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Redox interaction between selenite and mackinawite in

2	cement pore water
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Abstract

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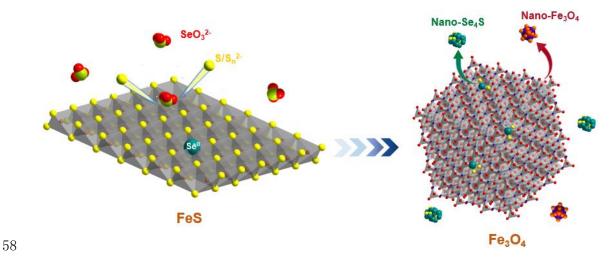
In cement-rich radioactive waste repositories, mackinawite (FeS) could form at the steel corrosion interface within reinforced concrete and potentially retard the transport of redox-sensitive radionuclides (e.g., ⁷⁹Se) in cement porous media. Redox interactions between selenite and mackinawite in hyperalkaline conditions remain unclear and require further investigations. Here, using a comprehensive characterization on both aqueous and solid speciation, we successfully monitored the whole interaction process between selenite and mackinawite in hyperalkaline conditions. Results show similar chemical environments for SeO_3^{2-} and S^{2-}/S_n^{-2-} at the mackinawite-water interface, verifying an immediate reduction. After 192 h reaction, SeO_3^{2-} was reduced to Se^0 and SeS_2 , accompanied by the oxidation of S^{2-}/S_n^{2-} to $S_2O_3^{2-}$ and Fe(II) to Fe(III) in mackinawite. Aqueous results showed that ~99 % of aqueous selenium was present as Se₄S nano-particles due to the dissolution of Se from the solid. In parallel, $\sim 62\%$ of S^{2-}/S_n^{2-} was released into the solution, with mackinawite transforming to magnetite, Fe(OH)₃ and FeS₂O₃⁺ complexed to Cl⁻ or OH species, and magnetite subsequently dispersed in the solution. This study provides valuable data about the retardation mechanisms of redox-sensitive radionuclides by soluble iron sulfides, which is critical to advance our understanding on reactive concrete barriers used in nuclear waste disposal systems.

- Synopsis: The comprehensive investigation of the fate of Se, S and Fe under hyperalkaline conditions promotes a better description of cement-rich radioactive waste repository environments.
- Key words: mackinawite, selenite, cement pore water, interaction mechanism, reactant fate

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Table of Contents (TOC)



1. Introduction

The trace element selenium (Se) is essential for human health, but highly toxic when exceeding the appropriate intake range. ¹⁷⁹Se is one of the ²³⁵U fission products, with a long half-life of 4.8 × 10⁵ years. Its safe disposal is a requisite to the sustainable development of nuclear industry. Oxidized species of selenium are highly mobile, which has stimulated the research on selenium interactions with potential anthropogenic or natural barriers such as clays²⁻⁴, cements⁵⁻⁷ or Fe phases^{4,8-12}. The various oxidation states of selenium (-II, -I, 0, +IV, +VI) and their complexation ability with inorganic and organic matter make selenium chemistry complex. Under oxidizing conditions, Se often occurs as highly soluble and thus mobile oxyanionic forms, i.e., SeO₃²⁻ and SeO₄²⁻. In contrast, under reducing conditions selenium adopts oxidation states of 0, -I and -II, having much lower solubility and mobility. This makes reductive immobilization an important pathway to decrease Se transport in the environment. Determining the ability of different barrier materials to reduce selenium species is therefore of high importance for the safety assessment of radioactive waste disposal initiatives.

engineered barrier for deep underground radioactive waste disposal in France. Mackinawite (FeS), a precursor of thermodynamically more stable iron sulfides minerals (e.g., greigite and pyrite), can be present naturally in host rocks of claystone and granite, and also in engineered concrete structures, such as at steel corrosion interfaces. 13 Its high reactivity and large specific surface area lead to a high potential to immobilize soluble ions, e.g., heavy metals and radionuclides ¹⁴⁻¹⁸. Previous studies have shown that Se(IV) can be reduced to a tetragonal FeSe-like phase after being in contact with mackinawite under acidic and anoxic conditions. However, red Se(0) nanoparticles could be formed though the reduction of Se(IV) by mackinawite at higher pH. 10 Along with the Se(IV) reduction, mackinawite was oxidized to green rust firstly and then to magnetite or goethite. Besides, both aqueous Fe(II) and S(-II) were able to reduce Se(IV) to Se(0) under highly alkaline conditions (e.g., at pH ~13.29)¹⁹. Meanwhile, other studies observed that Se(IV) could be reduced to elemental Se or Se-S precipitates (e.g. Se_nS_{8-n}) by aqueous sulfide species. ²⁰ However, to the best of our knowledge, little data on selenite sorption on mackinawite in the presence of cement pore water (at high pH) has been reported in literature. More specifically, precise information about the detailed interaction mechanism and the fate of each reactant in abiotic reduction processes by mackinawite under hyperalkaline conditions have not been fully unravelled. Considering that the geochemistry of selenite has close relationship with the chemical cycles of iron and sulfur in cement-rich repositories, a deeper knowledge about the interplay between selenite and mackinawite under the chemical environment of underground concrete structures (i.e., hyperalkaline and reducing conditions) is strongly required. In this study, we have investigated the interaction mechanism between SeO₃²⁻ and mackinawite in cement pore water (CPW), the solution from the interconnected pore structure produced by cement hydration, using a combination of X-ray absorption spectroscopy (XAS), X-ray photoelectron spectroscopy (XPS), transmission electron

The CEM-V/A cement, highly resistant to sulfate corrosion, is being evaluated as an

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microscopy (TEM) and wet chemistry methods. This work contributes to advance the understanding on the interaction mechanisms between redox-sensitive radionuclide oxyanions and iron sulfides and to promote our knowledge on redox chemical processes within reactive concrete barriers.

2. Materials and Methods

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- **2.1 Materials and Chemicals.** All chemicals used for synthesis of mackinawite (FeS) and stock solution (CPW) were analytical reagents. Boiled and N2-degassed Milli-Q water (18.2 M Ω ·cm) was used for all solutions and suspensions. Sodium selenite (Na₂SeO₃·5H₂O) and chemicals used for FeS synthesis and stock solution were purchased from Sigma-Aldrich. Synthetic fresh CPW at pH ~13.5 according to the recipe from CEA²¹ was used as matrix solution in all reactors, with its chemical composition shown in Table S1. All experiments were performed in a 99.99% N_2 -filled glovebox ($O_2 < 2$ ppm, using NaOH as the CO_2 trap) to prevent oxidation and possible CO2 contamination. The mackinawite suspension was prepared following the method reported previsouly¹⁰. More details are shown in Text S1 of the Supporting Information. The synthesized mackinawite solid was prepared on a silicon plate, sealed in the air-tight powder holder, and then checked by powder X-ray diffraction (XRD) (Bruker axs, D8 advance) with Vortex-EX detector (Hitachi) under Cu Ka radiation, showing no distinct impurity diffraction peak (Figure S1). The specific surface area of mackinawite was determined to be 102.3 m²/g by the Brunauer-Emmett-Teller (BET) N₂ absorption method. The surface site density was assumed to be 4 site nm⁻² and thus the corresponding surface-site concentration was 1.36 mM. The size and shape of the mackinawite particles were further characterized by TEM as illustrated in Figure S2. 2.2 Wet chemistry experiments. Sorption experiments of SeO₃²⁻ on mackinawite
- 2.2 Wet chemistry experiments. Sorption experiments of SeO₃²⁻ on mackinawite were performed at 25 °C under constant shaking and anoxic conditions (< 2 ppm O₂).
- 129 The system was tightly closed, except during sampling, to avoid CO₂ contamination.
- The solid-to-liquid (S/L) ratio in all the reaction suspensions was set to 2 g/L. The

