (1973) refitted by Dickson and Millero (1987), the borate acidity constant from Lee et al.
1005 (2010) and the CO₂ solubility coefficient of Weiss (1974). 15/04/2008 10:00:00 (TU) –
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1007 11:00, 06/11/2008 08:00 – 06/11/2008 17:00, 29/01/2009 06:00 – 29/01/2009 14:00,
1008 15/04/2009 13:00 – 16/04/2009 13:00, 25/06/2009 06:00 – 26/06/2009 06:00 and 04/09/2009
1009 05:00 – 05/09/2009 05:00. Notice that November 2008 and January 2009 were 12 hour-cycles

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1011 **Table 2** Gas transfer velocities (k_{600}) and air-sea CO₂ fluxes computed at high tide day and
1012 night periods (Fig. 3, the four-hour emerged-periods around low tide at each cycle were
1013 removed) for each cycle in the Arcachon flat (mean ± standard deviation and range between

1014 brackets). U_{10} values were computed from wind speed values measured at 9 meters high by the
1015 Lège-Cap Ferret Meteo-France station, 12 km far from the 24-H site) using the Amorochó and
1016 DeVries (1980) equation. k_{600} values were estimated according to Wanninkhof et al. (1992),
1017 Raymond & Cole (2001) and Abril et al. (2009) equations (W92, RC01 and A09, respectively,
1018 see M&M). Air-sea CO_2 fluxes were then calculated from k_{600} , water and air pCO_2 values. A
1019 mean air pCO_2 value of 390 ppm was chosen according to Eddy Covariance (EC) measurements
1020 deployed at 1.8 km far from the tidal creek in 2008 and 2009 (Polsenaere et al. 2012a). For
1021 comparison, U_{10} and air-sea CO_2 flux values obtained in July 2008 from simultaneous EC
1022 measurements (Polsenaere et al. 2012a) in July 2008 are shown in red. No k_{600} calculations
1023 were attempted from these simultaneous EC flux and water pCO_2 measurements at this season
1024 due to strong spatial heterogeneity in water bodies highlighted in the present study (see
1025 Discussion section 4.4.2.)

1026

1027 **Table 3** Seasonal, tidal and diurnal pCO_2 and water-atmosphere variation comparisons across
1028 tidal flat, bay, marsh-estuary and mangrove systems of the coastal zone (^adiurnal and/or tidal,
1029 ^bspatial and/or longitudinal sampling methodology)

1030

1031 **Table 4** Comparison between day and night pCO_2 values averaged over a same salinity range
1032 according to tidal phases (flooding/ebbing) for each 2008-2009 cycle. The non-parametric
1033 Mann-Whitney (p-values: *** = 0.0002; **** < 0.0001) test was used. / symbol indicates
1034 cycles (flooding and/or ebbing periods) where comparisons were not possible due to not enough
1035 data (no data > 5 min. with no significant salinity value difference between day and night
1036 periods). To the contrary, time periods (min.) (> 5 min.) where comparisons were possible are
1037 specified for each cycle. Temperature comparisons are also given along with the pCO_2 increase
1038 or decrease only due to day through nighttime temperature difference (calculated with the CO_2

1039 System Calculation program (version 2.1.) (Lewis and Wallace 1998) from constant T, S, TA
1040 and DIC values measured during each cycle (see Table 1 and M&M section)

1041

1042 **Fig. 1** Localization of the 24 hour-cycle site. (a) The Arcachon lagoon with the subtidal zone
1043 (channels and creeks) and the intertidal mudflat area; (b) the 24 hour-cycle site (24H,
1044 44°42.400'N 01°07.500'W in the subtidal creek always immersed); the EC station (EC,
1045 44°42.9858'N 01°08.6160'W, 1.8 km far from the 24-H site, in the intertidal area emerged for
1046 approximately four hours and immersed for approximately nine hours, Polsenaere et al. 2012a);
1047 the meteo-france station (A) (MF, 44°37.900'N 01°14.900'W, 12.450 km far from the 24-H
1048 site). The *Zostera noltei* seagrass meadow is derived from the SPOT image of the 22 June 2005;
1049 it occupies 60 % of the intertidal area (shades of green show the differences in seagrass density,
1050 brown and yellow represent muddy and sandy areas)