initially added concentration of SeO₃²⁻ was 4.70 mM. Prior to introducing SeO₃²⁻, mackinawite was equilibrated with CPW for 24 h. The pH of suspension was measured by a combined glass Micro-pH electrode (Metrohm 6.0234.100) during the reaction after its calibration by pH 4.00, 7.01, 10.00 and 12.00 standard solutions. A combined Pt-ring ORP electrode (Metrohm 6.0451.100) was used for the Eh measurement, after being calibrated with Zobell's solution (200 mV at 25 °C).²² Samples of 10 mL of the suspension were filtered by 0.22 µm nitrocellulose membrane by vacuum filtration at reaction times of 1, 17, 24, 50, 73, 96, 138 and 192 hours. Extreme care was taken to minimize the potential for sample oxidation during subsequent solid characterizations (see the details in Text S2). Total aqueous concentrations of S, Fe and Se in the filtrates were analyzed by inductively coupled plasma optical emission spectrometry (ICP-OES) with a Varian 720-ES apparatus. The concentrations of SeO_3^{2-} and sulfur species, including S^{2-}/S_n^{2-} , SO_3^{2-} and SO_4^{2-} , were analyzed by ion chromatography (Dionex ICS-6000) with an DionexTM IonPacTM AS9-HC IC Columns with 0.8 mL/min Na₂CO₃ (12.5 mM) as eluent. The retention time and calibration curves for SeO₃²⁻ and sulfur species were shown in Figure S3 and the standards of IC were cross-checked by ICP-OES. 2.3 X-ray Absorption Spectroscopy (XAS). Iron and selenium K-edge X-ray absorption near-edge structure (XANES) and extended X-ray absorption fine structure (EXAFS) spectra were collected at the Core Level Absorption & Emission Spectroscopy (CLÆSS, BL22) beamline at the Spanish synchrotron ALBA-CELLS, Barcelona, Spain. Elemental Se foil and Fe foil were used for energy calibrations in parallel with the measurements at the Se K-edge (12.658 keV) and Fe K-edge (7.112 keV), respectively. Si(111) and Si(311) double crystal monochromators were used with approximately 0.3 eV resolution at 2.5 keV. A silicon drift detector (KETEK

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GmbH AXAS-M with an area of 80 mm²) was employed to collect the fluorescence

signal. All the samples for XAS were double-face sealed using polyimide tape,

mounted on a sample holder, and measured in fluorescence mode, except for Se and

Fe references, which were prepared as pellets with a cellulose matrix and measured in transmission mode. Before being transferred into the vacuum experimental chamber, samples were stored under N_2 atmosphere. For EXAFS signal collection, a liquid N_2 cryostat was used to lower the temperature to 77 K, in order to minimize the effects of thermal disorder due to atomic vibrations. The data integration and reduction of XANES and EXAFS spectra were performed by the Demeter software package.²³ Linear combination fits (LCF) were applied to identify and quantify the components in samples. Regarding the quantitative EXAFS fit of Se samples, radial distribution functions were obtained by Fourier Transform (FT) of k^3 -weighted EXAFS oscillations (k-range: 3.0–12.0 Å⁻¹) using a Kaiser-Bessel window. FEFF8.4 was used to calculate the theoretical backscattering paths to perform the fits of radial distribution functions.²⁴ Only the first atomic shell ($R+\Delta R$: 1.0–2.7 Å) was fitted considering the quality of the EXAFS data.

2.4 Transmission Electron Microscopy (TEM). The pristine mackinawite, as well as the solids and nano-sized colloids after reaction, were analyzed by TEM (JEM-2100F) with energy dispersive X-ray spectroscopy (EDS). The solid samples were dispersed in ethanol and sonicated for 5 min, while no treatment was performed for the supernatant. A drop of treated samples was deposited on ultra-thin carbon grids for the measurement. TEM images were collected at 200 keV. Gatan Digital Micrograph software was used to analyze the images.

2.5 **X-ray photoelectron spectra (XPS)**. XPS was recorded with a Kratos Axis Ultra electron spectrometer equipped with a delay line detector. The reacted solid was collected and dried in the glove box and then brought to the XPS facility using an anaerobic jar. To avoid any oxidation, the samples were quickly transferred to the XPS chamber. A monochromatic Al K α source operated at 150 W, a hybrid lens system with a magnetic lens providing an analysis area of 0.3 mm \times 0.7 mm, and a charge neutralizer were employed for the measurement. The energy step was set to 0.1 eV for collecting Fe and S data. The binding energy was always calibrated using

187 C 1s peak (284.8 eV). The XPS spectra of Fe 2p and S 2p were fitted using the
188 CasaXPS software with Gaussian Lorentzian function through
189 background-subtraction corrections using a Shirley-type optimization.

3. Results and Discussion

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3.1 Selenium reduction kinetics

The pH of the suspension stayed constant for the entire duration of the experiment, whereas the Eh decreased during the first 50 h and then remained stable until 192h (Figure S4). The Eh decrease could result from the continuous consumption of oxidants (i.e., Se(IV)), in good accordance with the Se(IV) concentration decrease as shown in Figure 1. Aqueous concentrations of SeO₃² and total Se during the reaction with mackinawite are shown in Figure 1. Nearly all the aqueous SeO₃² was removed by mackinawite after 96 h. Specifically, during the first 24 h the aqueous concentration of total Se was equal to that of SeO₃²⁻, indicating that SeO₃²⁻ was the only species in aqueous phase and that the decreased amount of SeO₃²⁻ was retained in the bulk solid phase. More intriguingly, after reacting for 50 h the aqueous concentration of SeO₃²⁻ continued to decrease while the total concentration of selenium as determined in the filtrate was increasing with reaction time. This suggests that the reduction products of SeO₃²⁻ could be nano-sized, such as nano-Se⁰, resulting in a colloidal dispersion in the suspension, which could not be separated out by the 0.22 µm membrane. After reaction for 73 h, the concentration of Se(IV) decreased to 0.17 mM but the Se_{tot} concentration was 3.01 mM, which implied a concentration of 2.84 mM aqueous Se nanoparticles in the solution. Although previous studies also showed that mackinawite had an extremely strong affinity towards Se(IV), the formation of nano-Se⁰ colloids has been rarely reported.^{2,25,26} The distinct reaction kinetics and products observed here could be probably related to the hyperalkaline nature of CPW. The disassociation of nano-Se⁰ from the solid adsorbent was likely to be accompanied with a dissolution of makinawite, which will be discussed afterwards in detail. Meanwhile, the filtrate colors turned yellow after 24 h reaction from