1051

1052 **Fig. 2** *In situ* water temperatures (dashed black curve), pCO₂ (black curve with black dots) and
1053 derived temperature-normalized pCO₂ (NpCO₂, i.e. pCO₂ variations due to biological activity
1054 or non-temperature effects, black dotted curve with empty black dots, annual mean temperature
1055 over 2008/09 17.7 ± 4.2°C, dotted line) and thermally forced seasonal pCO₂ according to
1056 Takahashi et al. (2002) (TpCO₂, i.e. pCO₂ variations due to thermal effects, grey dotted curve
1057 with grey dots, annual mean temperature over 2008/09, 17.7 ± 4.2°C, dotted line and pCO₂ 496
1058 ± 36 ppmv, dashed line); mean temperature and pCO₂ over the whole 2008 and 2009 period
1059 were chosen for NpCO₂ and TpCO₂ computations due to sampled cycle strategy and for better
1060 consistency in year comparisons. Minimum and maximum pCO₂ values (associated to the mean
1061 pCO₂ value) measured during each tidal cycle are also represented (vertical black lines)

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1063 **Fig. 3** Tidal and diurnal variations in inorganic carbon and associated parameters measured
1064 during eight 24 hour-cycles to six 24h and two 12h cycles in the Arcachon flat. Water height
1065 (H), salinity (S), temperature (T), partial pressure of CO₂ in the water (pCO₂), total alkalinity
1066 (TA), dissolved inorganic carbon (DIC) and dissolved inorganic carbon isotopic ratio ($\delta^{13}\text{C}$ -
1067 DIC); (a) 15/04/2008 10:00:00 (TU) – 16/04/2008 10:00:00, (b) 02/07/2008 10:00 – 03/07/2008
1068 10:00, (c) 18/09/2008 11:00 – 19/09/2008 11:00, (d) 06/11/2008 08:00 – 06/11/2008 17:00, (e)
1069 29/01/2009 06:00 – 29/01/2009 14:00, (f) 15/04/2009 13:00 – 16/04/2009 13:00, (g)
1070 25/06/2009 06:00 – 26/06/2009 06:00 and (h) 04/09/2009 05:00 – 05/09/2009 05:00. T and S
1071 were measured every minute by the YSI multiparameter probe on board and H was measured
1072 every hour by the SHOM. The other parameters were sampled on board every hour. Clear and
1073 grey areas represent daytime and nighttime periods, respectively. Notice the same y-axis scale
1074 between cycles was chosen to better see seasonal variations and that November 2008 and
1075 January 2009 were 12 hour-cycles

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1077 **Fig. 4** Cross-correlation plots of TA (a), DIC (c) and $\delta^{13}\text{C}$ -DIC (d) values versus salinity and
1078 apparent 0 end-member versus *in situ* catchment TA value comparison (b). TA versus S slopes
1079 are all significantly positive, DIC versus S slopes are significantly positive except in September
1080 2009 (negative); and $\delta^{13}\text{C}$ -DIC versus S slopes are positive and significant except in April 2008,
1081 November 2008, September 2009. (b) Apparent 0 end-member TA and *in situ* catchment TA
1082 correspond to Y-intercept of TA versus S cross-correlations and *in situ* TA measurements
1083 carried out at the same time over the Arcachon lagoon watershed watercourses (see Polsenaere
1084 et al. 2012b) computed here as discharge-weighted TA means, respectively

1085

1086 **Fig. 5** Diurnal/tidal plots of water temperature and pCO₂ versus salinity for the four periods:
1087 HT_N high tide at night (dark blue triangle), LT_N low tide at night (green inversed triangle),

1088 LT_D low tide at day (yellow inversed triangle) and HT_D high tide at day (clear blue triangle).
 1089 (a, b) April 2008, (c, d) July 2008, (e, f) September 2008, (g, h) November 2008, (I, j) January
 1090 2009, (k, l) April 2009, (m, n) June 2009 and (o, p) September 2009. Notice the same y-axis
 1091 pCO₂ scale between all cycles was chosen to better observe diurnal and tidal variations. The
 1092 same x-axis temperature and salinity scales between April 2008-2009, July 2008-June 2009 and
 1093 September 2008-2009 were also chosen to better observe variations during the same seasons
 1094 from 2008 to 2009

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Table 1

	<i>T</i> (°C)	<i>S</i>	δ ¹³ C-DIC (‰)	TA (mmol kg ⁻¹)	pCO ₂ (ppmv)	DIC (mmol kg ⁻¹)
<i>April 2008</i>	1441	1441	23	25	1441	25
	13.9 ± 0.6	29.2 ± 0.6	-1.1 ± 1.1	2.081 ± 0.028	474 ± 14	1.948 ± 0.024
	(12.8 ~ 15.4)	(26.2 ~ 30.1)	(-3.7 ~ -0.4)	(2.000 ~ 2.119)	(442 ~ 496)	(1.884 ~ 1.985)
<i>July 2008</i>	1464	1464	24	25	1434	25
	22.2 ± 0.6	31.2 ± 0.6	-0.5 ± 0.4	2.131 ± 0.035	461 ± 14	1.926 ± 0.028
	(21.2 ~ 23.4)	(30.2 ~ 32.6)	(-1.7 ~ 0.2)	(2.072 ~ 2.211)	(432 ~ 499)	(1.879 ~ 1.995)
<i>September 2008</i>	1428	1428	25	25	1436	25
	19.2 ± 0.2	33.6 ± 0.3	-0.6 ± 0.3	2.255 ± 0.021	515 ± 36	2.062 ± 0.017
	(18.8 ~ 20.2)	(33.1 ~ 34.5)	(-1.5 ~ -0.1)	(2.223 ~ 2.301)	(405 ~ 597)	(2.017 ~ 2.086)
<i>November 2008</i>	541	541	10	10	541	10
	12.3 ± 0.2	32.2 ± 0.3	-0.6 ± 0.3	2.202 ± 0.020	463 ± 19	2.050 ± 0.022
	(12.0 ~ 12.6)	(31.8 ~ 32.6)	(-1.3 ~ -0.3)	(2.180 ~ 2.234)	(419 ~ 497)	(2.021 ~ 2.081)
<i>January 2009</i>	490	480	9	9	480	9
	8.9 ± 0.4	23.2 ± 1.5	-1.0 ± 0.5	1.646 ± 0.086	480 ± 15	1.594 ± 0.076
	(8.3 ~ 10.3)	(20.4 ~ 25.6)	(-2.1 ~ -0.5)	(1.493 ~ 1.733)	(452 ~ 503)	(1.458 ~ 1.669)
<i>April 2009</i>	1518	1518	25	25	1518	25
	13.7 ± 0.3	30.1 ± 1.0	-0.7 ± 0.1	2.006 ± 0.053	525 ± 14	1.891 ± 0.045
	(12.7 ~ 14.5)	(24.8 ~ 31.5)	(-0.9 ~ -0.5)	(1.840 ~ 2.101)	(494 ~ 568)	(1.752 ~ 1.968)
<i>June 2009</i>	1441	1441	24	25	1441	25

	21.5 ± 0.6 (20.0 ~ 23.3)	31.0 ± 0.4 (30.2 ~ 32.2)	-0.2 ± 0.2 (-0.4 ~ 0.2)	2.099 ± 0.029 (2.044 ~ 2.156)	490 ± 27 (419 ~ 545)	1.916 ± 0.023 (1.866 ~ 1.951)
<i>September 2009</i>	<i>1441</i>	<i>1440</i>	<i>25</i>	<i>25</i>	<i>1435</i>	<i>25</i>
	20.9 ± 0.3 (20.2 ~ 21.5)	34.0 ± 0.3 (33.5 ~ 34.8)	-0.2 ± 0.3 (-0.90 ~ 0.4)	2.218 ± 0.020 (2.187 ~ 2.251)	530 ± 39 (453 ~ 601)	2.021 ± 0.021 (1.978 ~ 2.051)
<i>2008-2009 Average</i>	<i>9764</i>	<i>9753</i>	<i>165</i>	<i>169</i>	<i>9726</i>	<i>169</i>
	17.7 ± 4.2 (8.3 ~ 23.4)	30.5 ± 3.4 (20.4 ~ 34.8)	-0.6 ± 0.4 (-3.7 ~ 0.4)	2.079 ± 0.036 (1.493 ~ 2.301)	496 ± 36 (405 ~ 601)	1.926 ± 0.032 (1.458 ~ 2.086)