transparent (1 h) to orange (50 h), olive (73 h) and black (138 h and 192 h), as shown in Figure S5. The orange color at 50 h might suggest that SeO₃²⁻ was reduced to red Se⁰ nanoparticles that were released to the solution afterwards. Han et al and Breynaert et al. also observed that the makinawite suspension turned red after interacting with Se(IV) regardless of the ratios of Se(IV)/FeS; they speculated that Se(0) particles formed through Fe-mediated Se(IV) reduction.^{2, 25} Therefore, the total Se concentration of the filtrate took into account both aqueous SeO₃²⁻ and dispersed nanosized Se⁰ species. At 192 h, the concentrations of Se_{tot} and Se(IV) measured in the filtrate were 4.40 mM and 0.031 mM, respectively, indicating that ~ 0.3 mM (6 %) of Se was adsorbed on the solid and ~99 % of aqueous Se was reduced species. The mass balance of Se in the filtrate (measured) and filtered solid (calculated) at different reaction times is listed in Table S4. Moreover, the filtrate color continued to change, and turned olive and black from 73 to 192h, indicating the possible formation of green rust and magnetite. This was due to the oxidation of FeS and the subsequent release of iron nanoparticles into the solution, which caused the increased of measured Fe concentrations (Figure 1). The chemical interactions between mackinawite and SeO₃²⁻ in CPW (pH ~13.5) were simulated by Phreegc modelling using the Andra THERMOCHIMIE database.²⁷ The experimental aqueous Fe concentration (0.62 mM at 192 h) is comparable with but lower than the calculated values of 1.5 mM Fe(OH)₃ and 1.7 mM Fe(OH)₄, respectively. This is probably due to the lack of thermodynamic equilibrium. Note that the measured Fe concentrations in the filtrates also include the contributions from nano-sized magnetite as indicated by the dark colour of filtrates in Figure S5. The precise forms of Se-solid species were further explored using XAS.

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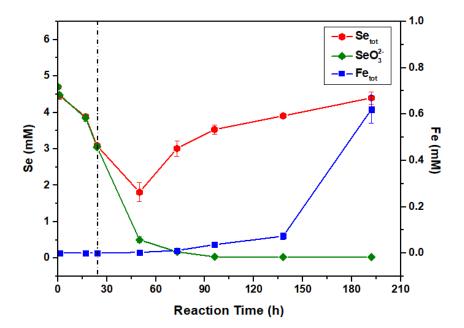


Figure 1. Measured concentration of SeO₃²⁻ by IC, total Se and Fe by ICP-OES in CPW as a function of the reaction time.

In XAS, the XANES region of the probed element can be used as a fingerprint of the chemical species, which depends both upon the valence state and the coordination environment. Se K-edge XANES spectra were collected for solid samples at different reaction times. LCF results of the solids and the references are shown in Figure 2. After reacting for 1 h, SeO₃²⁻ was mainly reduced into 35.8(3.0)% FeSe₂, 35.5(9.2)% FeSe and 19.2(1.0)% SeS₂, while only 4.9(2.9)% Se⁰ and 4.9(1.2)% SeO₃²⁻ could be identified, indicating that SeO₃²⁻ can be quickly adsorbed on the surface and reduced to Se(0, -I, -II). From 24 h to 196 h, all the spectra showed similar features, with ~63% SeS₂ and 30-32% Se⁰ identified to be the primary solid Se species. In contrast, previous studies at lower pH demonstrated that nano Se⁰ was the main reduced Se product. ^{10, 19, 25}. Kang et al. found that Se(-II) was transformed to elemental Se in the presence of remaining non-reduced selenium oxyanions. ²⁹ Therefore, the reduced species of FeSe and FeSe₂ after reacting for 1 h could be oxidized by the residual SeO₃²⁻ forming Se(0) at 24h and the corresponding reaction would be:

$$SeO_3^{2-}(aq) + 2FeSe_{(s)} + 3H_2O \rightarrow 3Se^0_{(s)} + 2Fe(OH)_{2(cr)} + 2OH^{-}$$

EXAFS data were fitted to determine the coordination environments of adsorbed

selenium, as shown in Figure S6. Fitting the main peak in R space allowed obtaining interatomic distances and coordination numbers (CN) for the first two coordination shells. As shown in Table S2, the spectrum of the 1 h sample yielded a CN ~1.7 Se-Fe pair at 2.31 Å and CN ~1.1 Se-Se pair at 2.61 Å. The Se-Se backscattering path used was the average from the Se-Se paths of FeSe, FeSe₂, SeS₂ and Se⁰. Regarding samples from 24, 73 and 192 h, apart from the Se-Se pair that was fitted at ~2.3 Å with a CN ~ 1.5, a backscattering signal from Se-S pair was also observed, resulting in a CN of ~ 0.6 -0.9 with $d_{\text{Se-S}} = \sim 2.1$ Å. The EXAFS fits also confirmed the existence of SeS_x species in the solid product. Diener et al. reported that selenium, belonging to the family of chalcophile, can incorporate into the structure of mackinawite and pyrite by replacing S sites and consequently forming zero-valent Se. Also, the decreased coordination number of the neighbour Se atoms around the incorporated Se⁰ site indicated a small cluster size, 30 in good agreement with our study. With regard to the SeS_x precipitate, the zero-valence selenium and sulfur are miscible in all proportions and can form cyclic-structures (e.g. Se_nS_{8-n}). Previous studies have showed that Se_nS_{8-n} are susceptible to form and precipitate as a result of Se(IV) reduction by sulfide, though the resulting sulfur-to-selenium ratio varied from 1.7 to 2.3.31 However, TEM imaging and EDS mapping analysis did not succeed in identifying Se⁰ nanoparticles and Se-S precipitates on the surface of solid sample for 192 h (the expected shape of Se-S particles is irregular and the elemental Se can adopt nanospheres and nanorod-like morphologies). 32, 33 Furthermore, no obvious selenium peaks of 50 and 192 h filtered solids were checked by XRD probably due to the poor crystallinity or detection limit (Figure S7). The Se was therefore likely to be present forming very small precipitates or amorphous atomic level cooperation with S (Figure S8). By combining with the information from Se K-edge XANES results, we proposed that SeO_3^{2-} was adsorbed onto surface sites corresponding to the S^{2-}/S_n^{2-} environment of mackinawite, and it was reduced by neighboring S²⁻/S_n²⁻ or Fe²⁺, forming a very fine Se⁰ phase or amorphous cluster followed by the formation of Se-S through the immediate bonding to Se⁰ and S. The S²-/S_n²- that was substituted by

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SeO₃²- dissolved into the solution (see Section 3.2). The Eh-pH diagrams of Se and Fe species calculated using Andra THERMOCHIMIE database are shown in Figure S9. It is worth noting that after 192 h zero-valent Se predominated among Se species, even though ferroselite (FeSe₂) was predicted to be more stable under the experimental conditions of pH ~13.5 and Eh ~-0.4 V as indicated in Figure S9(a). The zero-valent Se is therefore a metastable phase, with slow transformation kinetics from Se⁰ to Se(-I). These results allow making a hypothesis to explain the fact mentioned above: the total concentration of selenium was higher than that of SeO₃²⁻ after reacting for 24 h. It can be speculated here that ≡Fe-S-Se⁰ could break down to its constituent parts in CPW, i.e. into \equiv Fe-S and nano-Se⁰, or \equiv Fe-OH and nano-sized SeS species. After 192 h, the nanoparticles present in the filtrate were characterized by (S)TEM (Figure 3). As it can be seen, the size of nanoparticles was approximately ~20 to 60 nm (Figure 3a). The nanoparticles formed at 192 h consisted of Se, S, Fe and O with a ratio of Se to S roughly at 4:1 according to EDS analysis (Figure S10). Generally, elemental Se occurs in three forms: trigonal Se (t-Se), monoclinic Se (m-Se) and amorphous Se (a-Se). The interplanar spacings of trigonal Se that is considered as the most stable allotropic form are 4.96 Å and 3.78 Å along and perpendicular to the helicoid growth direction, respectively,³⁴ which was different from that (4.3 Å) observed in nanoparticles in the filtrate as shown in Figure 3b. Taking into account the EDS results, the interplanar spacing may belong to Se₄S crystalline phase, whereas no data is available to identify the lattice plane of Se₄S. Se species in the filtered solids and filtrates at different reaction times are shown in Table S3. The high-resolution transmission electron microscopy (HRTEM) image (Figure 3c) illustrated that the interplanar spacing was ~2.5 Å, corresponding to the (331) lattice plane of magnetite (Fe₃O₄).³⁵ This indicates that mackinawite transformed to nanosized magnetite during the reaction, resulting in the increased Fe concentration with time shown in Figure 1.