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1109 **Table 2**

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		U_{10} (m s ⁻¹)	K_{600} (cm h ⁻¹)	Air-Sea CO ₂ fluxes (mmol m ⁻² h ⁻¹)				
			W92	RC01	A09	W92	RC01	A09
Apr. 08	<i>Day</i>	4.42 ± 1.42 (2.02 ~ 6.06)	6.59 ± 3.71 (1.26 ~ 11.37)	9.9 ± 4.46 (3.87 ~ 15.91)	17.04 ± 4.92 (8.76 ~ 22.74)	0.19 ± 0.1 (0.05 ~ 0.34)	0.29 ± 0.12 (0.14 ~ 0.48)	0.5 ± 0.12 (0.32 ~ 0.69)
	<i>Night</i>	2.88 ± 0.7 (2.02 ~ 4.04)	2.71 ± 1.27 (1.26 ~ 5.05)	5.38 ± 1.34 (3.87 ~ 7.85)	11.74 ± 2.4 (8.76 ~ 15.72)	0.07 ± 0.04 (0.02 ~ 0.14)	0.13 ± 0.05 (0.07 ~ 0.22)	0.29 ± 0.11 (0.16 ~ 0.44)
Jul. 08	<i>Day</i>	4.46 ± 2.34 (2.02 ~ 11.1)	7.71 ± 9.88 (1.26 ~ 38.21)	14.92 ± 24.73 (3.87 ~ 93.05)	17.16 ± 8.06 (8.69 ~ 40.09)	0.18 ± 0.23 (0.03 ~ 0.9)	0.36 ± 0.58 (0.09 ~ 2.19)	0.41 ± 0.2 (0.2 ~ 0.94)
	<i>Night</i>	5.05 ± 1.43 (3.03 ~ 6.06)	8.37 ± 4.03 (2.84 ~ 11.37)	12.12 ± 4.94 (5.51 ~ 15.91)	19.2 ± 4.91 (12.24 ~ 22.68)	0.21 ± 0.11 (0.07 ~ 0.31)	0.31 ± 0.13 (0.14 ~ 0.43)	0.49 ± 0.13 (0.32 ~ 0.61)
EC data Tidal flat	<i>Day</i>	3.86 ± 1.29 (1.29 ~ 7.71)				-2.25 ± 2.7 (-12.38 ~ 2.1)		
	<i>Night</i>	4.3 ± 2.21 (0.69 ~ 10.14)				0.12 ± 1.37 (-2.56 ~ 1.76)		
Sep. 08	<i>Day</i>	2.48 ± 0.83 (1.01 ~ 4.04)	2.1 ± 1.33 (0.32 ~ 5.05)	4.72 ± 1.41 (2.72 ~ 7.85)	10.32 ± 2.86 (5.24 ~ 15.73)	0.08 ± 0.05 (0.01 ~ 0.16)	0.19 ± 0.06 (0.1 ~ 0.3)	0.42 ± 0.15 (0.2 ~ 0.68)
	<i>Night</i>	2.31 ± 1.39 (1.01 ~ 5.05)	2.17 ± 2.66 (0.32 ~ 7.9)	4.82 ± 2.95 (2.72 ~ 11.17)	9.76 ± 4.81 (5.29 ~ 19.22)	0.08 ± 0.11 (0.01 ~ 0.32)	0.19 ± 0.12 (0.09 ~ 0.45)	0.38 ± 0.19 (0.2 ~ 0.78)
Nov. 08	<i>Day</i>	7.74 ± 1.77 (5.05 ~ 10.09)	19.37 ± 8.37 (7.9 ~ 31.58)	33.33 ± 19.51 (11.17 ~ 65.36)	28.38 ± 6.12 (19.08 ~ 36.54)	0.53 ± 0.22 (0.25 ~ 0.87)	0.92 ± 0.52 (0.36 ~ 1.81)	0.79 ± 0.14 (0.61 ~ 1.01)
	<i>Night</i>							
Jan. 09	<i>Day</i>	2.27 ± 0.97 (1.01 ~ 3.03)	1.82 ± 1.25 (0.32 ~ 2.84)	4.4 ± 1.36 (2.72 ~ 5.51)	9.58 ± 3.3 (5.27 ~ 12.19)	0.06 ± 0.04 (0.01 ~ 0.11)	0.14 ± 0.06 (0.08 ~ 0.21)	0.29 ± 0.13 (0.16 ~ 0.46)
	<i>Night</i>							
Apr. 09	<i>Day</i>	5.35 ± 3.37 (1.01 ~ 10.09)	12.03 ± 12.55 (0.32 ~ 31.58)	22.84 ± 25.61 (2.72 ~ 65.36)	20.02 ± 11.43 (5.24 ~ 36.09)	0.56 ± 0.59 (0.01 ~ 1.59)	1.06 ± 1.2 (0.12 ~ 3.3)	0.93 ± 0.55 (0.23 ~ 1.82)
	<i>Night</i>	4.04 ± 1.81 (2.02 ~ 7.07)	5.9 ± 5.23 (1.26 ~ 15.48)	9.43 ± 6.95 (3.87 ~ 22.65)	15.58 ± 6.18 (8.65 ~ 25.9)	0.23 ± 0.2 (0.06 ~ 0.6)	0.38 ± 0.26 (0.18 ~ 0.87)	0.63 ± 0.21 (0.4 ~ 1)
Jun. 09	<i>Day</i>	3.62 ± 1.32 (1.01 ~ 5.05)	4.55 ± 2.65 (0.32 ~ 7.9)	7.39 ± 2.92 (2.72 ~ 11.17)	14.18 ± 4.51 (5.25 ~ 19.06)	0.13 ± 0.08 (0.01 ~ 0.25)	0.22 ± 0.09 (0.09 ~ 0.35)	0.43 ± 0.14 (0.17 ~ 0.6)
	<i>Night</i>	3.7 ± 1.17 (3.03 ~ 5.05)	4.53 ± 2.92 (2.84 ~ 7.9)	7.4 ± 3.27 (5.51 ~ 11.17)	14.54 ± 4.02 (12.21 ~ 19.18)	0.12 ± 0.08 (0.07 ~ 0.22)	0.2 ± 0.09 (0.13 ~ 0.31)	0.4 ± 0.12 (0.29 ~ 0.53)
Sep. 09	<i>Day</i>	9.31 ± 3.1 (5.05 ~ 13.12)	29.51 ± 18.64 (7.9 ~ 53.37)	81.73 ± 80.78 (11.17 ~ 188.61)	32.09 ± 10.07 (19.18 ~ 46.85)	1.02 ± 0.65 (0.23 ~ 1.83)	2.8 ± 2.68 (0.32 ~ 6.48)	1.11 ± 0.47 (0.48 ~ 1.65)
	<i>Night</i>	2.88 ± 0.91 (2.02 ~ 4.04)	2.8 ± 1.7 (1.26 ~ 5.05)	5.48 ± 1.78 (3.87 ~ 7.85)	11.71 ± 3.14 (8.71 ~ 15.75)	0.11 ± 0.06 (0.05 ~ 0.21)	0.23 ± 0.06 (0.14 ~ 0.33)	0.49 ± 0.11 (0.32 ~ 0.66)

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Table 3

Study Site	Seasonal pCO ₂ range (ppmv)	Diurnal-tidal pCO ₂ range (ppmv)	Water-atmosphere CO ₂ flux (mmol m ⁻² h ⁻¹)	Reference
Arcachon tidal flat (France)	461 ± 14 (July 2008) – 530 ± 39 (September 2009)	(452 - 503) (January 2009) to (405 - 597) (September 2008)	0.27 ± 0.22 to 0.56 ± 0.54	This study ^a
Aiguillon tidal flat (France)	(390 - 609) (spring - summer 2017) (215 - 1929) (summer - fall 2018)		0.14 (-0.075 to 1.13, summer - winter)	Polsenaere et al. (2018) ^b ; Coignot et al. (2020; in prep.) ^b
Aiguillon tidal flat (France)		(466 - 1024) (HT-LT) (summer 2018)	1.00 ± 0.82	Ternon et al. (2018) ^a
Duplin River salt marsh-estuary (Georgia, USA)		(500 - 4000) (HT-LT, winter) to (1600 - 12,000) (HT-LT, summer) (2014)	-0.7 to -5.5 and -0.6 to -3.9	Wang et al. (2018) ^{a,b}
Tidal marsh channel of the Fier d'Ars (Ré Island, France)	385 ± 60 (267 - 522) (summer) to 460 ± 58 (334 - 569) (autumn)	(377 - 510) (HT-LT, winter 2018) to (334 - 569) (HT - LT, Autumn 2018)	-0.18 ± 0.18 (spring) to 0.10 ± 0.08 (winter)	Mayen et al. (in prep.) ^a
Amazon estuary - tidal forest channel		5000 ± 300 - 2320 ± 40 (LT - HT, February 2007)		Abril et al. (2013) ^a
Ras Dege (Tanzania) mangrove creek		500 - 5000 (September 2005)	0.04-3.33	Bouillon et al. (2007) ^a
Nagada (Papua New Guinea) and the Gaderu (India) mangrove creeks		(540 - 1680) and (1380 - 4770) (July, August 2000)	1.82 ± 1.38 and 2.33 ± 4.20	Borges et al. (2003) ^a
Bay of Brest (subtidal)	200 - 700 (summer-winter, 2013 to 2010)		-0.0083 to -0.38 (spring and summer) and 0.038 to 0.91 (fall and winter)	Bozec et al. (2011) ^a
Estuary of Guadalquivir	341 ± 18 (February 2007) - 411 ± 20 (November 2006)		-0.058 (February 2007) and 0.15 (November 2006)	Ribas-Ribas et al. (2011) ^b
Estuary and the Bay of Cadiz		(589 - 1244), (525 - 1405) and (742 - 1059) (HT-LT, summer 2015)	0.375 to 2.37	Burgos et al. (2018) ^{a,b}
Coastal systems of Cadiz Bay (Spain)				
Jiaozhou Bay (East China Sea)	315 - 720 (autumn 2007) and 145 - 315 (winter 2008)		0.12 (autumn) and -0.68 (winter)	Zhang et al. (2012) ^b
Tampa and Florida bays (US)	414 ± 86 (October 2003); 351 ± 72 (March 2000)	(262 - 580) (October 2003); (260 - 497) (March 2000)		Yates et al. (2007) ^a
Macrophyte meadow (Baltic Sea)		281 ± 88 (July); 219 ± 24 (August); 1488 ± 574 (September 2011)		Saderne et al. (2013) ^a
<i>Zostera marina</i> meadow, Eastern shore of Virginia (US)	425 (April) - 490 (June 2015)	(193 - 731) (April); (256 - 859) (June 2015)		Berg et al. (2019) ^a
Guanabara Bay (Brazil)	353 ± 141 and 194 ± 127 (S3); 380 ± 286 and 203 ± 159 (S4); 364 ± 343 and 132 ±	591 ± 231 to 194 ± 114 (September 2013); 163 ± 40 to 116 ± 25 (January 2014); 346 ± 166 to 146	-2.87 (S3); -2.16 (S4); -2.52 (S5)	Cotovicz et al. (2015) ^{a,b}