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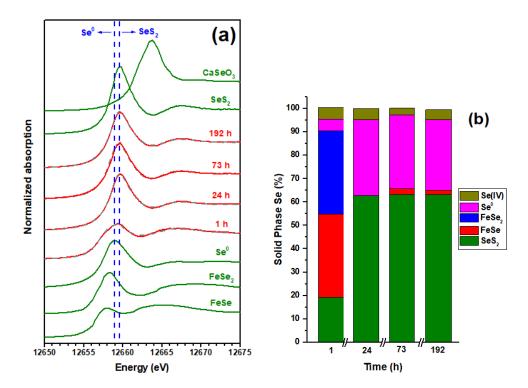


Figure 2. (a) Se K-edge normalized XANES spectra of solid samples after reacting for 1, 24, 73 and 192 h, comparing to Se references. (b) The fractions of FeSe, FeSe₂, Se⁰, SeS₂ and CaSeO₃ components in each Se-solid sample determined by LCF method.

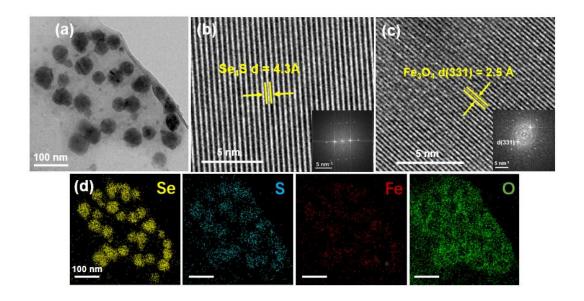


Figure 3. Transmission electron micrographs of nanoparticles found in the filtrate after reaction for 192 h. (a) Transmission electron microscopy image; high resolution transmission electron microscopy (HR-TEM) image on two spots on a single particle (b) SeS_4 and (c) Fe_3O_4 with the corresponding Fast Fourier Transform (FFT); (d)

Elemental mapping by energy-dispersive X-ray spectroscopy analysis with Se in yellow, S in blue, Fe in red and O in green.

3.2. Dissolution kinetics of mackinawite with/without SeO₃²-

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The ion concentrations of CPW equilibrated with mackinawite for 24 hours are shown in Table S4. 1.28 mM Ca²⁺ was removed from the original CPW by mackinawite after 24 h equilibration, nearly equal to the value of the mackinawite surface-site concentration of (1.36 mM). The pH value (~13.5) of CPW was much higher that the point of zero charge of mackinawite (7.5).³⁶ It is reasonable to think thus that the negatively charged surface of mackinawite favored the adsorption of Ca²⁺ from CPW leading to the formation of ≡FeSCa²⁺ surface species. 0.24 mM S²⁻ and almost no Fe were detected after 24 h, suggesting that mackinawite was not congruently dissolved. However, Phreege calculations predicted that 0.06 mM of both HS⁻ and Fe²⁺ should be released once the dissolution reached equilibrium. A plot of S concentration versus reaction time is shown in Figure 4. It can be seen that when no SeO₃²⁻ was added, the S²-/S_n²- concentration increased very slowly and aqueous SO₄²- (blank) was kept constant with increasing reaction time. This suggests a limited dissolution of mackinawite and no significant amount of aqueous SO_4^{2-} adsorbed onto mackinawite. The color of filtrate after 192 h was clear and the measured Eh of the suspension was about -0.402 V, which is closed to the Eh value (-0.398 \pm 0.005 V) of the suspension with SeO₃²⁻ (Figure S4). In contrast, in the presence of SeO₃²⁻, a critical threshold of reaction time (73 h) was observed based on the concentration change of S²⁻/S_n²⁻. Dissolved concentration of S²⁻/S_n²⁻ originating from the solid increased quickly during the first 73 h, remaining roughly constant at 13-14 mM thereafter; the same tendency for the total concentration of sulfur was observed. This supports the hypothesis that 2 to $3 S^2 / S_n^2$ units were substituted by one SeO_3^2 atom according to the concentrations of the two species. The measured concentrations of SO_4^{2-} , S^{2-}/S_n^{2-} , SO_3^{2-} and total S with the addition of SeO₃²⁻ are shown in Table S6. After reaction for 1 h, the summation (1.74 mM) of SO_4^{2-} , S^{2-}/S_n^{2-} and SO_3^{2-} concentrations by IC was lower than the total concentration of S (2.34 mM) by ICP. This suggests that other soluble S species, probably S^0 , existed in solution. However, it is almost the same value between the total S concentration and the sum from 24 h to 192 h. It means that no other S species were present in the filtrates. 14.12 mM S^2/S_n^2 was detected in the filtrate and the dissolution efficiency was ~62% after 192h. The mass balance of S in the filtrate (measured) and filtered solid (calculated) at different reaction times is listed in Table S4. Han et al. proposed a mechanism by which Se(IV) was reduced to FeSe or FeSe₂ by mackinawite at pH 8 and the replaced sulfide by Se released into the solution, with no detection of aqueous sulfide.²⁵ The SO_4^{2-} concentration remained 1.55 ± 0.03 mM during the reaction, which means that no SO_4^{2-} adsorbed on the surface of the solid during reaction. A considerable amount of aqueous SO_3^{2-} was observed following the same kinetic tendency of sulfide, resulting in 1.20 mM SO_3^{2-} after 192 h. SO_3^{2-} can be present as intermediate but stable species at high pH, as the presence of S^{2-} seems to inhibit further oxidation of SO_3^{2-37}

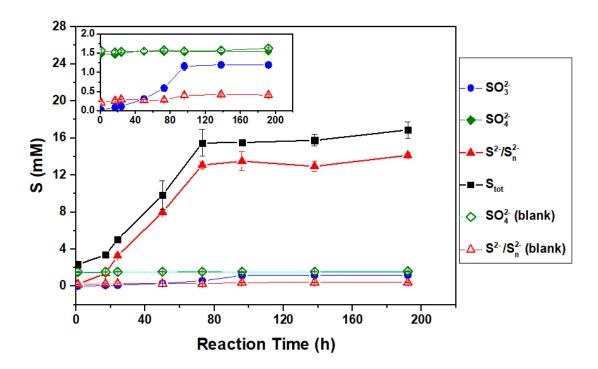


Figure 4. Measured concentration profiles of (poly)sulfide (S^{2-}/S_n^{2-}) , sulfate (SO_4^{2-}) and sulfite (SO_3^{2-}) of the experiment with SeO_3^{2-} and S^{2-}/S_n^{2-} (blank), SO_4^{2-} (blank) of

experiment without SeO₃²⁻ as a function of reaction time. The inset shows an enlarged view of the less abundant species.

3.3 Possible Reaction Involving Fe, S and Se

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S(2p) and Fe(2p_{3/2}) XPS spectra of Se-reacted mackinawite nanoparticles at different reaction times are shown in Figures 5a and 5b. In Figure 5a, to adequately fit the S(2p) spectrum of pure mackinawite (0h), three 2p_{3/2} components at 160.9 eV, 162.7 eV and 164.4 eV were required. The one at 160.9 eV, accounting for 71.3% of the total S(2p) signal, was typically attributed to sulfide(S²); the second component located at 162.7 eV may correspond to polysulfides (S_n²⁻), taking up 27.3%. The third was fitted to take into account the high energy tail of the S(2p) spectrum, attributed to elemental sulfur S⁰ (1.4%) (Table S7).³² The S:Fe ratio of 1.2 derived from TEM-EDS analysis experiments was consistent with the data of previous studies, 42 and suggests the formation of greigite (Fe₂^{III}Fe^{II}S₄) on the surface of mackinawite, although this solid was not identified by XRD. 41 Recently studies revealed the existence of Fe₃S₄ species in FeS_{nano}, as a solid-phase precursor of mackinawite, and of edge-sharing FeS₄ species to stabilize FeS structure to form mackinawite. 43-44 In our case, we speculate that Fe_3S_4 , as the representative species of solid S_n^2 , reacted with SeO_3^2 especially at 1 h. A set of redox and oxidation reactions susceptible to occur between Se couples and the considered solid product couples are listed in Table S8. After reacting with SeO₃²⁻, the S(2p) spectra exhibited different features, which were characterized by the energy increases (0.6-0.7 eV) of S^{2-} and S_n^{-2-} species and probably attributed to surface defects, Ca or Se impact. After 1 h reaction, there is evidence for thiosulfate $(S_2O_3^{2-})$ species at 168.0 eV by XPS and an obvious decrease of S_n^{2-} signal of solid coupled to an increase in S^0 and $S_2O_3^{\ 2^-}$ contributions, indicating that $S_n^{\ 2^-}$ (Fe_3S_4) on the surface of mackinawite was oxidized to S^0 and $S_2O_3^{2-}$ accompanied with SeO₃²⁻ reduction to FeSe, FeSe₂ and Se⁰ (Figure 2). The corresponding half reactions (1)-(5) are listed in Table S8. Oxidation reaction (4) and (5) suggest that S_n^{2-} (Fe₂^{III}Fe^{II}S₄) was the active reductant of aqueous SeO₃²⁻ within a short reaction

time and was responsible for the formation of FeSe and FeSe₂ at high pH via the formation of Fe(OH)2 and Fe(OH)3 species, which was also confirmed in acid solution. 45 Previous studies showed that SO₃²⁻ reacted with S⁰ to form thiosulfate (S₂O₃²⁻), which is thermodynamically stable and is an excellent complexing agent for metal ions.³⁸ Therefore, the S₂O₃²⁻ species detected in our system were probably formed though the disproportionation reaction. Also, $S_2O_3^{2-}/S_4O_6^{2-}$ was observed in the reaction between Se(IV) and pyrite by Han et al., 39 while Scheinost et al. reported that the sulfide in mackinawite was oxidized to elemental sulfur at pH 5.7, coupled to the reduction of Se(IV). 40 The content of S²-decreased from 67.6% at 1 h to 36.1% after 24 h contact, while an increase of $S_n^{\,2-}$ and $S_2O_3^{\,2-}$ content occurred simultaneously. As the oxidation state S within S^{2-} and S_n^{2-} is same, it suggests that S^{2-} was primarily oxidized to $S_2O_3^{2-}$ through the reduction of aqueous SeO_3^{2-} to zero-valent Se (SeS₂ and Se⁰). The increase of solid S_n^{2-} observed in S_{2p} XPS spectra suggests the formation of an intermediate amorphous solid phase Fe₂^{III}Fe^{II}S₄ (later transformed to pyrite), which could account for Fe(III) species in Fe(2p_{3/2}) XPS part. 46-48 Finck et al. reported that sulfide, instead of Fe(II), contributed to the reduction of SeO₃²⁻, as indicated by the detection of FeS₂ by XRD.¹⁹ Here, it should be noted as the content of Se-sorbed was quite lower than the mass of solid matrix and the primary Se XPS region is 3d, the interference of Se 3p_{3/2} (161.7 eV) could be neglected. After 73h, the percentage of S₂O₃²- peak area increased to 40.5% and 10.8% of S^0 species were also detected. In contrast, the percentages of S^{2-} and S_n^{2-} contributions decreased to 27.1% and 21.5%, respectively, suggesting that S^{2-} and S_n^{2-} were oxidized to $S_2O_3^{2-}$ and S_n^{0} (equation 5 and 6). From 73h to 192h, more S_n^{2-} was oxidized to S₂O₃², which should be responsible for the reduction of SeO₃². Usually, the oxidation of S^{2-} followed S^{2-} - S_n^{2-} - S^0 pathway during Se(VI) reduction and more than 90% of S^{2}/S_n^{2} was oxidized to S^0 with little amount of $S_2O_3^{2}$ and SO_4^{2-} at acid or weakly alkaline conditions. ^{45,49} In the present study, $S_2O_3^{2-}$ was the only oxidation product after 192 h reaction with Se(IV). The summed concentration of S^{2-}/S_n^{2-} , SO_3^{2-} and SO_4^{2-} is almost the same value of the total S (Figure 4),

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indicating that Se(VI)-sorbed onto mackinawite was reduced by solid S^{2-} , not by aqueous S^{2-} species even though plenty of aqueous S^{2-}/S_n^{2-} were present in solution. The reduction mechanism is not totally consistent with the one reported for U(VI) reduction by FeS by Hua et al, who found U(VI)-sorbed to be reduced by either S^{2-} -solid or dissolved $HS^{-,50}$ A direct reduction of Se(IV) by aqueous S^{2-} was shown to be the dominant reaction at low initial pH. Probably aqueous solid S^{2-} was more reactive for the reduction of Se(IV) than aqueous S^{2-} in CPW. Aqueous S^{2-} was stabled at pH 12.2 under reducing conditions for a long time. S^{2-}

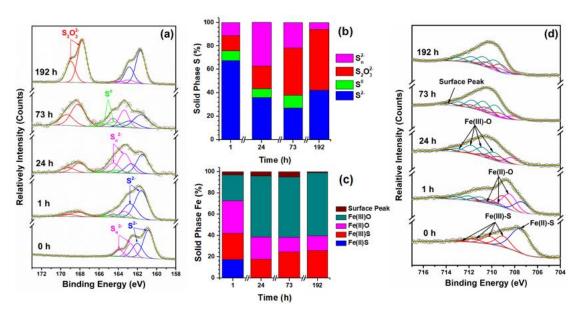


Figure 5. XPS spectra (a) narrow S(2p) and (d) $Fe(2p_{3/2})$ (d) XPS scan, and the fractions of corresponding components of (b) S and (c) Fe species. Circles were experimental data; bottom solid line was the Shirley background; dashed lines were the fit to each spectrum and top solid line was the sum of background and fitted peaks.