74 (S5) (winter and summer) ± 106 (February 2014);
 637 ± 421 to 265 ± 186
 (April 2014) (S4, S5)

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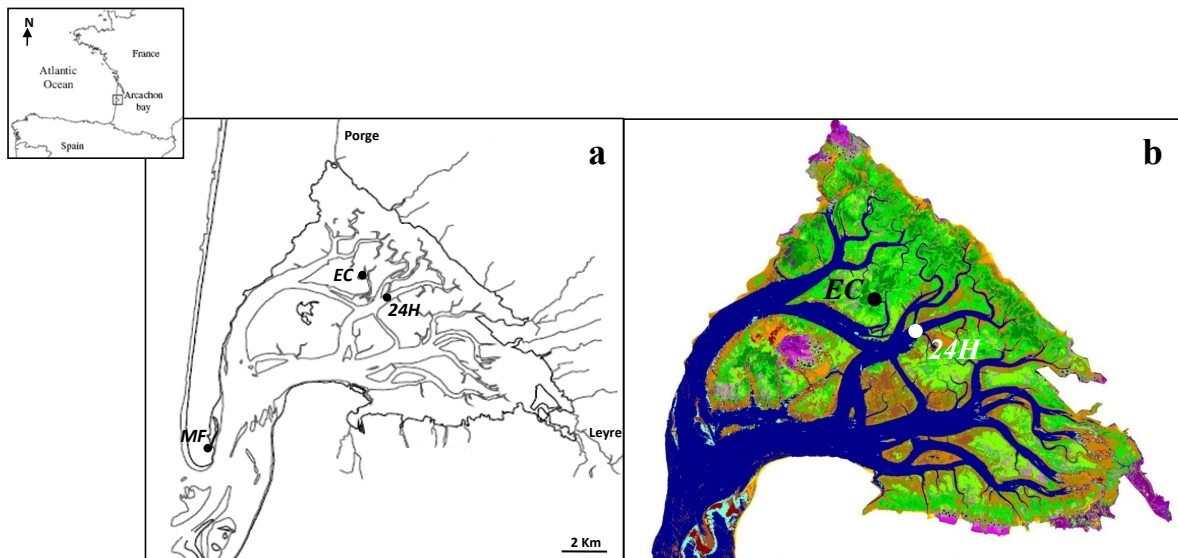
Table 4

		Day			Night			Significant differences (Day vs Night)			pCO ₂ increase or decrease only due to day through night T difference
		Salinity	pCO ₂ (ppmv)	T (°C)	Salinity	pCO ₂ (ppmv)	T (°C)	Salinity	pCO ₂	T	
<i>Apr. 2008</i>	Flooding	/	/	/	/	/	/	/	/	/	/
	Ebbing (30 min.)	29.27	481	14.4 7	29.27	471	13.5 6	No	Yes ****	Yes ****	-18
<i>Jul. 2008</i>	Flooding	/	/	/	/	/	/	/	/	/	/
	Ebbing (10 min.)	30.4	493	21.3 5	30.43	471	22.5 5	No	Yes ****	Yes ****	+21
<i>Sept. 2008</i>	Flooding (120 min.)	33.55	513	18.9 5	33.55	537	19.0 7	No	Yes ****	Yes ****	+2
	Ebbing	/	/	/	/	/	/	/	/	/	/
<i>Nov. 2008</i>	Flooding	/	/	/	/	/	/	/	/	/	/
	Ebbing	/	/	/	/	/	/	/	/	/	/
<i>Jan. 2009</i>	Flooding	/	/	/	/	/	/	/	/	/	/
	Ebbing	/	/	/	/	/	/	/	/	/	/
<i>Apr. 2009</i>	Flooding	/	/	/	/	/	/	/	/	/	/
	Ebbing	/	/	/	/	/	/	/	/	/	/
<i>Jun. 2009</i>	Flooding	/	/	/	/	/	/	/	/	/	/
	Ebbing (30 min.)	33.89	552	20.9 7	33.9	572	20.5 8	No	Yes ***	Yes ****	-8
	Ebbing (15 min.)	33.97	540	20.9 7	33.96	535	20.8 4	No	Yes ***	Yes ****	-3