The broad $Fe(2p_{3/2})$ spectrum of mackinawite (Figure 5d, 0 h reaction time) could be fitted by one Fe(II)-S peak at ~707.5 eV and one main peak of Fe(III)-S at 709.1 eV with three multiple peaks. ⁴¹ The area ratio of each multiple peak to the first Fe(III)-S component is equal to 0.66, 0.31, and 0.10 respectively, in good agreement with other studies on mackinawite. ⁴² The $Fe(2p_{3/2})$ peaks fitted in the current study are listed in

Table S9. Compared with the XPS result of pristine mackinawite, three multiple peaks of Fe(II)-O and four multiple peaks of Fe(III)-O were observed in the Fe 2p spectrum of SeO₃²-reacted mackinawite at 1 h. This suggests the oxidation of Fe(II) to Fe(III) (equation 7), coupled to SeO₃² reduction (equation 1). Also, the Fe(II) hydroxyl-complexes are important in alkaline solution and Fe(OH)2 is probably formed. An increased (33.2%) Fe(III)-O peak intensity and an absence of Fe(II)-S peak are observed after 24 h reaction compared with the 1 h sample, which was probably related to the oxidation of Fe(II) and the further release S^{2-} by the solid. The XPS peak of S²- to be show in Figure 5a at 24 h may be attributed to Fe(III)-S species. Similar results of Fe(II) oxidation to Fe(III) were observed in mackinawite reacted with Se (IV) at different pH values. ^{25,40} As reaction time increased to 73 h, the content of Fe(III)-O remained constant, indicating that no more SeO₃²⁻ was reduced by Fe(II). Mixtures of Fe(II)-O and Fe(III)-O observed on the solid surface may imply a formation of multi-valent iron (oxyhydr)oxides such as green rust, magnetite, or other phases containing both ferrous and ferric sites.⁵² Furthermore, the presence of Fe(III)-S species was probably due to the transformation of FeS to amorphous Fe2^{III}Fe^{II}S4 resulting in the enrichment of S combined with Fe(III) or/and the formation of Fe^{III}S₂O₃⁺ species via Se reduction. Lan et al. suggested that Fe₃S₄ was present beneath Fe(III) (hydr)oxide precipitation during FeS transformation at pH 8.⁵³ The XRD patterns of 50 and 192 h solid samples show a poor crystallinity and differed from the original mackinawite in Figure S7. The XRD pattern from the 50 h sample probably indicated that the formation of poor crystallized Fe₃S₄ and Fe(OH)₃. in agreement with the XPS results. The very broad XRD reflections in 192 h sample is attributed to magnetite and Fe(OH)₃ with very poor crystallinity. Normalized Fe K-edge XANES spectra of the 192 h solid sample and of the references, including mackinawite, magnetite, and hematite, are shown in Figure S11, together with the LCF result. The result shows that the species present in the solid product was similar to magnetite, in good accordance with black color of the filtrate at 192 h.

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Furthermore, the ratio of Fe(II)-O/Fe(III)-O identified by XPS at 192 h was 0.23, less than the idea value of 0.5, and the presence of Fe(III)-S species suggested the final solid contained amorphous Fe(OH)₃ and FeS₂O₃⁺ complexed to Cl⁻ or OH⁻ species except very poor crystallization magnetite. The Eh-pH diagram of Fe species shown in Figure S9b indicated that mackinawite tends to dissolve and to form magnetite. It is well known that Fe(OH)₂, an extremely redox-sensitive phase in solution, could lead via the Shikorr reaction⁵⁴ to a disproportion reaction to magnetite under anaerobic condition. Furthermore, the interaction of SeO₃²⁻ with Fe(II) could produce Se⁰ and metastable green rust (GR), which is know to have the potential to reduce SeO₃²⁻ to Se⁰ via being oxidized to magnetite.⁵⁵

4. Environmental relevance

It is critical to waste depository safety to improve our understanding of reductive precipitation of mobile radionuclides in hyperalkaline conditions, and specifically to measure the surface complexes, aqueous speciation and solubility in order to decipher immobilization pathways and the underlying interaction mechanisms when contaminations are in contact with cement media.⁵⁶ Mackinawite, as one of the steel corrosion products and one of the common naturally-occurring iron sulfides, showed an extremely high sorption and reduction ability towards SeO₃²⁻ within a relatively short interaction time (~96 hours). SeO₃²-can be reduced to a mixture of SeS₂ and Se⁰ solid species, as confirmed by Se K-edge XAS. However, the zero-valent selenium species shows a relatively high solubility and thus mobility in CPW, with ~99% of reduction products released to the solution in the form of Se₄S nanoparticles confirmed by TEM-EDS after 192 hours. The surface chemistry of the mackinawite evolved during the reaction in the following wasy: 2 to 3 S²-/S_n²- units were substituted by one SeO_3^{2-} atom. 62% of initial solid S^{2-}/S_n^{2-} was released into solution when in presence of SeO₃² and mackinawite was transformed to magnetite, Fe(OH)₃ and FeS₂O₃⁺ complexed to Cl⁻ or OH⁻ species after 192 h. The current study provides

500	valuable information about the reduction mechanisms of redox-sensitive radionuclides
501	in presence of iron sulfides that have strongly reduction capacity.
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503	ASSOCIATED CONTENT
504	Supporting Information Available. Additional materials referenced in the text are
505	available free of charge.
506	The composition of CPW. XRD patterns, and TEM micrograph of synthesized
507	mackinawite. The calibration curves of SO_4^{2-} , SO_3^{2-} , S^{2-}/S_n^{2-} , SeO_3^{2-} and the relative
508	retention time of the ions. Picture of the filtrate color. Se K-edge EXAFS of solid
509	samples. STEM image and EDS mapping results of solid sample. Se and Fe species
510	Eh-pH diagrams. Aqueous concentration of SO_4^{2-} , SO_3^{2-} , S^{2-}/S_n^{2-} in the filtrate.
511	Binding energies (BE), Peak full width at half maximum (FWHM) and peak areas for
512	$S(2p)$ and $Fe(2p_{3/2})$. Fe K-edge normalized XANES spectra and LCF result of solid
513	samples.
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Reference