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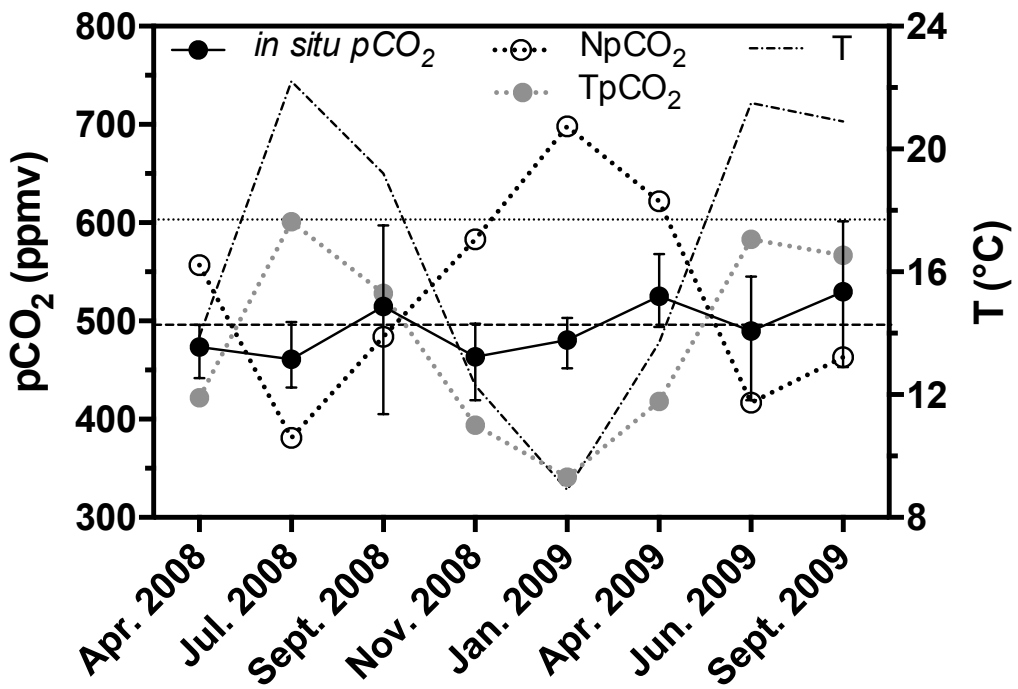
Fig. 1



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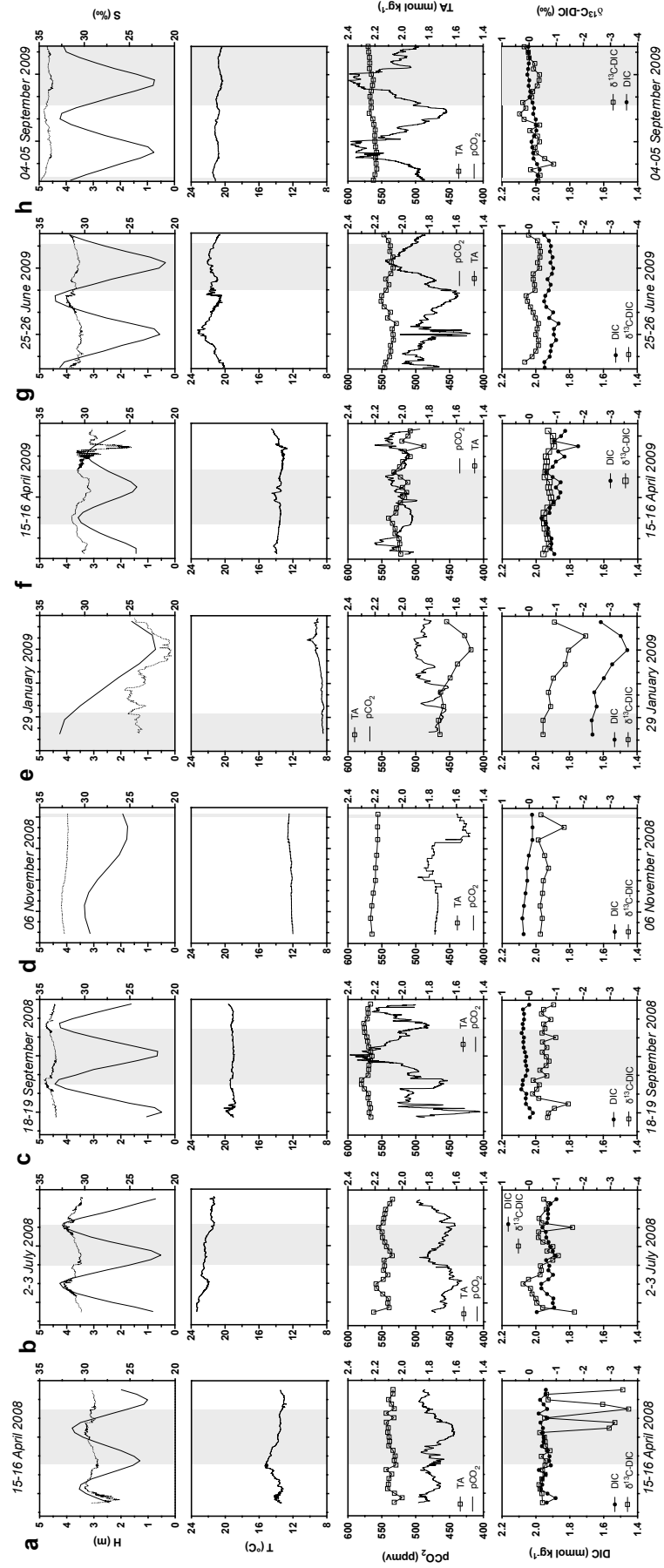
Fig. 2