- 530 1. Kausch, M. F.; Pallud, C. E., Modeling the impact of soil aggregate size on
- selenium immobilization. *Biogeosciences* **2013**, *10*, (3), 1323-1336.
- 532 2. Breynaert, E.; Bruggeman, C.; Maes, A., XANES-EXAFS analysis of se
- solid-phase reaction products formed upon contacting Se(IV) with FeS2 and FeS.
- 534 Environ. Sci. Technol. **2008**, 42, (10), 3595-3601.
- 535 3. Bruggeman, C.; Maes, A.; Vancluysen, J.; Vandenmussele, P., Selenite reduction
- 536 in Boom clay: Effect of FeS2, clay minerals and dissolved organic matter. *Environ*.
- 537 *Pollut.* **2005,** *137*, (2), 209-221.
- 538 4. Charlet, L.; Scheinost, A. C.; Tournassat, C.; Greneche, J. M.; Gehin, A.;
- Fernandez-Martinez, A.; Coudert, S.; Tisserand, D.; Brendle, J., Electron transfer at
- 540 the mineral/water interface: Selenium reduction by ferrous iron sorbed on clay.
- 541 *Geochim. Cosmochim. Acta* **2007,** *71*, (23), 5731-5749.
- 542 5. Sun, W. L.; Renew, J. E.; Zhang, W. L.; Tang, Y. Z.; Huang, C. H., Sorption of
- 543 Se(IV) and Se(VI) to coal fly ash/cement composite: Effect of Cat(2+) and high ionic
- 544 strength. Chem. Geol. **2017**, 464, 76-83.
- 6. Ma, B.; Fernandez-Martinez, A.; Wang, K. F.; Made, B.; Henocq, P.; Tisserand,
- 546 D.; Bureau, S.; Charlet, L., Selenite Sorption on Hydrated CEM-V/A Cement in the
- 547 Presence of Steel Corrosion Products: Redox vs Nonredox Sorption. Environ. Sci.
- 548 *Technol.* **2020,** *54*, (4), 2344-2352.
- 549 7. Bonhoure, I.; Baur, I.; Wieland, E.; Johnson, C. A.; Scheidegger, A. M., Uptake
- of Se(IV/VI) oxyanions by hardened cement paste and cement minerals: An X-ray
- absorption spectroscopy study. Cem. Concr. Res. 2006, 36, (1), 91-98.
- 8. Onoguchi, A.; Granata, G.; Haraguchi, D.; Hayashi, H.; Tokoro, C., Kinetics and
- mechanism of selenate and selenite removal in solution by green rust-sulfate. R. Soc.
- 554 *Open Sci.* **2019,** *6*, (4).
- 9. Xu, L.; Huang, Y. H., Kinetics and mechanism of selenite reduction by zero
- valent iron under anaerobic condition activated and enhanced by dissolved Fe(II). Sci.
- 557 *Total Environ.* **2019,** *664*, 698-706.
- 10. Scheinost, A. C.; Charlet, L., Selenite reduction by mackinawite, magnetite and
- 559 siderite: XAS characterization of nanosized redox products. Environ. Sci. Technol.
- 560 **2008**, *42*, (6), 1984-1989.
- 561 11. Wu, J.; Zeng, R. J., In Situ Preparation of Stabilized Iron Sulfide
- Nanoparticle-Impregnated Alginate Composite for Selenite Remediation. *Environ*.
- 563 *Sci. Technol.* **2018,** *52*, (11), 6487-6496.
- 12. Wu, J.; Zhao, J.; Li, H.; Miao, L. Z.; Hou, J.; Xing, B. S., Simultaneous Removal
- of Selenite and Selenate by Nanosized Zerovalent Iron in Anoxic Systems: The
- Overlooked Role of Selenite. *Environ. Sci. Technol.* **2021,** *55*, (9), 6299-6308.
- 13. Ma, B.; Charlet, L.; Fernandez-Martinez, A.; Kang, M.; Made, B., A review of
- 568 the retention mechanisms of redox-sensitive radionuclides in multi-barrier systems.
- 569 Appl. Geochem. **2019**, 100, 414-431.

- 14. Kirsch, R.; Fellhauer, D.; Altmaier, M.; Neck, V.; Rossberg, A.; Fanghanel, T.;
- 571 Charlet, L.; Scheinost, A. C., Oxidation State and Local Structure of Plutonium
- Reacted with Magnetite, Mackinawite, and Chukanovite. *Environ. Sci. Technol.* **2011**,
- 573 *45*, (17), 7267-7274.
- 15. Yalcintas, E.; Scheinost, A. C.; Gaona, X.; Altmaier, M., Systematic XAS study
- on the reduction and uptake of Tc by magnetite and mackinawite. *Dalton Trans.* **2016,**
- 576 *45*, (44), 17874-17885.
- 16. Wharton, M. J.; Atkins, B.; Charnock, J. M.; Livens, F. R.; Pattrick, R. A. D.;
- 578 Collison, D., An X-ray absorption spectroscopy study of the coprecipitation of Tc and
- 579 Re with mackinawite (FeS). *Appl. Geochem.* **2000,** *15*, (3), 347-354.
- 580 17. Zhou, J. M.; Chen, S.; Liu, J.; Frost, R. L., Adsorption kinetic and species
- variation of arsenic for As(V) removal by biologically mackinawite (FeS). Chem.
- 582 *Eng. J.* **2018**, *354*, 237-244.
- 18. Jeong, H. Y.; Klaue, B.; Blum, J. D.; Hayes, K. F., Sorption of mercuric ion by
- 584 synthetic manocrystalline mackinawite (FeS). Environ. Sci. Technol. 2007, 41, (22),
- 585 **7699-7705**.
- 19. Finck, N.; Dardenne, K., Interaction of selenite with reduced Fe and/or S species:
- 587 An XRD and XAS study. J. Contam. Hydrol. **2016**, 188, 44-51.
- 588 20. Jung, B.; Safan, A.; Batchelor, B.; Abdel-Wahab, A., Spectroscopic study of
- 589 Se(IV) removal from water by reductive precipitation using sulfide. Chemosphere
- 590 **2016,** *163*, 351-358.
- 591 21. Mace, N.; Nerfie, P.; Nathalie Coreau, N.; Thory, E.; Bescop. L. P.; Touze. G.;
- 592 Spécifications techniques: Préparation et caractérisations des PCH CEM V/A
- ROMBAS dans le cadre du GL CTEC DPC/SECR/ST/2015/057 indice A. 2015.
- 594 22. Markelova, E.; Parsons, C. T.; Couture, R.-M.; Smeaton, C. M.; Madé,
- 595 B.; Charlet, L.; Van Cappellen, P. Deconstructing the redox cascade: what role do
- 596 microbial exudates (flavins) play?. *Environ. Chem.* **2017**, 14 (8), 515–524.
- 597 23. Ravel, B.; Newville, M., ATHENA, ARTEMIS, HEPHAESTUS: data analysis
- for X-ray absorption spectroscopy using IFEFFIT. J. of Synchrotron Radiat. 2005, 12,
- 599 537-541.
- 600 24. Ankudinov, A. L.; Ravel, B.; Rehr, J. J.; Conradson, S. D. Real-space
- 601 multiple-scattering calculation and interpretation of x-ray-absorption near-edge
- 602 structure. Phys. Rev. B: Condens. Matter Mater. Phys. **1998**, 58 (12), 7565–7576.
- 603 25. Han, D. S.; Batchelor, B.; Abdel-Wahab, A., Sorption of selenium(IV) and
- selenium(VI) to mackinawite (FeS): Effect of contact time, extent of removal,
- 605 sorption envelopes. J. Hazard. Mater. **2011**, 186, (1), 451-457.
- 26. Mitchell, K.; Couture, R. M.; Johnson, T. M.; Mason, P. R. D.; Van Cappellen, P.,
- Selenium sorption and isotope fractionation: Iron(III) oxides versus iron(II) sulfides.
- 608 Chem. Geol. 2013, 342, 21-28.
- 609 27. Giffaut, E.; Grive, M.; Blanc, P.; Vieillard, P.; Colas, E.; Gailhanou, H.;
- 610 Gaboreau, S.; Marty, N.; Made, B.; Duro, L., Andra thermodynamic database for
- performance assessment: ThermoChimie. *Appl. Geochem.* **2014,** 49, 225-236.