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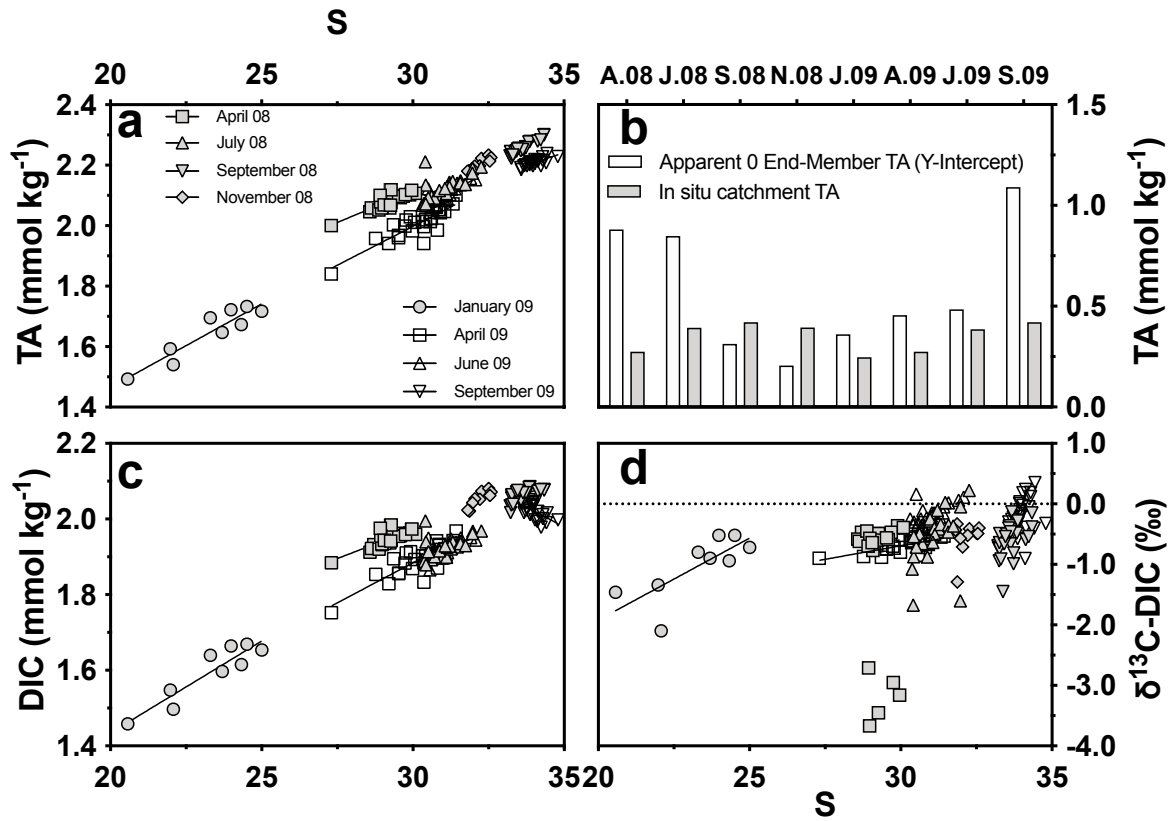
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Fig. 3



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Fig. 4



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Fig. 5