- 612 28. Ma, B.; Fernandez-Martinez, A.; Mancini, A.; Lothenbach, B., Spectroscopic
- 613 investigations on structural incorporation pathways of Fe-III into zeolite frameworks
- 614 in cement-relevant environments. Cem. Concr. Res. 2021, 140.
- 615 29. Kang, M. L.; Ma, B.; Bardelli, F.; Chen, F. R.; Liu, C. L.; Zheng, Z.; Wu, S. J.;
- 616 Charlet, L., Interaction of aqueous Se(IV)/Se(VI) with FeSe/FeSe2: Implication to Se
- 617 redox process. J. Hazard. Mater. **2013**, 248, 20-28.
- 618 30. Diener, A.; Neumann, T.; Kramar, U.; Schild, D., Structure of selenium
- 619 incorporated in pyrite and mackinawite as determined by XAFS analyses. J. Contam.
- 620 *Hydrol.* **2012,** *133*, 30-39.
- 31. Geoffroy, N.; Demopoulos, G. P., The elimination of selenium(IV) from aqueous
- 622 solution by precipitation with sodium sulfide. J. Hazard. Mater. 2011, 185, (1),
- 623 148-154.
- 624 32. Sasaki, K.; Blowes, D. W.; Ptacek, C. J., Spectroscopic study of precipitates
- 625 formed during removal of selenium from mine drainage spiked with selenate using
- 626 permeable reactive materials. *Geochem. J.* **2008**, *42*, (3), 283-294.
- 627 33. Prasad, K. S.; Vaghasiya, J. V.; Soni, S. S.; Patel, J.; Patel, R.; Kumari, M.;
- Jasmani, F.; Selvaraj, K., Microbial Selenium Nanoparticles (SeNPs) and Their
- 629 Application as a Sensitive Hydrogen Peroxide Biosensor. *Appl. Biochem. Biotechnol.*
- 630 **2015,** *177*, (6), 1386-1393.
- 631 34. Zhenghua, W.; Shiyu, Z. H. U., Rapid growth of t-Se nanowires in acetone at
- 632 room temperature and their photoelectrical properties. Front. Optoelectron. China
- 633 **2011,** *4*, (2), 188-194.
- 634 35. Turgeman, R.; Gedanken, A., Crystallization of ZnO on crystalline magnetite
- nanoparticles in the presence of ultrasound radiation. Cryst. Growth Des. 2006, 6,
- 636 (10), 2260-2265.
- 637 36. Wolthers, M.; Charlet, L.; Van der Linde, P. R.; Rickard, D.; Van der Weijden, C.
- 638 H., Surface chemistry of disordered mackinawite (FeS). Geochim. Cosmochim. Acta
- 639 **2005,** *69*, (14), 3469-3481.
- 640 37. Lothenbach, B.; Rentsch, D.; Wieland, E., Hydration of a silica fume blended
- low-alkali shotcrete cement. *Phys. Chem. Earth* **2014**, 70-71, 3-16.
- 38. Beat M., Chapter 3-Properties. Sulfur, Energy, and Environment. 1997, 38-116.
- 643 39. Han, D. S.; Batchelor, B.; Abdel-Wahab, A., Sorption of selenium(IV) and
- selenium(VI) onto synthetic pyrite (FeS2): Spectroscopic and microscopic analyses. J.
- 645 *Colloid Interface Sci.* **2012**, 368, 496-504.
- 40. Scheinost, A. C.; Kirsch, R.; Banerjee, D.; Fernandez-Martinez, A.; Zaenker, H.;
- Funke, H.; Charlet, L., X-ray absorption and photoelectron spectroscopy investigation
- of selenite reduction by Fe-II-bearing minerals. J. Contam. Hydrol. 2008, 102, (3-4),
- 649 228-245.
- 41. Herbert, R. B.; Benner, S. G.; Pratt, A. R.; Blowes, D. W., Surface chemistry and
- 651 morphology of poorly crystalline iron sulfides precipitated in media containing
- 652 sulfate-reducing bacteria. *Chem. Geol.* **1998,** *144*, (1-2), 87-97.

- 42. Mullet, M.; Boursiquot, S.; Abdelmoula, M.; Genin, J. M.; Ehrhardt, J. J., Surface
- 654 chemistry and structural properties of mackinawite prepared by reaction of sulfide
- 655 ions with metallic iron. Geochim. Cosmochim. Acta 2002, 66, (5), 829-836.
- 43. Matamoros-Veloza, A.; Cespedes, O.; Johnson, B. R. G.; Stawski, T. M.;
- 657 Terranova, U.; de Leeuw, N. H.; Benning, L. G., A highly reactive precursor in the
- iron sulfide system. *Nat. Commun.* **2018**, 9.
- 659 44. Beauvais, M. L.; Chupas, P. J.; O'Nolan, D.; Parise, J. B.; Chapman, K. W.,
- Resolving Single-layer Nanosheets as Short-lived Intermediates in the Solution
- 661 Synthesis of FeS. ACS Mater. Lett. **2021**, *3* (6), 698-703.
- 45. Kang, M. L.; Bardelli, F.; Ma, B.; Charlet, L.; Chen, F. R.; Yang, Y. Q., The
- influence of pH and reaction time on the formation of FeSe2 upon selenite reduction
- by nano-sized pyrite-greigite. *Radiochim. Acta* **2016,** *104* (9), 649-656.
- 46. Benning, L. G.; Wilkin, R. T.; Barnes, H. L., Reaction pathways in the Fe-S
- system below 100 degrees C. Chemical Geology **2000**, 167 (1-2), 25-51.
- 47. Hunger, S.; Benning, L. G., Greigite: a true intermediate on the polysulfide
- pathway to pyrite. Geochemical Trans. 2007, 8.
- 48. Bourdoiseau, J. A.; Jeannin, M.; Remazeilles, C.; Sabota, R.; Refait, P., The
- 670 transformation of mackinawite into greigite studied by Raman spectroscopy. J.
- 671 Raman Spectrosc. **2011**, 42 (3), 496-504.
- 672 49. Ma, B.; Kang, M. L.; Zheng, Z.; Chen, F. R.; Xie, J. L.; Charlet, L.; Liu, C. L.,
- 673 The reductive immobilization of aqueous Se(IV) by natural pyrrhotite. J. Hazard.
- 674 *Mater.* **2014,** 276, 422-432.
- 50. Hua, B.; Deng, B. L., Reductive Immobilization of Uranium(VI) by Amorphous
- 676 Iron Sulfide. Environ. Sci. Technol. 2008, 42 (23), 8703-8708.
- 51. Lothenbach, B.; Le Saout, G.; Ben Haha, M.; Figi, R.; Wieland, E., Hydration of
- 678 a low-alkali CEM III/B-SiO2 cement (LAC). Cem. Concr. Res. 2012, 42 (2), 410-423.
- 679 52. Han, D. S.; Batchelor, B.; Abdel-Wahab, A., XPS analysis of sorption of
- selenium(IV) and selenium(VI) to mackinawite (FeS). Environ. Prog. Sustain. Energy
- 681 **2013,** *32*, (1), 84-93.
- 682 53. Lan, Y.; Butler, E. C., Iron-Sulfide-Associated Products Formed during
- Reductive Dechlorination of Carbon Tetrachloride. Environ. Sci. Technol. 2016, 50
- 684 (11), 5489-5497.
- 685 54. Ma, M.; Zhang, Y.; Guo, Z. R.; Gu, N., Facile synthesis of ultrathin magnetic
- 686 iron oxide nanoplates by Schikorr reaction. *Nanoscale Res. Lett.* **2013**, 8, 1-7.
- 55. Borsig, N.; Scheinost, A. C.; Shaw, S.; Schild, D.; Neumann, T., Retention and
- 688 multiphase transformation of selenium oxyanions during the formation of magnetite
- via iron(II) hydroxide and green rust. *Dalton Trans.* **2018**, *47*, (32), 11002-11015.
- 690 56. Altmann, S., 'Geo'chemical research: A key building block for nuclear waste
- 691 disposal safety cases. J. Contam. Hydrol. **2008**, 102, (3-4), 174-179.